Original Article

Author for correspondence:
J.J. Bell, E-mail: james.bell@vuw.ac.nz

Domination of mesophotic ecosystems in the Wakatobi Marine National Park (Indonesia) by sponges, soft corals and other non-hard coral species

J.J. Bell[1], J. Jompa[2], A. Haris[2], S. Werorilangi[2], M. Shaffer[1] and C. Mortimer[1]

[1]Victoria University of Wellington, School of Biological Sciences, Wellington, 6140, New Zealand and [2]Research and Development Centre on Marine, Coastal and Small Islands, Hasanuddin University, Makassar, Indonesia

Abstract

Mesophotic ecosystems have been relatively poorly studied in the Indo-Pacific and in particular within the Coral Triangle region. Here we used a mini-ROV to explore the changes in major benthic groups at two sites (~200 m apart) in the Wakatobi Marine National Park, SE Sulawesi, Indonesia spanning shallow water coral reefs (5 m) to deeper water mesophotic ecosystems (80 m). We found very similar patterns at both sites where coral cover peaked at 15 m, declined rapidly by 30 m, and was virtually absent at 50 m. As coral declined there was a marked increase in sponges, soft corals and other encrusting organisms (including ascidians, bryozoans, tubeworms, gorgonians and molluscs). Importantly, our results differ from most previous studies in other geographic locations where hard corals extend much deeper. It is unclear what drives this difference but it may be related to higher levels of turbidity and therefore reduced light penetration in the Wakatobi compared with other areas, which limits the vertical extent of coral development.

Introduction

There is increasing interest in characterizing the biodiversity and ecology of mesophotic ecosystems and particularly mesophotic coral reefs (MCRs) that span the 30–150 m depth range. The increasing availability of remotely operated vehicles (ROV) means these reefs are becoming increasingly accessible (e.g. Kahng *et al.*, 2014; Englebert *et al.*, 2017). However, despite this, there are still some highly diverse geographic areas where mesophotic ecosystems have yet to be explored and characterized. Kahng *et al.* (2010) reviewed the distribution of MCR studies and at that time, no studies had been conducted in Indonesia, where biodiversity is considered to be among the highest anywhere in the world. To date, there have still been no published studies exploring Indonesian mesophotic ecosystems to see if they are dominated by coral as they are in other regions. The Wakatobi Marine National Park (WMNP) in SE Sulawesi (Indonesia) hosts rich and diverse shallow water benthic communities, but they have experienced large declines in coral cover in recent decades (McMellor & Smith, 2010). Despite the strong focus on shallow water communities in the WMNP (e.g. Bell & Smith 2004; Crabbe & Smith, 2005; Powell *et al.*, 2014), very little is known of the ecosystems below 18 m. The aim of this study was to explore and characterize deeper water (to 80 m) benthic communities at two sites in the WMNP.

Materials and methods

The WMNP (Figure 1) is Indonesia's third largest marine national park and was gazetted in 1996 (Clifton & Unsworth, 2010). The park is located in the Coral Triangle Region and supports highly diverse marine communities, but is also inhabited by over 90,000 people who are heavily dependent on reef resources for food and income. Declines in hard coral cover in shallow waters have been documented in the park since 2002 (McMellor & Smith, 2010). We conducted a number of preliminary ROV deployments at two sites, known locally as Ridge 1 and Outer Pinnacle, which are ~200 m apart (Figure 1), and are part of the same fringing reef system. For a full description of the environmental conditions and biological communities found in the shallow areas at Ridge 1 see Powell *et al.* (2014). Briefly, Ridge 1 has lower turbidity, sedimentation rates and higher chlorophyll concentrations at 10 m than other sites locally with relatively high coral cover (estimated at 35% at 10–15 m in 2010; Powell *et al.*, 2014). No environmental data have previously been collected at Outer Pinnacle.

To sample the deepwater communities, we used the mini ROV, 'SAL', which is a Deep Trekker DTG2 (see http://www.deeptrekker.com). The ROV is rated to 150 m depth (although the maximum depth of continuous reef sampled in this study was 80 m). The unit is equipped with a low resolution internal camera for orientation, an externally mounted GoPro 4 capable of up to 4 K video resolution and a laser for sizing. Although continuous reef continued to ~100 m at these sites where the reef became more broken, we only sampled to a maximum depth of 80 m to avoid getting the ROV and its cable caught in the overhangs that were dominant between
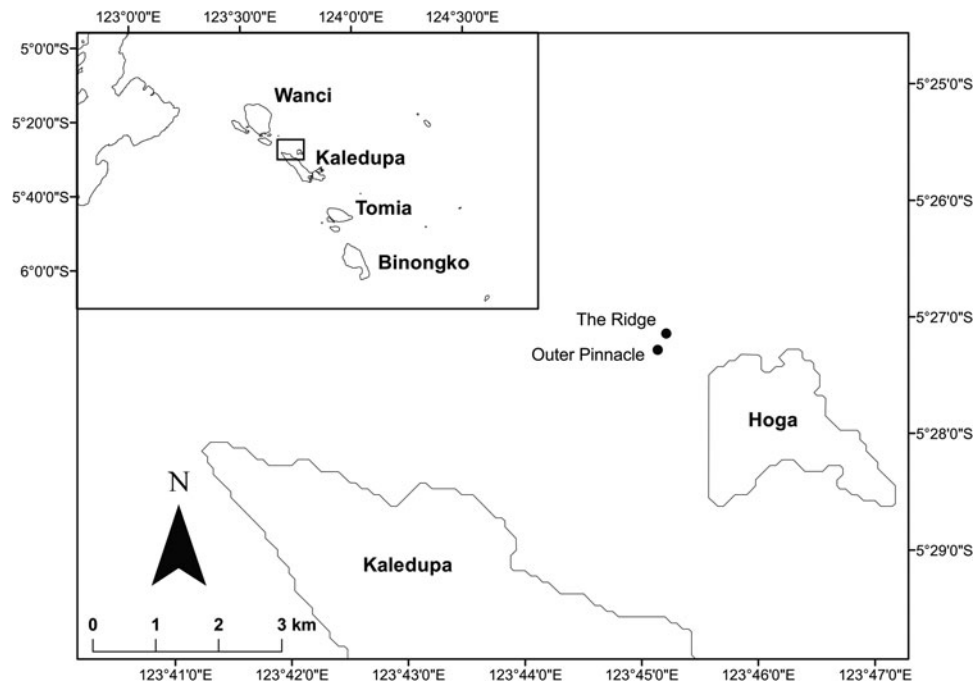
Fig. 1. Map of the Wakatobi Marine National Park, Indonesia, and location of two sites where the mesophotic reefs were sampled (adapted from Marlow *et al.*, 2018).

80–100 m. For quantitative analyses, we deployed the ROV to 80 m, and kept the ROV at that depth while moving horizontally along the reef with the camera directly focused toward the reef. At each sampling point the ROV was held as stationary as possible at a distance of 30–40 cm from the substratum for ~15 s to try and collect a steady image for processing later. After 15 s the ROV was then moved 1–2 m along the reef at the same depth and the same process repeated. After filming for ~5 min the ROV was manoeuvred to a shallower depth and the sampling continued. Sampling was undertaken at 5, 10, 15, 30, 50 and 80 m depth. The video was played back in VLC media player, and still images (N = 5–7 from each depth) were extracted from the video by haphazardly stopping the recording where the frames were in focus and the laser positioning indicated the area being examined was ~0.5 m^2 giving a total sampled area of ~2.5–3.5 m^2 at each site, at each depth. These images were collected over a 5-day period in June 2017. Coral Point Count (CPCe) analysis (Kohler & Gill, 2006) was used to measure the percentage cover of different major groups of benthic organisms within the frame grabs obtained from the ROV footage. CPCe randomly allocates points over a given image and the user then manually identifies the substrate type or benthic taxa beneath each point. The software uses this input to estimate substrate composition across the entire frame grab (percentage cover of each substrate/benthic group). We overlaid 50 points on each frame grab (N = 5–7) from each depth (5, 10, 15, 30, 50 and 80 m) at each site (N = 2). The benthic categories used in this study were hard coral, soft coral, sponge, crustose calcareous algae, bare substrate and 'other', which included a range of tubeworms, molluscs, ascidians, bryozoans, gorgonians and other algae. Gorgonians are likely under-represented in our sampling as there were many large examples, but we specifically avoided these areas to avoid damage and entanglement.

Multivariate analyses were conducted within the PRIMER-E v6 software package with the Permutational Multivariate Analysis of Variance (PERMANOVA) add-on (Clarke & Gorley, 2006) to compare the benthic community composition between depths and sites. The benthic community data analyses were based on resemblance matrices calculated from square-root transformed data using Bray–Curtis similarity coefficients. Data were transformed to reduce the influence of abundant and rare groups
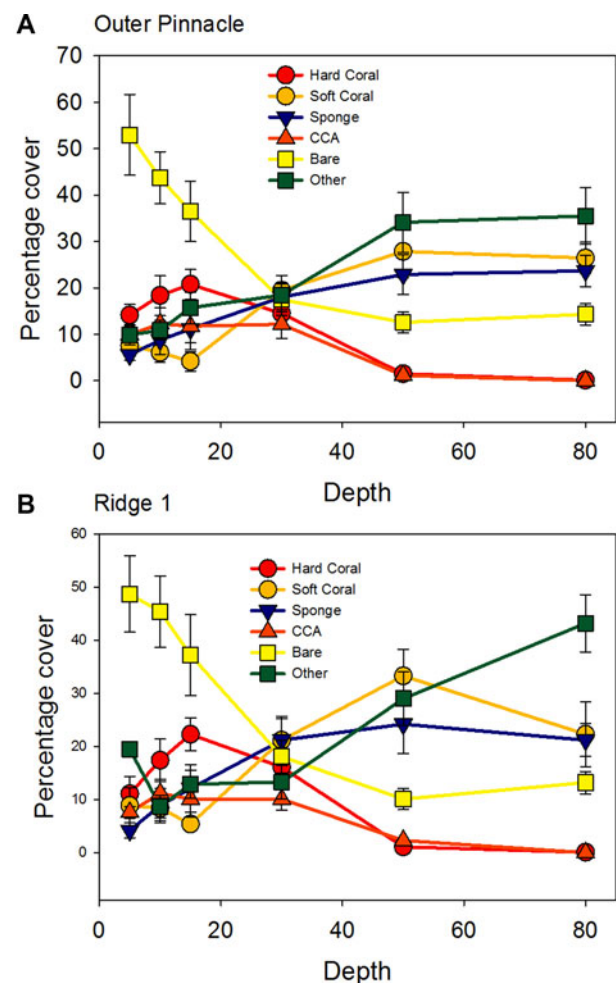

Fig. 2. Changes in the benthic organisms with depth (metres) at two sites (~200 m apart) in the Wakatobi Marine National Park, Indonesia.
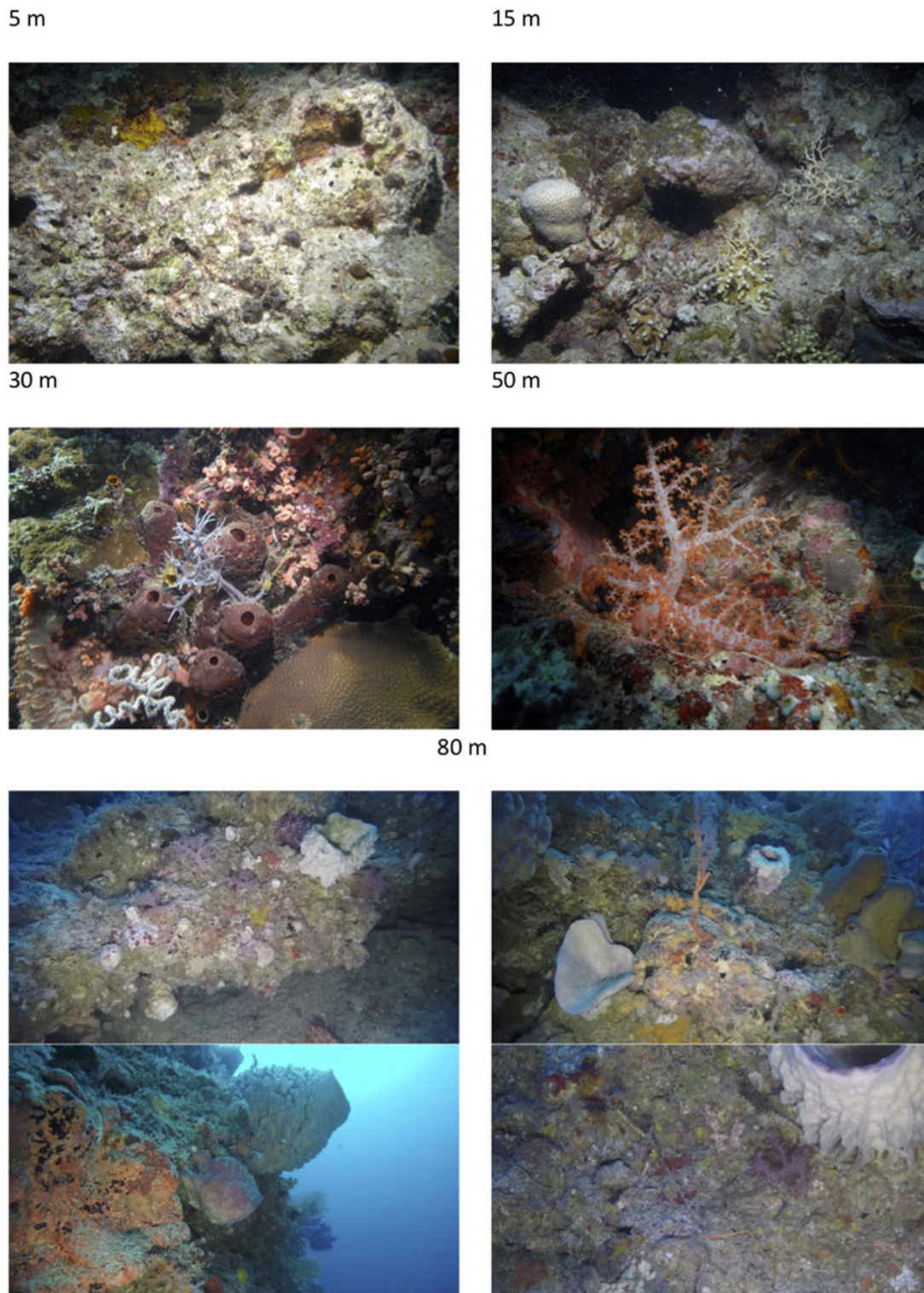
Fig. 3. Examples of the benthic communities at 5, 15, 30, 50 and 80 metres. Note that the pictures taken at 5 and 10 m were taken on scuba. 5 m – dominated by bare space; 15 m – dominated by coral and bare space; 30 m – dominated by mixed community of coral, sponges, CCA and soft coral (hard coral 15%); 50 m – dominated by mixed community of sponges, soft coral, and other benthic encrusting animals (hard coral <1%); dominated by mixed community of sponges, soft coral and other benthic encrusting animals.

(Clarke & Warwick, 2001). Differences in assemblage composition among depths and sites was assessed using PERMANOVA with depth and site as factors.

Results

The two areas where the ROV was deployed showed very similar changes in benthic community composition with increasing depth (Figures 2 and 3). This was confirmed by the PERMANOVA, which showed a significant difference between depths (PERMANOVA, $F_{(5,58)} = 38.904$, $P = 0.001$), but not between sites (PERMANOVA, $F_{(1,58)} = 1.04$, $P = 0.31$). Both sites were dominated by bare space at 5, 10 and 15 m (~50–30%), bare space then decreased considerably with increasing depth, to be colonized by a range of benthic organisms. Coral cover increased with depth, reaching a maximum at 15 m, but then

declined substantially by 30 m, to be replaced by sponges, soft corals and other benthic organisms. We found <1% coral cover at 50 m and below, and much less bare space at 30 m and deeper compared with shallower water. The complexity of the reef also decreased with depth, with mostly encrusting, two-dimensional forms dominating below 50 m. The exception was the presence of large (greater than 1.5 m in diameter in many cases) barrel sponges (*Xestospongia* spp.) and sea fans (see Figure 3).

Discussion

Here we present the first description of the organisms inhabiting mesophotic habitats in Indonesia. However, it is important to note that we have only quantitatively examined a very small area of the total reef area at the two sites and it is possible these areas are not representative of the wider Wakatobi. However, other ROV deployments across other areas at our sampling sites, along with shorter non-quantitative deployments at other sites in the Wakatobi by the authors also support the low coral abundance that we reported below 50 m. At 50 m and below we found a dominance of sponges, soft corals, gorgonians and other benthic, non-coral organisms.

Other studies from mesophotic depths in the Caribbean (e.g. Hoeksema *et al.*, 2017), Hawaiian Islands (Rooney *et al.*, 2010) and the Great Barrier Reef (Englebert *et al.*, 2017) have reported the presence of coral at much great depths (>120 m) than we reported here and there are a number of potential explanations for this difference including: (1) light penetration to support corals being much more limited in the Wakatobi compared with other sites; (2) increased sedimentation with depth, which smothers corals; (3) damage to the hard coral as a result of recent bleaching events in the Wakatobi in 2016 (authors' unpublished data); and (4) regional-scale differences between the mesophotic environments in the Wakatobi and other geographic areas. Reduced light penetration seems the most plausible explanation since benthic communities at 50 m and below seem well developed, with large sponges and soft corals, and CCA abundance also showed a similar overall pattern of decline with depth. Initial observations of the videos also suggest that the assemblage composition for specific taxa is also different in deeper water compared with shallow water, particularly for sponges and soft corals, where many species at 50–80 m are not seen near the surface. It is also possible that sedimentation could influence the benthic community distribution, as we found some evidence of sediment accumulation on rock surfaces (<3%), although it did not appear to increase with depth.

In shallow water we found a high proportion of bare space that probably represents the long and continued history of exploitation of the shallow water areas on reefs generally in the Wakatobi (see McMellor & Smith, 2010). Interestingly, our estimates of coral cover are markedly lower than those reported in 2013, suggesting on-going reef degradation in shallow water. As our study is the first to sample mesophotic ecosystems in the Wakatobi, it is unclear if these deeper water areas are being influenced by surface water anthropogenic activities, and future monitoring of these deeper water areas is required.

Sponges were an important component of the benthic community below 50 m and there were many large barrel sponges. Interestingly, unlike their shallow water counterparts, these sponges were all 'bleached' white, suggesting they have much reduced abundance of cyanobacteria in their outer tissue layer. Since there is no reason to believe this is not the natural state of these sponges, it does suggest *Xestospongia* spp. are not particularly dependent on their cyanobacteria for nutrition. However, *Xestospongia muta* from the Caribbean has been reported to undergo both cyclical bleaching and reduction in cyanobacterial abundance, but no stress is caused to the sponge (López-Legentil *et al.*, 2008). While it is possible that cyclical bleaching is occurring in the *Xestospongia* spp. in the Wakatobi, it seems unlikely since no shallow water barrel sponges were 'bleached'.

In conclusion, our results show that the deeper water communities in the Wakatobi are dominated by soft corals, sponges and other benthic organisms rather than hard corals like other mesophotic habitats. Further research is needed to determine how these reefs function, how they differ from shallow water reefs, and if they are able to support shallower water communities, for example through larval subsides. There is also a need to quantify the importance of environmental parameters that might explain the depth-related patterns we have observed, as currently data are only available for shallow water (10–15 m) environments (see Powell *et al.*, 2014; Marlow *et al.*, 2018). Importantly, there has been discussion in the literature about the potential for shallow water coral reefs to transition to reefs dominated by other groups of organisms including sponges and soft coral (Bell *et al.*, 2013), and understanding the ecology of these deeper water reefs could provide insights into how altered shallow water reefs might function since they are already dominated by these groups.

References

Bell JJ, Davy SK, Jones T, Taylor MW and Webster NS (2013) Could some coral reefs become sponge reefs as our climate changes? *Global Change Biology* 19, 2613–2624.

Bell JJ and Smith D (2004) Ecology of sponge assemblages (Porifera) in the Wakatobi region, south-east Sulawesi, Indonesia: richness and abundance. *Journal of the Marine Biological Association of the United Kingdom* 84, 581–591.

Clarke KR and Gorley RN (2006) *PRIMER v6: User Manual/Tutorial*. Plymouth: Primer-E.

Clarke KR and Warwick RM (2001) *Change in Marine Communities: An Approach to Statistical Analysis and Interpretation*, 2nd Edn. Plymouth: PRIMER-E.

Clifton J and Unsworth RKF (2010) Introduction to the Wakatobi National Park. In Clifton J, Unsworth RKF and Smith DJ (eds), *Marine Conservation and Research in the Coral Triangle: The Wakatobi National Park*. New York, NY: Nova Publishers, pp. 1–9.

Crabbe MJC and Smith DJ (2005) Sediment impacts on growth rates of *Acropora* and *Porites* corals from fringing reefs of Sulawesi, Indonesia. *Coral Reefs* 24, 437–441.

Englebert N, Bongaerts P, Muir PR, Hay KB, Pichon M and Hoegh-Guldberg O (2017) Lower mesophotic coral communities (60–125 m depth) of the northern Great Barrier Reef and Coral Sea. *PLoS ONE* 12, e0170336.

Hoeksema BW, Bongaerts P and Baldwin CC (2017) High coral cover at lower mesophotic depths: a dense *Agaricia* community at the leeward side of Curaçao, Dutch Caribbean. *Marine Biodiversity* 47, 67–70.

Kahng SE, Copus JM and Wagner D (2014) Recent advances in the ecology of mesophotic coral ecosystems (MCEs). *Current Opinion in Environmental Sustainability* 7, 72–81.

Kahng SE, Garcia-Sais JR, Spalding HL, Brokovich E, Wagner D, Weil E, Hinderstein L and Toonen RJ (2010) Community ecology of mesophotic coral reef ecosystems. *Coral Reefs* 29, 255–275.

Kohler KE and Gill SM (2006) Coral Point Count with Excel extensions (CPCe): a Visual Basic program for the determination of coral and substrate coverage using random point count methodology. *Computers and Geosciences* 32, 1259–1269.

López-Legentil S, Song B, McMurray SE and Pawlik JR (2008) Bleaching and stress in coral reef ecosystems: hsp70 expression by the giant barrel sponge *Xestospongia muta*. *Molecular Ecology* **17**, 1840–1849.

Marlow J, Schönberg CHL, Davy SK, Haris A, Jompa J and Bell JJ (2018) Bioeroding sponge assemblages: the importance of substrate availability and sediment. *Journal of the Marine Biological Association of the United Kingdom*. https://doi.org/10.1017/S0025315418000164.

McMellor S and Smith DJ (2010) Coral reefs of the Wakatobi: Abundance and biodiversity. In Clifton J, Unsworth RKF and Smith DJ (eds), *Marine Research and Conservation in the Coral Triangle: The Wakatobi National Park*. New York, NY: Nova Publishers, pp. 11–26.

Powell A, Smith DJ, Hepburn LJ, Jones T, Berman J and Bell JJ (2014) Reduced diversity and high sponge abundance on a Sedimented Indo-Pacific Reef System: implications for future changes in environmental quality. *PLoS ONE* **9**, e85253.

Rooney J, Donham E, Montgomery A, Spalding H, Parrish F, Boland R, Fenner D, Gove J and Vetter O (2010) Mesophotic coral ecosystems in the Hawaiian Archipelago. *Coral Reefs* **29**, 361–367.