

Abundance, distribution and conservation of Rio Branco Antbird *Cercomacra carbonaria* and Hoary-throated Spinetail *Synallaxis kollari*

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Summary

Cercomacra carbonaria and *Synallaxis kollari* are passerine birds endemic to the gallery forests of Roraima state in northernmost Brazil and adjacent Guyana. The IUCN Red List of Threatened Species lists both as Vulnerable but they have been removed from Brazil's list of threatened species because of data deficiency. They are poorly known, reflecting both Roraima state's distance from Brazil's main population centres and the inaccessibility of their habitat. In 2004 and 2005, we conducted bird surveys along the major rivers that provided previous sightings, and expanded records from only a handful to several dozens. We found *C. carbonaria* at 29% of the points surveyed, and estimated its local population density at approximately 80 individuals km^{-2} and total population size to exceed 15,000 individuals. The species has 723 km^2 of available habitat, 8% of which is inside conservation units. We found *S. kollari* at 44% of the points surveyed, and estimated its local population density as approximately 60 individuals km^{-2} , with an estimated total population size exceeding 5,000 individuals. It has 206 km^2 of available habitat, none of which is inside conservation units. We recommend that *C. carbonaria* be down-listed on the IUCN Red List to the Near Threatened category, and that *S. kollari* be listed as Endangered. Both species live in areas vulnerable to habitat loss. We also recommend that both species re-enter the Brazilian list of threatened species and highlight the importance of indigenous reserves to their conservation.

Resumo

Cercomacra carbonaria e *Synallaxis kollari* são aves passeriformes das matas de galeria do Estado de Roraima em sua parte mais ao norte do Brasil e da Guiana adjacente. Ambas as espécies são catalogadas como "Vulneráveis" pela lista vermelha de espécies ameaçadas da IUCN, mas foram excluídas da lista de espécies ameaçadas do Brasil devido à insuficiência de dados. Estas espécies são pouco conhecidas, refletindo tanto a distância do Estado de Roraima dos principais centros urbanos quanto a inacessibilidade do seu hábitat. Em 2004 e 2005, realizamos censos de aves anteriormente avistadas ao longo de rios maiores, aumentando assim os registros de apenas alguns poucos números para diversas dezenas. Encontramos *C. carbonaria* em 29% dos pontos amostrados e estimamos sua densidade local em ~ 80 indivíduos km^{-2} e o tamanho populacional total em mais de 15.000 indivíduos. A espécie possui 723 km^2 de habitat disponível, 8% do qual está incluído em unidades de conservação. Encontramos *S. kollari* em 44% dos pontos amostrados e estimamos sua densidade populacional local em ~ 60 indivíduos km^{-2} , com uma estimativa de tamanho populacional total em mais de 5.000 indivíduos. Esta espécie tem 206 km^2 de habitat disponível, nenhum dos quais incluído em unidades de conservação. Recomendamos que *C. carbonaria* seja rebaixada para a categoria 'Quase Ameaçada' e *S. kollari* categorizada

como 'Ameaçada' na lista vermelha da IUCN. Ambas as espécies vivem em áreas vulneráveis em relação à perda de seu habitat. Recomendamos, também, que ambas as espécies sejam re-incluídas na lista de espécies ameaçadas do Brasil, destacando a importância das reservas indígenas para a sua conservação.

Introduction

Roraima state is located in extreme northern Brazil, bordering Venezuela to the north and west, Guyana to the north and east, and Amazonas state to the south. Roraima has a more heterogeneous vegetation than would be expected from its location in the Amazon Basin, including evergreen tropical forests, semi-deciduous forests, swamps, white sand forests (*campina* and *campinarana*) and savannas (*lavrado*) (Furley and Mougeot 1994). Nearly all water bodies in the savannas are fringed by gallery forests, which do not flood during the high-water season (unlike *várzea* forest) and are characterized by shrubby vegetation with a dense understorey dominated by vines. Stattersfield *et al.* (1998) identified the Rio Branco gallery forest of Roraima as an Endemic Bird Area (EBA 063) based on the presence of two endemic passerines specialized to this linear riverine habitat: the Rio Branco Antbird (*Cercomacra carbonaria*) and the Hoary-throated Spinetail (*Synallaxis kollari*). Wege and Long (1995) recognize four Key Areas for Threatened Birds within the Rio Branco gallery forest EBA: Boa Vista and Rio Mucajaí, based on the presence of *C. carbonaria*, and Conceição do Mau and Rio Surumu, based on the presence of *S. kollari*. Although these key areas are in the northern, more populated, portion of the state, they are not formally protected within conservation units.

According to the map of Monteiro and Sawyer (2001), which combines demographic and socio-economic indices, Boa Vista, Rio Surumu and Rio Mucajaí are in the counties with the highest anthropogenic pressures in Roraima state. Conceição do Mau, a small town within the Raposa Serra do Sol Indigenous Reserve, is subject to lower levels of such pressures.

Zimmer *et al.* (1997) identified uncontrolled fires and conversion to rice production as the main threats facing the gallery forests of Roraima. Since their assessment, however, the establishment of new rice plantations along riverbanks has increased dramatically (Cordeiro 2005) and these are now the single major threat to these forests.

Cercomacra carbonaria occurs in the gallery forests along the entire Rio Branco and its major tributaries, while *S. kollari* is restricted to the tributaries (Remsen 2003, Zimmer and Isler 2003, Naka *et al.* 2006). Parker *et al.* (1996) considered *C. carbonaria* to be "fairly common" and *S. kollari* "probably rare" and ranked both species as a medium-level priority for research and conservation. IUCN lists both species as Vulnerable based on their estimated small ranges, continuing habitat loss and population decline (IUCN 2004, BirdLife International 2006a, b). Information about these species is still very limited, as evidenced by the quality of the data available to assess their conservation status (BirdLife International 2006a, b). The only pertinent literature published since the Collar *et al.* (1992) assessment is a study on *C. carbonaria* vocalization and behaviour (Zimmer *et al.* 1997) and a few additional records (Forrester 1992, Stotz 1997, Grosset and Minns 2002, Santos 2003). As a consequence of clear deficiency in data, *C. carbonaria* and *S. kollari* were removed from Brazil's most recent official list of threatened species (Brasil 2003, Machado *et al.* 2005).

Following the recommendations of Collar *et al.* (1992), we investigated both *C. carbonaria* and *S. kollari* to revise and offer new information on the species' geographic range limits, investigate abundance at the local and regional levels, and estimate available habitat and global population size.

Methods

Bird surveys: regional abundance

Following explorations in July 2003, we conducted fieldwork from 10 July to 24 August 2004. These months represent the height of the Roraima rainy and high water level season (Barbosa

1997), which coincide with the breeding season of both species (Zimmer *et al.* 1997, Vale *et al.* 2005). Here we use the terms “right bank” and “left bank” according to the traditional Amazonian system, based on the hand each bank faces when descending the watercourse (Whittaker and Oren 1999).

We determined the presence or absence of *C. carbonaria* and *S. kollari* along the major rivers where they occur with field assistance of a very experienced boatman, Claudiomiro Parente (C. P.). We surveyed 157 km on the right bank of the mid- and lower Rio Uraricoera, 114 km on the right bank of the mid- and lower Rio Tacutu, and 70 km on the left bank of the mid- and lower Rio Surumu (Figure 1A). On 13 July 2005 we surveyed an additional point on the Rio Uraricoera, 3 km away from the uppermost point surveyed in 2004, where C. P. reported hearing both species in August 2003.

We used a boat to survey the strip of gallery forest along these rivers, stopping every 3 km to perform playback. At each point, we broadcast a 3:47 minute pre-set sequence of vocalization intercalated with silence for each species using a Sony TCM-5000 EV tape recorder. *C. carbonaria*'s sequence was always broadcast before *S. kollari*'s. We created the sequence in the computer using a commercial recording of *C. carbonaria*'s song (Isler and Whitney 2002) and a recording of *S. kollari*'s song made by Jeremy Minns (Grosset and Minns 2002). Each point was visited only once. No work was done under rainy or windy conditions.

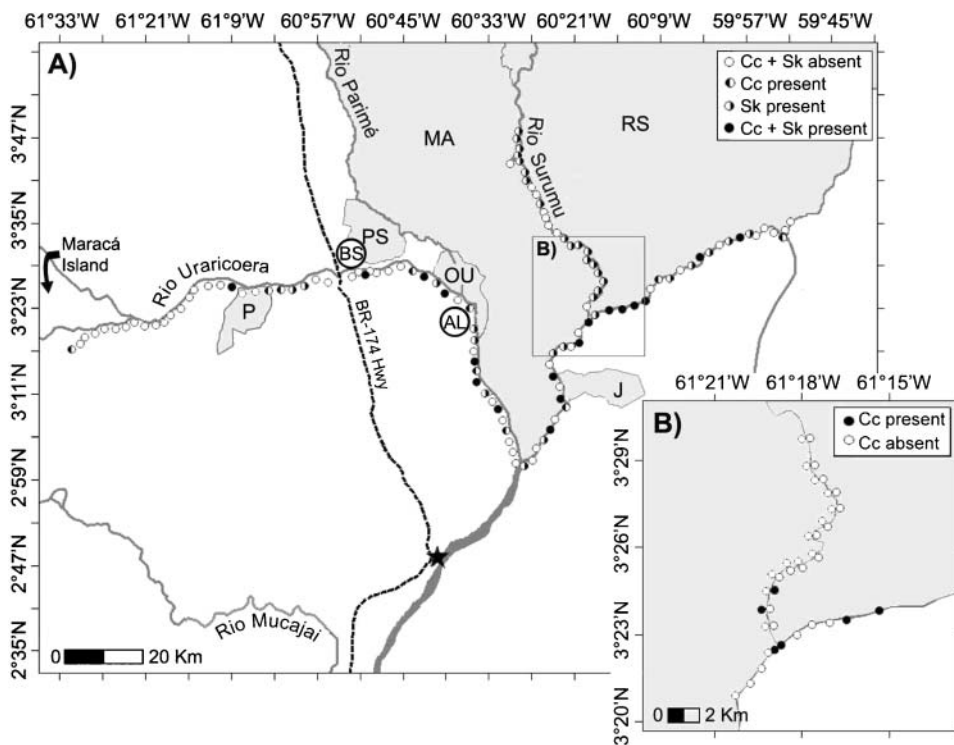


Figure 1. Regional abundance survey. (A) Rio Uraricoera, Rio Tacutu and Rio Surumu survey (3 km sampling regime). (B) Rio Surumu survey (1 km sampling regime). Species: Cc, *Cercomacra carbonaria*; Sk, *Synallaxis kollari*'s. Local abundance survey localities: AL, Alagadiço; BS, Bridge site. Indigenous Reserves within species range (grey): J, Jaboti Indigenous Reserve; OU, Ouro Indigenous Reserve; P, Pium Indigenous Reserve; PS, Ponta da Serra Indigenous Reserve; RS, Raposa Serra do Sol Indigenous Reserve; MA, São Marcos Indigenous Reserve.

We used playback to improve our chances of detecting the birds. Playback is especially recommended for secretive species and dense habitats (Johnson *et al.* 1981), which is the case for both species studied. The preferred habitat of *C. carbonaria* and *S. kollari* is nearly impenetrable (Zimmer *et al.* 1997), and although both species are secretive, we found in our 2003 explorations that they respond well to playback.

We divided the number of points where a species was present by the total points surveyed to estimate of the probability of finding the species in a given site within its geographic range. It is important to note that although presence data are definite, absence data are not. If a given species responded to playback, we were sure it was present at that point but if it did not respond it could either have been absent or present but unresponsive. Therefore, our estimate of the probability of finding a species in a given site is likely to be underestimated.

Bird surveys: local abundance

We used playback counts to estimate local abundance (Bibby *et al.* 2000). The fieldwork on local abundance ran from 10 July to 5 August 2004. The surveys took place in two sites on the Rio Uraricoera: 'Alagadiço' and 'Bridge' sites (Figure 1A). Both sites had a strip of dense gallery forest dominated by shrubs and vines, with an average canopy height of 4 m, and a sharp edge with savanna originally converted to rice plantation and cattle pasture and subsequently abandoned. We chose these sites based on the confirmed presence of both species in July 2003 exploratory surveys. Alagadiço (03°22' N, 60°35' W) is located on the right bank of the Uraricoera River, inside Fazenda Truarú, a farm 67 km north of Boa Vista. The patch of forest surveyed at Alagadiço was 4 km long, had an area of 39.6 ha, and a width ranging from 3 to 300 m. This site was relatively undisturbed by human visitation, having at the time of the study only one household within the 330 km² of Fazenda Truarú. The Bridge site (03°27' N, 60°54' W) is 86 km north of Boa Vista, on the left margin of the Rio Uraricoera near the BR-174 highway bridge that spans the river. The patch of gallery forest surveyed at the Bridge site was 3.5 km long, had an area of 37.5 ha, and a width ranging from 11 to 220 m.

At both sites, we flagged points at 200 m intervals, as suggested in Hutto *et al.* (1986) and Gutzwiller (1991), along a pre-existing trail parallel to the inland margin of the gallery forest. We surveyed the maximum number of points we could fit into each patch of gallery forest: 20 points (4 km) in Alagadiço and 17 points (3.5 km) in the Bridge site. Each site was visited three times. During each visit, two surveys were carried out by two observers starting at the same time from opposite ends of the line, such that each point was surveyed twice in each visit. The three visits and two surveys per visit yielded six surveys per site.

The surveys started at 08h00 and took 3–5 hours to complete. We standardized the time spent at each point by the length of the playback (3:47 minutes). In contrast, the time spent between points varied considerably between sites and observers due to heterogeneity in terrain condition and observers' travel speed. In the field, the observer broadcasted the same pre-set vocalizations used at the regional survey at each survey point, recording all individuals heard or seen within 100 m from the observer during the playback.

Bird surveys: geographic range limits

We determined the species geographic range limits from: (1) the literature, (2) fieldwork carried out in 2004 during the regional abundance bird survey (mentioned above), and (3) fieldwork carried out in 2005 in key areas, using the same 3 km and playback sampling regime as in the regional abundance survey.

Between 5 and 14 September 2005, we surveyed 23 km upriver on the Rio Mau starting at the uppermost record in this river for *C. carbonaria* and *S. kollari* (O'Shea *et al.* 2007), 55 km upriver on the Rio Mucajaí starting at the uppermost record for *C. carbonaria* on this river

(Santos 2003), and 46 km upriver on the Rio Parimé starting at its mouth with the Rio Uraricoera.

Data analysis: local population density

We used the local abundance data to estimate population density at the Alagadiço and Bridge sites. The total number of individuals at a site (n) was defined as the sum of the maximum number of individuals recorded at each point surveyed for that site. While Zimmer *et al.* (1997) recorded *C. carbonaria* female response to playback, in our study only males responded. This may be a consequence of female occupation with nesting duties during the breeding season. There is no sexual dimorphism in *S. kollari* and we do not know of differential response to playback vocalization (Grosset and Minns 2002). In Manú National Park, Peru, both male and female of the genera *Synallaxis* respond to playback (J. Terbough, pers. comm. 2006). We considered that both male and female *S. kollari* responded to playback, using the total number of individuals recorded as the estimate of the number *S. kollari* individuals at each site. We therefore used the total number of individuals recorded for *S. kollari* estimates and twice the total number recorded for *C. carbonaria* estimates.

Both species often came close to the observer without vocalizing in response to broadcasting. This usually enabled first detection just a few metres from the observer. Therefore, the distance between the observer and the bird was omitted from population density estimates, as it would artificially inflate the estimate. At each point, we only recorded birds that were within a 100 m radius, considering birds beyond this distance as belonging to adjacent points.

We calculated male population density (D) per site as the total number of individuals (n) divided by the area surveyed, that area being within a 100 m radius of the survey point.

We used regression and one-way ANOVA to test whether the number of missed individuals at a given point was influenced by: (1) the time of day at which the point was surveyed, and (2) whether it was the first or second time this point had been surveyed that day. We estimated the number of individuals missed per point in each survey by subtracting the number of individuals recorded at each point in that survey from the maximum number of individuals recorded for that point in all surveys. The regression of the number of individuals missed against the time of day revealed no relationship for any of the species in either site ($r^2_{\max} < 0.01$, $P_{\min} > 0.05$), possibly because point counts supplemented by playback equalize detection rates at different times of day (Sliwa and Sherry 1992). At both sites and for both species, a single-factor ANOVA showed no difference in number of missed individuals in points surveyed for the first versus second time that point was surveyed (ANOVA $P_{\min} > 0.05$), which indicates that the simultaneous surveys did not influence the results.

Data analysis: available habitat and global population size

Zimmer *et al.* (1997) consider *C. carbonaria* habitat to be confined to 0.5 km from the riverbanks where the species occur. Because *S. kollari* is believed to share the same habitat as *C. carbonaria*, we used the Zimmer *et al.* (1997) assessment of habitat for both species. We determined the potential available habitat as the entire area within 0.5 km from water, either along riverbanks or islands, within the known geographic range of the species. We used three mosaics of Landsat images (Mr-Sid N-20-00, N-21-00, S-20-00) and geographic information system (GIS) buffering techniques. These mosaics are from year 2000, cover an area of 555 km × 768 km, and have a resolution of 14.25 m.

From the total potential habitat, we visually identified and removed areas that did not correspond to gallery forest, which is the preferred habitat of *C. carbonaria* and *S. kollari*. We divided the non-gallery-forest areas (hereafter called 'non-habitat') into natural landscapes (savanna, sand banks and water bodies) and human-modified landscapes (urban areas, roads and farms). We calculated available habitat as the total potential habitat area minus the non-habitat

areas. Although the visual identification of habitat and non-habitat areas is straightforward, the distinction between natural and human-modified non-habitat areas can be subjective and the results should be interpreted with caution.

We calculated the area of available habitat within conservation units and indigenous reserves using the GIS database of the SIGLAB at the Instituto Nacional de Pesquisas da Amazônia (INPA) and the online interactive maps of Instituto Socioambiental.

We multiplied the available habitat by the mean population density (between Alagadiço and Bridge sites) to estimate species' maximum global population size. This is the *maximum* global population because it considers the entire available habitat to be occupied by the species. We calculated a minimum global population size by multiplying the available habitat by the probability of finding the species in a given site. This is a *minimum* population size because the probability of finding the species in a given site is likely to be underestimated. The species' global population size should lie somewhere between the estimated maximum and minimum population size.

Results

Regional abundance

S. kollari was present at 44% of the 121 points surveyed during the 2004 fieldwork, being most common along the Rio Surumu (67% of 27 points surveyed), followed by the Rio Tacutu (54% of 41 points) and the Rio Uraricoera (36% of 36 points surveyed up to the species' range limit on this river). *C. carbonaria* was present in 24% of the 121 points surveyed, being more abundant on the Rio Tacutu (37% of 41 points surveyed), followed by the Rio Uraricoera (12% of 53 points) and absent from the Rio Surumu (Figure 1A). To confirm the absence of *C. carbonaria* from the Rio Surumu, we revisited this river 21 days later, stopping every 1 km on both margins for 16 km. In the second, more thorough, survey of the Rio Surumu we found *C. carbonaria* at three points, the one farthest upriver being only 2.8 km away from the mouth of the Surumu (03°24' N, 60°19' W) (Figure 1B).

Local abundance

Table 1 shows the number of individuals recorded and the estimated population density for Alagadiço and the Bridge sites. *S. kollari* had a mean density estimate of approximately 60 individuals km⁻² and *C. carbonaria* of approximately 75 individuals km⁻². The Bridge site had the highest male population density for both species (Table 1).

Both *C. carbonaria* and *S. kollari* were responsive to playback. *S. kollari* was especially bold, sometimes positioning itself less than 1 metre away from the observer. Sightings (as opposed to hearings) represented 29% of *C. carbonaria* records and 46% of *S. kollari* records. We commonly recorded solitary individuals of both species (70% of *C. carbonaria* and 55% of *P.*

Table 1. Abundance of *Cercomacra carbonaria* and *Synallaxis kollari* in the Alagadiço (20 survey points) and Bridge (17 survey points) sites.

Site	<i>Cercomacra carbonaria</i>		<i>Synallaxis kollari</i>	
	<i>n</i>	<i>D</i>	<i>n</i>	<i>D</i>
Alagadiço	42 ^a	67	35	56
Bridge	44 ^a	82	34	64
Mean	43	76	34.5	60

N, number of individuals recorded; *D*, population density expressed as individuals/km².

^aTwice the number of male individuals.

kollari records) and paired individuals (27% of *C. carbonaria* and 39% of *P. kollari* records). The maximum group size was three individuals for *C. carbonaria* and five for *S. kollari*.

Geographic range limits

Table 2 presents the distribution of both species from the literature and data herein. In the 2004 survey, M. M. V. and J. B. B. recorded both *C. carbonaria* and *S. kollari* up to 111 km on the Rio Uraricoera (03°28' N, 61°09' W). The surveys detected neither species at the remaining 16 points surveyed upriver (c. 46 km). This is the most upstream record of *S. kollari* on the Rio Uraricoera to date. On 13 July 2005, M. M. V. and S. L. P. confirmed the presence of two *C. carbonaria* males at a point 3 km farther upstream than the farthest point surveyed in 2004 (03°18' N, 61°31' W), as reported by C. P. The record was on the right bank of the Rio Uraricoera, approximately 24 km upriver from the easternmost tip of Maracá Island, in an area highly disturbed by cattle ranching. This is the farthest upstream record to date for *C. carbonaria* on the Rio Uraricoera.

Table 2. Geographic range limits, available habitat, and area within conservation and indigenous reserves for *Cercomacra carbonaria* and *Synallaxis kollari*.

River	Geographic range	Length (km)	Habitat (km ²)	Area within reserve (%)
<i>Cercomacra carbonaria</i>				
Branco	Source ^a down to 01°16' S/ 61°50' W ⁽¹⁾	562	467 (11)	18% (VNP, CSP, NSP)
Mau	Mouth up to 03°42' N/ 59°40' W ⁽²⁾	80	25	54% (RS)
Mucajaí	Mouth up to 02°41' N/ 61°16' W ⁽³⁾	95	68	0
Tacutu	Mouth up to 03°34' N/ 59°53' W ⁽⁴⁾	120	36 (1)	57% (RS, MA, J)
Uraricoera	Mouth up to 03°18' N/ 61°31' W ⁽⁵⁾	169	102 (13)	65% (MA, OU, P)
<i>Synallaxis kollari</i>				
Cotingo	Mouth up to 04°10' N/ 60°32' ⁽⁶⁾	46	8	100% (RS)
Mau	Mouth up to 03°42' N/ 59°40' W ⁽²⁾	80	25	54% (RS)
Parimé	Mouth up to 03°34' N/ 60°44' W ⁽⁵⁾	46	10	70% (MA, PS, OU)
Surumu	Mouth up to 04°12' N/ 60°48' W ⁽⁷⁾	183	42	100% (RS, MA)
Tacutu	Mouth up to 03°34' N/ 59°53' W ⁽⁴⁾	120	36 (1)	57% (RS, MA, J)
Uraricoera	Mouth up to 03°28' N/ 61°09' W ⁽⁵⁾	116	71 (13)	88% (MA, OU, P)

Geographic range, species' upper and lower range limits along rivers; Length, extent of range along rivers; Habitat, area of available habitat along rivers within riverbanks and (islands).

Sources of geographic range limit data: (1) 10 km upriver from mouth of Rio Branco, Brazil (Naka *et al.* 2006), (2) Good Hope, Guyana (O'Shea *et al.* 2007), (3) Fazenda Paraense, Brazil (Santos 2003), (4) near Conceição do Mau, Brazil (Forrester 1992), (5) data herein, (6) near Contão, Brazil (Grosset and Minns 2002), (7) near Surumu town, Brazil (Grosset and Minns 2002).

Reserves: CSP, Caracará State Park; NSP, Niquiá State Park; VNP, Viruá National Park (see Figure 1 for indigenous reserves).

^aThe source of the Rio Branco is the confluence between the Rio Uraricoera and the Rio Tacutu.

On 5 September 2005, M. M. V. failed to record *C. carbonaria* or *S. kollari* individuals upriver from their previous uppermost record in the Rio Mau (Good Hope – Guyana, 3°53' N, 59°35' W; O'Shea et al. 2007). Likewise, on 14 September 2005 M. M. V. recorded no individuals upriver from the uppermost records of *C. carbonaria* in the Rio Mucajaí (Fazenda Paraense – Brazil, 2°41' N, 61°16' W; Santos 2003). Therefore, we have confirmed the known uppermost records for the species in the Rio Mucajaí and Rio Mau.

On 7 September 2005, M. M. V. recorded *C. carbonaria* individuals up to 4 km (03°25' N, 60°36' W) and *S. kollari* up to 35 km (03°34' N, 60°44' W) from the mouth of the Rio Parimé. These are the first records for both species on the Rio Parimé and represent their known geographic limits on this river.

We do not consider *C. carbonaria* to occur in the Rio Parimé and Rio Surumu, nor *S. kollari* to occur in the Rio Branco. Our study shows that *C. carbonaria* occurrence in the Rio Parimé and Rio Surumu is restricted to the area a few kilometres from the respective mouths of these rivers and, thus, should be considered within the influence of the Rio Uraricoera and Rio Tacutu, respectively. Similarly, the only *S. kollari* record in the Rio Branco is on Forte São Joaquim (J. Natterer 1836, Vienna Museum of Natural History), at about 1 km from the Rio Branco's source (the confluence of the Rio Uraricoera and Rio Tacutu), and should be considered under the influence of the Rio Tacutu.

Available habitat and global population size

Table 2 shows the amount of available habitat for *C. carbonaria* and *S. kollari* along the rivers at which they occur. *C. carbonaria* has 723 km² of available habitat (gallery forest), with only 25 km² of this within islands (Table 2). The available habitat represents 72% of the species' potential habitat (all habitat within 0.5 km from the river margin). The non-habitat is mostly covered by natural landscapes (savanna, sand beaches and water bodies), but 37% is covered by human-modified landscapes (urban areas, roads and farms). Eight per cent of the available habitat of *C. carbonaria* is within conservation units and 15% within indigenous reserves (Table 2). The combination of *C. carbonaria* available habitat (723 km²) and its population density (75 individuals km⁻²) translates into a maximum estimate exceeding 50,000 individuals. If we factor in the species' regional abundance (found in 29% of points surveyed) we have a global population size of about 15,000 individuals.

Synallaxis kollari has 206 km² of available habitat (gallery forest), with only 14 km² of this within islands (Table 2). This habitat represents only 36% of the total potential habitat within the species' range. As with *C. carbonaria*, the non-habitat is mostly covered by natural landscapes, with 32% covered by human-modified landscapes. There is no *S. kollari* habitat within conservation units but 57% is within indigenous reserves (Table 2). The combination of available habitat (206 km²) and population density (60 individuals km⁻²) translates into a maximum estimate exceeding 12,000 individuals. If we factor in the species' regional abundance (found in 44% of points surveyed) we estimate a population size of about 5,000 individuals.

Discussion

Local abundance

Few studies of population density in riverine habitats have been attempted in the Amazon. The methods we used have obvious and well-known flaws, some of which probably cancel each other out. Logistical issues limit what we could do, of course. Circular plots suffer the obvious limitation that an $x\%$ error in estimating the maximum distance at which one regularly detects the bird, yields an approximately $x^2\%$ error in density. We took care to judge whether birds were inside or outside a 100 m radius. With circular plot methods, there is always the possibility of the systematic error of some fraction of the birds present not calling or, in this case,

responding to the tape. This might be especially true for birds at the farthest distances from the observer. To reduce that error, we surveyed each site several times and took the maximum estimate. Clearly, if birds were not detected until we had surveyed each point several times, we might not detect all the birds present and so *underestimate* the correct density. In fact, for both species, for survey points that eventually held birds, we detected birds at roughly 60% on the first visit, and only 5% needed five visits to find the birds. Sixth visits detected no extra birds.

We found that the birds responded strongly to the tape and sometimes came close to it. That poses the concern that the tape might attract individuals from territories outside the 100 m survey radius: we would then *overestimate* the correct density. Our experience was that we readily heard birds up to 100 m away in their habitat and the tape did not obviously bring them towards it from large distances away. Were it to be the case, it would take several surveys at each point before birds moved in and we recorded the maximum number of birds at each point. In fact, we recorded the maximum numbers at each point at 45% of the survey points (for both species) on the first visit, needing no more than five visits for *S. kollari* and six visits only at one point for *C. carbonaria*.

Another check on the estimates is to compare them with other studies. Taken at face value, a density of 76 and 60 individuals km⁻² (roughly 38 and 30 pairs) for *C. carbonaria* and *S. kollari*, respectively, is comparable to abundance of birds in similar guilds and environments. In western Amazonia, for example, Robinson and Terborgh (1997) found that early forests along the river have significantly more common species than do mature forests inland. That these habitats were rich in arthropods explained an average density of 30 pairs (= 60 birds) km⁻² for understory insectivorous species, which is strikingly similar to our estimate, all things considered.

Kratter (1995) estimated the population density of *Formicarius rufifrons*, a threatened bird endemic to western Amazonia. The species is similar to *C. carbonaria* and *S. kollari* in having a very small geographic range and a preferred habitat restricted to the thin strip of forest along a few rivers. There are numerous methodological differences between his study and ours, but the salient difference is the rarity of *F. rufifrons* compared with either *C. carbonaria* or *S. kollari*. Kratter found only five territories during 180 days of fieldwork – a measure of how time-consuming his method was. He estimated population density as about 1 territory km⁻¹ of river length surveyed or about 3 territories km⁻².

In examining relative abundances, Parker *et al.* (1996) consider *C. carbonaria* to be “fairly common” and *S. kollari* as “probably rare” (and *F. rufifrons* to be “rare”). However, using their definition of “common” as those species that “occur throughout their ranges in moderate to large numbers and are found easily during brief periods of fieldwork”, our results suggest that both species of the present study are locally “common”.

Conservation

The greatest human threats to Roraima occur in the northern portion of the state, where the capital, Boa Vista, and most agricultural and ranching activities are located. Based on demographic and socio-economic indices, all rivers in this region, with the exception of the Rio Mau, are considered to be under pressure (Monteiro and Sawyer 2001). These include the upper Rio Branco (north of Caracaraí), Rio Uraricoera, Rio Tacutu, Rio Surumu and Rio Cotingo. The Rio Mau and the lower Rio Branco (south of Caracaraí) suffer lower pressure. Because *S. kollari* is restricted to northern Roraima, its entire available habitat is under pressure. *C. carbonaria* occupies large tracts of available habitat in the southern portion of its range that are under low pressure.

Birdlife International (2006a) lists *C. carbonaria* as Vulnerable based on its small geographical range and estimated population declines. The species is supposed to have a declining population of between 4,200 and 4,700 individuals, all within a single subpopulation (Birdlife International 2006a), based on species distribution data up to Grosset and Minns (2002). Since its initial designation, however, the known distribution of *C. carbonaria* has more than doubled through

new records by Santos (2003) on the Rio Mucajá, Naka *et al.* (2006) on the Rio Branco, O'Shea *et al.* (2007) on the Rio Mau and Tacutu, and the present study on the Rio Uraricoera and Rio Parimé. Of special relevance is the record by Naka *et al.* (2006) that extended the known distribution approximately 300 km southward, including large tracts of forests south of Caracará (previously known as the species' southernmost limit) that, as mentioned above, are under considerably less threat than those in the north (Monteiro and Sawyer 2001). *C. carbonaria* may, therefore, have greater ecological plasticity than previously thought, occurring in gallery forests to the north and short *várzea* forests to the south (Naka *et al.* 2006).

The increase in the known range of *C. carbonaria* should lower the species' inferred level of threat, as it probably possesses better population trends than previously thought, though the human threat is still present. The species should still remain on the Red List, where we suggest it be down-listed to Near Threatened, as it fulfils the relevant distribution criteria (it is restricted to a single river and its tributaries) and has only limited habitat protection within conservation units. Furthermore, our estimate of available habitat of 723 km² fulfils one only of two required criteria for Vulnerable species, namely that it have an area of occupancy smaller than 2,000 km² under continual decline (IUCN 2001). Our estimated area is about 3 times larger than that estimated by Birdlife International (2006a). It is important to note that only about 3% of *C. carbonaria* available habitat is within islands, lessening their importance for the species conservation, previously suggested by Zimmer *et al.* (1997).

S. kollari was listed as Endangered, as it was known to have a small range with records from only five localities (BirdLife International 2004). In the 2006 assessment, the species was down-listed to Vulnerable because of additional recent sightings (BirdLife International 2006b). We consider the down-listing to be premature: although *S. kollari* is more regionally abundant than previously thought, it still has a very small range and strict ecological requirements. Our estimate of 206 km² of available habitat fulfils the criterion of an area of occupancy smaller than 500 km² (IUCN 2001), which, together with a declining and fragmented habitat, makes *S. kollari* eligible for Endangered status. This small range is sufficient to raise considerable concerns about the species' survival. None of the *S. kollari* range is formally protected within a conservation unit, and all of it is within the region of highest anthropogenic pressure in Roraima. The gallery forests within the species' range are rapidly being converted into rice plantations. Zimmer *et al.* (1997) mention "some conversion to agriculture, especially rice cultivation" as a possible threat to *C. carbonaria* and *S. kollari*. In 1999, however, mechanized irrigated rice cultivation boomed in Roraima, and by 2002, the production was already 3-fold what these authors had witnessed (Cordeiro 2005). The rice plantations are concentrated on the margins of Rio Branco's main tributaries and rely heavily on fertilizers, pesticides, and herbicides (Cordeiro 2005). Local authorities have generally failed to investigate reports by local indigenous populations of occasional bird and fish die-offs near these plantations (CIR 2001, 2003). The margins of the Rio Tacutu, Rio Surumu and Rio Mau have been completely converted to rice plantations. In 2005, M. M. V. recorded 16 irrigation pipes for rice along 110 km of the Rio Tacutu (between its mouth and Conceição do Mau).

C. carbonaria and *S. kollari* were removed from Brazil's official list of threatened species (Brasil 2003) based on data deficiency (Machado *et al.* 2005). The present study provides enough new information for both species to be reconsidered under the Brazilian List of Threatened Species.

The importance of indigenous reserves

The role of indigenous reserves in the conservation of *C. carbonaria* and *S. kollari* has been overlooked. Previous authors systematically failed to acknowledge that most of the species records have been in indigenous reserves. Records from the Rio Uraricoera and Vila Surumu (Grosset and Minns 2002) are in the São Marcos Indigenous Reserve; while the records from Conceição do Mau (Zimmer *et al.* 1997, Forrester 1992), Flexal, Limão (T. D. Carter 1927,

American Museum of Natural History) and Contão (Grosset and Minns 2002) are in the Raposa-Serra do Sol Indigenous Reserve.

Most gallery forests in the northern, more disturbed, area of Roraima are within indigenous reserves. Although these are not part of the Brazilian network of conservation units, indigenous people have had a much better record of maintaining the ecosystems around them than other populations in the Amazon (Fearnside 2003, Nepstad *et al.* 2006). Rice cultivation is the main threat to *C. carbonaria* and *S. kollari* habitat. In Roraima, most rice production is carried out illegally by non-indigenous people within indigenous land (Rohter 2004). Producers have already been evicted from the São Marcos Indigenous Reserve but have not yet been evicted from Raposa-Serra do Sol, for which final legislation was completed on 13 April 2005. In the context of the conservation of *C. carbonaria*, and especially of *S. kollari*, the final legalization of Raposa-Serra do Sol Indigenous Reserve provides added protection.

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