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Trophic structure of soft-bottom macrobenthos in an inlet in north-western Spain

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The trophic composition of macrobenthic communities in intertidal and subtidal soft-bottoms of the Ensenada de San Simón (north-western Spain) was found to be related to a number of environmental variables. Distribution and abundance of trophic groups have been studied, to provide essential baseline information for monitoring the area, after its designation as a Natura 2000 Special Conservation Zone. Analyses of trophic data showed a numerical predominance of the herbivores in the inner part of the inlet, while sites at the oceanic-influenced area were numerically dominated by surface-deposit feeders. These dominances were mainly due to Hydrobia ulvae in the intertidal area and to polychaetes at the subtidal one. Both univariate and multivariate statistical analyses showed that the sediment composition (organic matter and silt-clay contents) and temperature of the bottom water influenced the benthic macrofauna and were correlated with their trophic composition, abundance and distribution.

Keywords: trophic structure, soft bottoms, benthos, macrofauna, estuary

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INTRODUCTION

The composition of sublittoral benthic macrofauna depends on interaction between biotic (food availability, colonizer larvae, predation and competition) and abiotic factors, such as salinity, temperature, depth, oxygen concentration, current speed, stochastic events or sediment grain size (Sanders, 1968; Cocito et al., 1990). Another important factor is the marine bottom relief, which affects deposition and suspension of particles, which in turn affect the trophic structure of the communities (Pearson, 2001). The effects of contamination on trophic structure can be difficult to detect in estuaries, since many opportunistic taxa show a facultative resistance to pollution stress (Rakocinski et al., 2000). Nevertheless, trophic structure has been largely studied since it is a good indicator of physical characteristics in the marine bottom environment (Lastra et al., 1991). Soft bottoms can be characterized relating the distribution of the inhabitant trophic categories with abiotic parameters (Olabarria et al., 1999). Furthermore, the morphology that species adopt for optimizing food capture, offers indirect information about the physical characteristics of the environment (Lastra et al., 1991). Sanders (1958) observed that suspension feeding organisms are typically found in high energy environments with coarse sediments, while deposit feeders are more abundant in low energy environments with high percentage of fine particles.

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The Ensenada de San Simón is located in the inner part of the Ría de Vigo, between $42^{\circ}17' - 42^{\circ}21'$ N and $8^{\circ}37' - 8^{\circ}39'$ W (Figure 1). Benthic communities of several estuaries have been studied in recent years, but despite the profusion of scientific studies in the Ensenada de San Simón (Nombela et al., 1992; Alvarez-Iglesias et al., 2003), none of them have analysed the patterns of spatial distribution of its benthic fauna or the trophic behaviour of these populations. Soft-bottoms of the Ensenada de San Simón are mainly muddy with high organic matter content (Vilas et al., 1995). Intertidal and shallow subtidal areas are colonized by the seagrasses Nanozostera noltii Horneman (Tomlison & Postuzny) and Zostera marina L. Culture of mussels on rafts is a common practice in large areas of the mouth of the inlet (Abella et al., 1996). Large freshwater input occurs in its innermost part, which translates into salinity fluctuations both tidally and seasonally (Nombela et al., 1992). These inflows, together with the numerous mussel rafts and the two harbours present in the inlet, are the main natural and anthropogenic stressors/ pressures in the inlet.

The 2252 hectares of the Ensenada de San Simón have been included in the Nature 2000 Network as a Special Conservation Zone because of the intertidal areas colonized by *Zostera marina* and *Nanozostera noltii*, which are used as resting and feeding habitats by bird populations. Research on the composition and distribution of macrobenthic communities in soft sediments is of great interest to properly manage the protected marine environments. So, there arose a need to improve the scientific knowledge of the benthic communities present in the Ensenada de San Simón. This knowledge will provide a reference baseline to analyse any possible impact



Fig. 1. Location of the Ensenada de San Simón (Ría de Vigo, Spain) and position of the 29 sampling sites, pointed line indicating the limit with intertidal level. Grey areas are mussel raft sites and black bars indicate harbour positions.

from future changes (effect of mussel rafts, overexploitation, contaminant spills and construction of new harbours).

By taking into account previous studies, we know that the community structure depends on abiotic factors (mainly grain size, calcium carbonate and organic matter composition) and that suspension feeders are dominant in sediments with larger grain size, while deposit feeders are dominant in the muddy sediments. But in the specific case of the Ensenada de San Simón, other extreme factors can determine the organization of the trophic communities. Depth, and derived desiccation stress in intertidal areas, will be hypothesized as the responsible factor for the benthic fauna distribution. The particular hydrodynamic conditions and the presence of seagrass meadows and mussel rafts will enrich organic matter content in the calmer areas of the inlet, where deposit feeders will find their optimal conditions. Our hypothesis is that deposit feeders, adapted to live in muddy sediments and related stable conditions, will be dominant in subtidal areas. Filter feeder species, adapted to coarser sediments and stressful intertidal conditions, will be present in inner parts. The presence of seagrass in intertidal and shallow subtidal areas will determine the distribution of herbivore species which feed on it and its epiphytes.

According to Rhoads & Young (1970) deposit feeders can destabilize the sediment and thus control fauna distribution on soft bottoms. This trophic amensalism hypothesis suggests that suspensivores would collapse their filtration mechanisms in the presence of deposit feeders, and so, the abundance of suspension feeders will be negatively correlated with deposit feeders. This assumption will also be tested in San Simón. Therefore, the principal aim of the present study is to characterize the trophic structure and ecology of species inhabiting intertidal and subtidal soft substrata throughout the Ensenada de San Simón and relate them with the measured environmental variables.

MATERIALS AND METHODS

Samples were collected from 29 sites during high tide in November and December 1999 (Figure 1). The map also shows the main natural and anthropogenic stressors/pressures in the inlet (harbours and rafts for mussel cultures). Five replicated samples were taken at each site, by means of a van Veen grab (0.056 m²). Samples were sieved through 0.5 mm mesh and the retained material was fixed in 10% buffered formalin. Fauna was sorted from the sediment and preserved in 70% ethanol. Temperature from bottom water and sediment and pH from sediment were measured in situ with a pHmeter Hanna HI 9025C. A sediment sample was taken at each site for later analyses. To determine content of calcium carbonate a sediment sample was treated with hydrochloric acid. Total organic matter content was estimated from the weight loss after placing samples in a furnace for 4 hours at 450°C. Granulometric fractions were determined according to Guitián & Carballas (1976) and sedimentary types according to Rodrigues & Quintinho (1985) and Junoy (1996).

Data analysis

Macrofaunal species were identified, abundance data were organized into matrices and replicate data were added to obtain a value for the whole sampling site. Each macrofaunal species was assigned to a trophic level following previously published papers (Fauchald & Jumars, 1979; Gambi & Giangrande, 1985a, b; Ambrogi et al., 1989; Lastra et al., 1991; Palacio et al., 1993; Troncoso et al., 1996; Olabarria et al., 1999; Giangrande et al., 2000). The trophic groups considered include carnivores (C), surface deposit-feeders (SD), subsurface deposit-feeders (SSD), suspension feeders (S), herbivores (H) and others (O), the latter including ectoparasites, omnivores and scavenging fauna. Several categories were defined combining the trophic and the major faunal groups (Polychaeta, Mollusca, Crustacea and others). Structure analyses were carried out with the data of the major trophic groups and at the level of these categories (e.g. suspension feeder molluscan, SM).

Median size grain (Q_{50} , mm) and sort coefficient (S_0) (Trask, 1932) were calculated for each sample. Sort coefficient was considered as $S_0 = \sqrt{Q_{25}/Q_{75}}$, where Q_{25} and Q_{75} are the 25th and 75th percentiles. Kurtosis (Kg) and skewness (Sk) coefficients were calculated according to Folk & Ward (1957).

Spearman rank correlations were used to examine relationships both between abiotic (depth, grain size, etc) and biotic variables (abundance of trophic groups) and within biotic variables (carnivores versus deposit feeders, etc) (SPSS 14.0 program) (Sokal & Rohlf, 1980).

Assemblages defined by trophic structure were determined through non-parametric multivariate techniques using the Plymouth Routines in Multivariate Ecological Research software package (PRIMER 5; Clarke & Warwick, 1994). A similarity matrix was performed using the Bray–Curtis coefficient (Bray & Curtis, 1957) after applying the fourth-root transformation to trophic abundances. From the similarity matrix, classification and ordination of the sites (direct analysis) and trophic groups or categories (inverse analysis) was performed through cluster analysis using the algorithm UPGMA and non-metric multidimensional scaling (MDS). The SIMPER program was used to identify trophic groups that contributed to the dissimilarity among the groups of previously determined sites.

Relationships between abundance of trophic groups and environmental variables were investigated by means of BIOENV procedure (PRIMER package) and canonical correspondence analysis (CANOCO for Windows 4.5 package; ter Braak & Prentice, 1988). Environmental variables expressed in percentages were previously transformed by log(x + 1) and all of them were normalized.

RESULTS

Sedimentary characterization

The soft bottoms of the Ensenada de San Simón were characterized by a predominance of muddy sediments with high organic matter and low calcium carbonate contents (Table 1). Sandy sediments were present in the eastern areas and in the tidal channels of the inner inlet, where low content of total organic matter were found. Site 26, located in the mouth of the inlet, had muddy sands with a large gravel fraction composed of shells of the mussels which were cultured in rafts and then thrown into the water after harvesting.

Composition and abundance of trophic groups

Sampling yielded 71,576 individuals, of which 29.87% of the specimens belonged to the trophic group surface deposit-feeders (SD), 28.72% were herbivores (H), 21.12% subsurface deposit-feeders (SSD), 12.77% others (O), 4.59% carnivores (C) and only 2.93% suspension feeders (S). Maximum densities were found at intertidal sites (6, 2 and 3, >18,500 individuals per m²) and minimum values in marginal areas close to the harbours (Sites 11, 24 and 29, <700 ind m⁻²) (Table 2).

The distribution of trophic groups can be seen in the maps (Figures 2 & 3). As a general pattern, herbivores were found to dominate the sandy areas of the Ensenada de San Simón, while surface and subsurface deposit-feeders dominated the subtidal muddy zone.

Herbivores presented a mean of 2531.79 individuals per m² (standard deviation, $\sigma = 7052.21$) due to high densities of the gastropod *Hydrobia ulvae* (Pennant, 1777) in intertidal areas. This species presented a total of 19,928 individuals, representing 96.94% of the trophic group to which it belongs and 27.84% of the total fauna. Maximum values of *H. ulvae*

Table 1. Depth of the water column during the sampling (metres, <2 m is intertidal) and characteristics of sediments at each sampled site.

Gravel (%) со Site Depth Sand (%) Silt-clay (%) Sedimentary type OM Q50 1.6 16.8 Mud 26.52 1 0.01 0.1 83.1 5.52 1.6 Mud 2 0.01 2.2 32.9 64.9 23.30 5.60 Sandy mud 1.6 0.08 0.00 56.7 19.05 6.12 3 43.3 4 1.6 0.32 17.8 74.0 8.2 Muddy sand 2.16 6.00 1.8 Muddy sand 5 1.25 30.0 64.3 5.7 4.90 7.33 Very coarse sand 1.6 6 1.15 76.8 0.95 11.98 21.1 2.1 7 0.15 0.3 Sandy mud 3.95 6.31 3.4 74.3 25.4 8 3.2 0.04 0.6 35.9 63.5 Mud 10.88 5.80 0.9 Mud 2.9 0.01 18.12 4.28 9 27.7 71.4 Mud 36.93 4.28 10 2.9 0.01 0.0 2.3 97.7 3.6 Mud 11 0.01 0.0 8.9 26.50 4.81 91.1 12 3.8 0.01 1.1 19.2 Mud 2.12 79.7 19.93 0.01 3.0 Mud 13 3.5 23.0 74.0 23.00 2.36 14 4.6 0.01 7.1 24.4 68.5 Mud 19.78 2.28 1.8 Coarse sand 0.74 1.00 8.35 15 3.5 94.4 2.1 16 4.2 0.01 1.0 83.5 Mud 21.47 15.5 4.53 17 3.7 0.02 4.7 31.1 64.2 Mud 18.93 5.90 18 0.01 20.0 78.1 Mud 15.20 4.5 1.9 4.52 Mud 86.1 19 4.7 0.01 0.0 13.9 21.05 4.53 11.8 Muddy sand 20 2.6 0.21 77.7 10.5 1.80 4.85 21 18 0.01 0.6 26.3 Mud 19.50 4.61 73.1 10.4 0.01 1.0 61.6 Mud 12.98 22 37.4 5.51 23 5.9 0.01 1.2 25.2 73.6 Mud 22.17 5.40 Mud 0.01 12.6 87.4 21.42 24 4.1 0.0 4.07 25 1.6 0.01 6.8 31.8 61.4 Mud 23.72 5.47 26 28.2 1.50 40.2 48.3 11.5 Muddy sand 7.22 40.46 11.5 0.01 28.2 62.5 Mud 10.60 8.61 27 9.3 28 78.6 Mud 4.61 0.01 22.32 4.7 2.4 19.0 68.2 Mud 29 2 0.01 0.1 31.7 14.33 4.45

CO, percentage of carbonate; OM, percentage of organic matter; Q₅₀, median particle size (mm).

Table 2. Total abundance data expressed as densities for each trophic group in each site (individuals per m^2).

Site	S	С	SD	SSD	Н	0	Total
1	14.3	28.6	403.6	507.1	2896.4	592.9	4442.9
2	78.6	1578.6	1032.1	2667.9	14800.0	3139.3	23296.4
3	67.9	1942.9	671.4	2389.3	10185.7	3246.4	18503.6
4	317.9	100.0	700.0	435.7	325.0	260.7	2139.3
5	707.1	514.3	3296.4	3114.3	1339.3	382.1	9353.6
6	625.0	639.3	67.9	1050.0	34964.3	589.3	37935.7
7	89.3	325.0	1467.9	4432.1	532.1	1039.3	7885.7
8	228.6	396.4	425.0	1407.1	628.6	342.9	3428.6
9	103.6	182.1	553.6	1917.9	135.7	392.9	3285.7
10	85.7	1203.6	521.4	803.6	103.6	3089.3	5807.1
11	17.9	17.9	132.1	75.0	117.9	71.4	432.1
12	50.0	35.7	532.1	100.0	3.6	171.4	892.9
13	35.7	42.9	207.1	210.7	35.7	803.6	1335.7
14	1707.1	382.1	8010.7	4100.0	235.7	2342.9	16778.6
15	182.1	717.9	1400.0	967.9	4571.4	1096.4	8935.7
16	67.9	132.1	4253.6	1875.0	85.7	214.3	6628.6
17	210.7	207.1	5189.3	1585.7	103.6	2664.3	9960.7
18	100.0	78.6	1014.3	350.0	50.0	328.6	1921.4
19	182.1	171.4	3360.7	1492.9	60.7	275.0	5542.9
20	267.9	321.4	4646.4	8746.4	60.7	482.1	14525.0
21	167.9	250.0	4135.7	1692.9	139.3	900.0	7285.7
22	800.0	625.0	10232.1	4010.7	175.0	3278.6	19121.4
23	271.4	207.1	4475.0	1550.0	139.3	1564.3	8207.1
24	7.1	14.3	303.6	71.4	17.9	250.0	664.3
25	110.7	67.9	332.1	142.9	439.3	714.3	1807.1
26	717.9	917.9	8807.1	2846.4	910.7	1946.4	16146.4
27	267.9	575.0	9246.4	4157.1	135.7	2235.7	16617.9
28	14.3	32.1	878.6	1053.6	50.0	207.1	2235.7
29	0.0	21.4	71.4	228.6	175.0	14.3	510.7

S, suspension feeders; C, carnivores; SD, surface-deposit feeders; SSD, subsurface-deposit feeders; H, herbivores; O, others.

density appeared at Stations 6, 2 and 3 (34,946.43, 14,800 and 10,153.57 ind per m² respectively).

Surface and subsurface deposit-feeders dominated in the subtidal muddy zone of the Ensenada de San Simón. Dominance of surface deposit-feeders increased from the inner part of the inlet towards the mouth, due to the higher density of the polychaetes *Pseudopolydora paucibranchiata* (Okuda, 1937), *Aphelochaeta marioni* (Saint-Joseph, 1894), Ampharetidae spp. or the crustacean *Microdeutopus* cf. *armatus* Chevreux, 1887; surface deposit-feeders also represented 37.32% of the number of species. Surface deposit-feeders presented a mean of 2633.36 ind per m² (standard deviation, $\sigma = 3082.93$). Sub-surface deposit-feeders increased their numerical dominance towards mid and external areas of the inlet.

The group others showed 25.95% of the total number of species, with 15.74% carnivores, 11.37% suspension feeders, 7.00% subsurface deposit-feeders and 2.62% herbivores.

Univariate and multivariate analysis

The abundance of suspension feeders was positively correlated (Spearman's correlation coefficient) with median particle size (P < 0.05), gravel content (P < 0.01) and abundance of surface (P < 0.01) and subsurface (P < 0.05) deposit-feeders. Carnivores showed positive correlations with abundance of herbivores (P < 0.05) and others (P < 0.01) abundances

and skewness coefficient (P < 0.05). Surface deposit-feeders showed positive correlations with depth (P < 0.01), gravel content (P < 0.05), subsurface deposit-feeders (P < 0.01) and others (P < 0.01) abundances. Subsurface deposit-feeders showed positive correlations with sand content (P < 0.05) and kurtosis coefficient (P < 0.05), and negative correlations with organic matter content (P < 0.05). Herbivores showed positive correlations with median grain size (P < 0.05) and negative with silt–clay content (P < 0.01).

The BIOENV routine had pointed out that sand-carbonate content was the combination of environmental variables that best groups the sites according to the trophic structure, but showed a very low coefficient (0.096). The combinations between bottom water temperature-sand content, bottom water temperature-sand content-sediment pH or bottom water temperature-sand and carbonate contents showed a coefficient >0.082. According to these results, it seems that the environmental variables measured do not explain the trophic structure observed in the inlet.

A dendrogram obtained through cluster analysis based on abundance data of the trophic categories showed three main groups or assemblages: A (Sites 2, 3, 10 and 15), B (Sites 5, 7, 8, 9, 14, 16, 17, 19, 20, 21, 22, 23, 26 and 27) and C (Sites 1, 4, 11, 12, 13, 24, 25 and 28) (Figure 4). A summary of the physical characteristics of these associations appears in Table 3. Ordination of sites through non-metric multidimensional analysis showed similar associations to the cluster analysis (Figure 5).

SIMPER analysis showed that the herbivores and the others were the trophic groups with a greater contribution to similarity (40%) for group A. Group B was mainly defined by surface and subsurface deposit-feeders (up to 40%), while deposit feeders and the others were the trophic groups with a greater contribution to similarity for group C.

Whenever inverse analysis was applied to the abundance data of the trophic groups, the obtained dendrogram showed high similarities between the surface and the subsurface deposit-feeders, and between suspension feeders and carnivores (Figure 6). After applying inverse analysis to abundance of trophic categories, the obtained cluster showed high affinities between the group others and the polychaetes-suspension feeders, the others-surface depositfeeders, the crustaceans-subsurface deposit-feeders and the molluscs-carnivores; another group was formed by the polychaetes-herbivores, others and carnivores, others-carnivores and the molluscs-others, suspension feeders and surface deposit-feeders. The polychaetes-deposit feeders, crustaceans-others and surface deposit-feeders and others-subsurface deposit-feeders also showed high similarities (Figure 7).

The first two axes of the canonical correspondence analysis accounted for 90.8% of the total variance of species – environment variables relationship, and 74.7% of the species variance. Eigenvalues and species – environment correlations appear in Table 4. Correlation values close to 1 indicated that abiotic variables were correctly chosen (ter Braak & Prentice, 1988). Forward selection indicated that temperature of the bottom water and kurtosis coefficient were the variables explaining most of the variance in the trophic data (P < 0.01) as well as the silt–clay and the organic matter content (P < 0.05). The scatter diagram showed an ordination of sites following these gradients on sediment characteristics (Figure 8).



Fig. 2. Abundance of individuals in the sampled sites of Ensenada de San Simón. Each trophic group considered is represented in a map: S, suspension feeders; C, carnivores; SD, surface-deposit feeders; SSD, subsurface deposit-feeders; H, herbivores; O, others.

DISCUSSION

The Ensenada de San Simón mainly presented muddy sediments with high content of organic matter. There was a predominance of deposit-feeders in subtidal muddy areas of the inlet and herbivores in the sandy intertidal ones. This distribution seems to be related with the communities cited by Cacabelos et al. (2008). In this paper, the authors found two communities in the inlet: a reduced Macoma community (Thorson, 1957) and a Syndosmya (=Abra) alba community (Petersen, 1918). The reduced Macoma community was present in the intertidal areas, where the herbivore Hydrobia ulvae was numerically dominant. The subtidal muddy bottoms were characterized by a Syndosmya (=Abra) alba community (Petersen, 1918), showing high densities of Pseudopolydora paucibranchiata, Aphelochaeta marioni or Microdeutopus cf. armatus. These last species are surface and subsurface deposit-feeders.

According to Rhoads & Young (1970), the heterotrophic benthic organisms are predominantly suspension feeders and detritivores. The suspension feeders are dominant in sandy sediments, while the surface deposit-feeders predominate in muddy bottoms (Maurer *et al.*, 1979; Gambi & Giangrande, 1985a, b; Dauvin, 1988). On the contrary, the surface deposit-feeders in the Ensenada de San Simón were positively correlated with gravel content and depth; these results contrast with previous works that pointed out that this trophic group is strongly related to silt-clay content (Sanders, 1968; Levinton, 1977; Parapar, 1991).

The organic matter and silt-clay contents found in intertidal areas were much higher than those recorded by Nombela & Vilas (1986–1987). This situation is not surprising. On the one hand the presence of phanerogams traps and stabilizes the sediment (Nombela *et al.*, 1995) and on the other hand, an intense culture of mussels on rafts is located at the study sites (Figure 1). Mussels produce high quantities of faecal pellets that substantially modify the sediment composition by increasing the clay fraction (Nombela *et al.*, 1987; León *et al.*, 2004) which in turn influences the benthic community and its trophic structure (Abella *et al.*,



Fig. 3. Dominant feeding guild for each sampled site and sediment grain size distribution in the Ensenada de San Simón (M, mud; SM, sandy mud; MS, muddy sand; CS, coarse sand; VCS, very coarse sand).

1996, Conde & Domínguez, 2004). Since the biodeposits produced by each one of the 76 rafts sited in the inlet could reach 190 kg dry weight day⁻¹ (Cabanas *et al.*, 1979), an important effect of this activity on the granulometric composition is expected. Moreover, most of the inlet is submitted to really slow hydrodynamics. Boaventura *et al.* (1999) stated that factors which affect the trophic distribution are current intensity and sediment depositions. Consequently, changes in trophic structure are expected, but no evaluation of this impact can be assessed since no data are available.



Fig. 4. Cluster analysis of trophic group abundance data using the Bray–Curtis similarity coefficient.

Table 3. Main characteristics of the assemblages determined by trophic composition in the Ensenada de San Simón, indicating means \pm standard deviations (SD) of biological (%) and physical characteristics.

	Group A	Group B	Group C	
Н	42.88 ± 27.88	4.27 ± 5.51	15.84 ± 21.16	
SD	8.18 ± 5.52	43.91 ± 17.49	33.74 ± 17.29	
SSD	12.26 ± 1.37	32.17 ± 15.54	17.79 ± 11.74	
С	11.51 \pm 6.34	4.05 ± 2.52	3.12 ± 1.39	
S	1.05 ± 0.84	3.77 ± 2.67	4.51 ± 4.46	
0	24.12 \pm 19.51	11.82 \pm 6.62	24.99 ± 17.04	
Q ₅₀	0.21 \pm 0.36	0.23 ± 0.49	0.04 \pm 0.10	
GR	1.43 ± 1.73	7.76 ± 12.31	3.69 ± 5.71	
Sand	46.60 ± 38.87	37.88 ± 20.61	25.03 ± 19.45	
Silt-clay	51.98 ± 40.06	54.36 ± 28.20	71.29 ± 25.12	
Bottom type	Mud-coarse	Mud-sandy	Mud	
	sand	mud		
Depth (m)	1.9 ± 0.6	7.5 ± 7.4	3.2 ± 1.3	
T ^a bottom water	15.2 ± 1.7	15.7 ± 1.8	15.0 ± 2.5	
T ^a sediment	14.4 \pm 0.6	15.3 ± 2.8	15.0 ± 2.5	
pH sediment	7.3 ± 0.3	7.5 ± 0.1	7.4 ± 0.2	
ОМ	20.07 ± 14.83	13.81 ± 7.22	20.09 ± 7.55	
CO3	6.09 ± 1.70	7.89 ± 9.49	4.39 ± 1.35	

H, herbivores; SD, surface deposit-feeders; SSD, subsurface deposit-feeders; C, carnivores; S, suspension feeders; O, others; Q_{50} , median grain size; GR, gravel content (%); OM, organic matter content (%); CO₃, carbonate content (%).

The gastropod *Hydrobia ulvae* is responsible for the numerical dominance of the herbivores in the intertidal area. Seagrasses and their epiphytes find optimal growth conditions here, and *H. ulvae* changes its trophic behaviour from surface deposit-feeder to herbivore within seagrass meadows (Jacobs *et al.*, 1983). Although many authors consider *H. ulvae* to be a surface deposit-feeder (Fenchel, 1977;



Fig. 5. Non-metric multidimensional scaling ordination of trophic groups in the Ensenada de San Simón.



Fig. 6. Classification of trophic groups determined in Ensenada de San Simón through inverse cluster classification analysis.



Fig. 7. Inverse cluster analysis of trophic categories data using the Bray-Curtis similarity coefficient. First letters: S, suspension feeders; C, carnivores; SD, surface-deposit feeders; SSD, subsurface deposit-feeders; O, others. Last letters: M, molluscs; P, polychaetes; C, crustaceans, O, others.

Olabarria *et al.*, 1999), in the present work it was assigned to the trophic level of herbivores. Most of the bottoms in which *H. ulvae* appeared were colonized by *Zostera marina* and *Nanozostera noltii* covered by epiphytes on which the gastropod feeds (Barnes, 1981; Levinton & Bianchi, 1981; Hootsmans & Vermaat, 1985).

The scarce carnivores of the Ensenada de San Simón (mainly polychaetes and nemerteans) were principally interstitial species, with movement and feeding behaviour which depend on the existence of interstices among sediment grains (Muniz & Pires, 1999). *Microphthalmus pseudoaberrans* Campoy & Viéitez, 1982 and *Parapionosyllis brevicirra* Day, 1954, were the most common species which appeared in intertidal areas with coarse sediment where a greater diversity of interstitial spaces is available to these small organisms (Olabarria *et al.*, 1998).

The presence of the suspension feeders was remarkable only in the mouth of the rivers Oitabén and Verdugo, at Site 14 and in the external area. This might be due to the influence of greater hydrodynamism (Nombela et al., 1987). The abundance of suspension feeders was related with the coarser fractions of sediments located in these bottoms, where there is no risk of refilling their gills (Levinton, 1977). Nevertheless, the suspension feeders did not show elevated densities in Ensenada de San Simón. The most common suspension feeder was the intertidal bivalve Cerastoderma edule, characteristic of the reduced Macoma community. Cerastoderma edule and Parvicardium exiguum (Gmelin in Linneo, 1791) display high tolerances to salinity variations, and time of emersion and current speed were the most important factors for their distribution. Among the subtidal suspension feeders Mysella bidentata (Montagu, 1803) and Calyptraea chinensis showed the highest densities in Ensenada de San Simón. These results seem to agree with the general ideas of Boaventura et al. (1999), who relate the filter feeders to current velocity, and confine herbivores to the euphotic zone.

In the inlet, the abundance of the suspensivores was positively correlated with the abundance of the surface deposit-feeders, contradicting the cited hypothesis of amensalism of trophic groups. Detritivores are generally the dominant group in sublittoral soft bottoms of estuaries and coastal ecosystems (Gaston & Nasci, 1988; Lastra *et al.*, 1991; Muniz & Pires, 1999; García-Arberas & Rallo, 2002).

Our findings suggest that sediment characteristics (grain size and organic matter content) and the bottom water temperature are the most important variables that explain the distribution of the deposit feeders. Surface deposit feeders are directly correlated with depth, showing their maximum abundances in the subtidal areas. These areas show higher and more stable bottom water temperature, a factor selected by all the multivariate analyses as an important variable explaining trophic structure. Stressful natural fluctuations impose physical limitations on the estuarine biota (Wilson & Jeffrey, 1994). So, in agreement with our first hypothesis, deposit feeders find their required stability in the subtidal zone.

The relationship between the gravel fraction and the deposit feeders can be explained by the contribution of Site 26, which surprisingly had coarse grain size sediment. The explanation for this is the high hydrodynamics of this area, located in the narrow communication path with the estuary and the important contribution of mussel shells from nearby. Muddy sediments are dominant in most of the sub-tidal area, and deposit feeders are dominant here just like on the soft bottoms of other estuaries and coastal ecosystems

 Table 4.
 Summary of canonical correspondence analysis for the Ensenada de San Simón.

Axes	Ι	II	III	IV
Eigenvalues	0.187	0.023	0.011	0.008
Species – environment correlations	0.946	0.852	0.730	0.764
Cumulative percentage variance of species data	66.4	74.7	78.6	81.4
of species-environment relation	80.7	90.8	95.5	98.8



Fig. 8. Canonical correspondence analysis ordination of environmental variables and sampled sites relative to axes I and II for the Ensenada de San Simón. Q_{50} , median grain size; S_0 , sort coefficient; GR, gravel content; S, sand content; S/C, silt and clay content; T^as, sediment temperature; T^ab, temperature of bottom water; pHs, sediment pH; OM, organic matter content; CO₃, carbonate content; Kg, kurtosis coefficient; Sk, skewness.

(Gaston & Nasci, 1988; Lastra *et al.*, 1991; Muniz & Pires, 1999; García-Arberas & Rallo, 2002).

To conclude, surface and subsurface deposit-feeders dominate in the subtidal muddy zone of the Ensenada de San Simón, while herbivores are dominant in the intertidal sandier areas. The distribution of the trophic groups seems to be primarily determined by sediment composition (organic matter and silt-clay contents) and temperature of bottom water. This study provides a very important baseline for future comparative studies, since anthropogenic-related changes in macrobenthic assemblage structure may reflect changes in trophic structure (Rakocinski et al., 1997). Mussel culture on rafts and the reduction in the hydrodynamic regime due to human (construction of dams and harbours) or natural causes (presence of meadows) will progressively increase the content of organic matter and fine particles in the sediment; both potential factors would affect the biodiversity of benthic organisms and their trophic behaviour.

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REFERENCES

Abella F.E., Parada J.M. and Mora J. (1996) Relationship between the macrobenthic community structure and the presence of mussel rafts

culture in the Ría de Vigo (NW Iberian Peninsula). *Crangon* 1, 111–118.

- Álvarez-Iglesias P., Rubio B. and Vilas F. (2003) Pollution in intertidal sediments of San Simón Bay (Inner Ría de Vigo): total heavy metal concentrations and speciation. *Marine Pollution Bulletin* 46, 491–521.
- **Ambrogi R., Bedulli D. and Occhipinti A.** (1989) Variazioni nella ripartizione tra gruppi trofici di organismi di fondo mobile nell'area marina del Delta Padano. *Oebalia* 16, 47–55.
- Barnes R.S.K. (1981) Factors affecting climbing in the coastal gastropod Hydrobia ulvae. Journal of the Marine Biological Association of the United Kingdom 61, 301–306.
- Boaventura D., Cancela da Fonseca L. and Teles-Ferreira C. (1999) Trophic structure of macrobenthic communities on the Portuguese coast. A review of lagoonal, estuarine and rocky littoral habitats. *Acta Oecologica* 20, 407–415.
- Bray R.J. and Curtis J.I. (1957) An ordination of the upland forest communities of southern Wisconsin. *Ecological Monographs* 27, 325-349.
- **Cabanas J.M., González J.J., Mariño J., Pérez A. and Román G.** (1979) Estudio del mejillón y de su epifauna en los cultivos flotantes de la ría de Arousa. *Boletín del Instituto Español de Oceanografía* 5, 45–80.
- **Cacabelos E., Gestoso L. and Troncoso J.S.** (2008) Macrobenthic fauna in the Ensenada de San Simón (Galicia, north-western Spain). *Journal of the Marine Biological Association of the United Kingdom* 88, 237–245.
- **Clarke K.R. and Warwick R.M.** (1994) *Change in marine communities: an approach to statistical analysis and interpretation.* UK: Natural Environmental Research Council, 144 pp.
- **Cocito S.S., Fanucci S., Niccolai I., Morri C. and Bianchi C.N.** (1990) Relationships between trophic organization of benthic communities and organic matter in Tyrrhenian Sea sediments. *Hydrobiologia* 207, 53–60.
- **Conde A. and Domínguez J.** (2004) Biodeposiciones del cultivo de mejillón de acuicultura en Galicia: ¿contaminante o recurso? *III Congreso Iberoamericano Virtual de Acuicultura*. http://www.revistaaquatic. com/civa2004/
- **Dauvin J.C.** (1988) Structure et organisation trophique du peuplement des sables grossiers à *Amphioxus lanceolatus–Venus fasciata* de la baie de Morlaix (Manche occidentale). *Cahiers de Biologie Marine* 29, 163–185.
- Fauchald K. and Jumars P.A. (1979) The diet of worms: a study of polychaete feeding guilds. *Oceanography and Marine Biology: an Annual Review* 17, 193–284.
- Fenchel T. (1977) Aspects of the descomposition of seagrasses. In McRoy C.P. and Helfferich C. (eds) *Seagrass ecosystems: a scientific perspective*. New York: Marcel Dekker, pp. 123–145.
- Folk R.L. and Ward W.C. (1957) Brazos river bar: a study in the significance of grain size parameters. *Journal of Sedimentary Petrology* 27, 3-26.
- **Gambi M.C. and Giangrande A.** (1985a) Analisi della struttura trofica del popolamento dei policheti nei fondi mobili di due aree del Mar Tirreno. *Oebalia* 11, 215–222.
- Gambi M.C. and Giangrande A. (1985b) Caratterizzazione e distribuzione delle categorie trofiche dei policheti nei fondi mobili del Golfo di Salerno. *Oebalia* 11, 223–240.
- **García-Arberas L. and Rallo A.** (2002) The intertidal soft-bottom infaunal macrobenthos in three Basque estuaries (Gulf of Biscay): a feeding guild approach. *Hydrobiologia* 475/476, 457–468.
- Gaston G.R. and Nasci J.C. (1988) Trophic structure of macrobenthic communities in the Calcasieu Estuary, Lousiana. *Estuaries* 11, 201–211.

- **Giangrande A., Licciano M. and Pagliara P.** (2000) The diversity of diets in Syllidae (Polychaeta). *Cahiers de Biologie Marine* 41, 55–65.
- Guitián F. and Carballas J. (1976) *Técnicas de análisis de suelos*. Santiago de Compostela, Spain: Pico Sacro, ed.
- Hootsmans M.J.M. and Vermaat J.E. (1985) The effect of periphytongrazing by three epifaunal species on the growth of *Zostera marina* L. under experimental conditions. *Aquatic Botany* 22, 83–88.
- Jacobs R.P.W.M., Hegger H.H. and Ras-Willems A. (1983) Seasonal variations in the structure of a *Zostera* community on tidal flats in the SW Netherlands, with special reference to the benthic fauna. *Proceedings of the Koninklijke Nederlandse Akademie van Wetenschappen*, Series C. 86, 347–375.
- Junoy J. (1996) La Ría de Foz, comunidades bentónicas. Servicio de Publicaciones de la Diputación Provincial de Lugo, Spain.
- Lastra M., Palacio J., Sánchez A. and Mora J. (1991) Estructura trófica infralitoral de la Bahía de Santander. *Cahiers de Biologie Marine* 32, 333–351.
- León I., Méndez G. and Rubio B. (2004) Fases geoquímicas del Fe y grado de piritización en sedimentos de la Ría de Pontevedra (NO de España): implicaciones del cultivo del mejillón en bateas. *Ciencias Marinas* 30, 585–602.
- Levinton J.S. (1977) Ecology of shallow water deposit-feeding communities Quisset Harbor, Masachusetts. In Coull B.C. (ed.) *Ecology of marine benthos*, 6. Columbia, SC: University of South Carolina Press, pp. 191–227.
- Levinton J.S. and Bianchi T.S. (1981) Nutrition and food limitation of deposit-feeders. I. The role of microbes in the growth of mud snails (Hydrobiidae). *Journal of Marine Research* 39, 531-545.
- Maurer D., Watling L., Leathem W. and Kinner P. (1979) Seasonal changes in feeding types of estuarine benthic invertebrates from Delaware Bay. *Journal of Experimental Marine Biology and Ecology* 36, 125–155.
- Muniz P. and Pires M.S. (1999) Trophic structure of polychaetes in the São Sebastião Channel (southeastern Brazil). *Marine Biology* 134, 517–528.
- Nombela M.A. and Vilas F. (1986–1987) Medios y submedios en el sector intermareal de la Ensenada de San Simón. Ría de Vigo (Pontevedra): Secuencias sedimentarias características. *Acta Geológica Hispánica* 21–22, 223–231.
- Nombela M.A., Vilas F., Rodríguez M.D. and Ares J.C. (1987) Estudio sedimentológico del litoral gallego: III. Resultados previos sobre los sedimentos de los fondos de la Ría de Vigo. *Thalassas* 5, 7–19.
- Nombela M.A., Vilas F., Alejo I., García-Gil S., Rubio B. and Pazos O. (1992). Oceanografía del transecto: Isla de San Simón-Muelle de San Adrián. Ensenada de San Simón (Ría de Vigo), NO de España. *Thalassas* 10, 77–88.
- Nombela M.A., Vilas F. and Evans G. (1995) Sedimentation in the mesotidal Rías Bajas of Galicia (north-western Spain): Ensenada de San Simón, Inner Ría de Vigo. *Special Publications of the International Association of Sedimentologists* 24, 133–149.
- **Olabarria C., Urgorri V. and Troncoso J.** (1998) An analysis of the community structure of subtidal and intertidal benthic mollusks of the Inlet of Baño (Ría de Ferrol). *American Malacological Bulletin* 14, 103–120.
- **Olabarria C., Urgorri V. and Troncoso J.** (1999) Trophic structure of the molluscan fauna in the Inlet of Baño (NW Spain): distribution, ordination and relationship to environmental parameters. *Bollettino Malacologico* 34, 87–96.
- Palacio J., Mora J., Lastra M. and Planas M. (1993) Estructura trófica de la macrofauna intermareal: evolución en un área afectada por

vertidos orgánicos. Publicaciones Especiales del Instituto Español de Oceanografía 11, 415-422.

- Parapar J. (1991) Anélidos poliquetos bentónicos de la Ría de Ferrol (Galicia). PhD thesis. University of Santiago de Compostela, Galicia, Spain.
- **Pearson T.H.** (2001) Functional group ecology in soft-sediment marine benthos: the role of bioturbation. *Oceanography and Marine Biology: an Annual Review* 39, 233–267.
- **Petersen C.G.J.** (1918) The sea-bottom and its production of fish-food. A survey of the work done in connection with the valuation of the Danish waters from 1883–1917. *Reports of the Danish Station* of Biology 25, 1–62.
- Rakocinski C.F., Brown S.S., Gaston G.R., Heard R.W., Walker W.W. and Summers J.K. (1997) Macrobenthic responses to natural and contaminant-related gradients in northern Gulf of Mexico estuaries. *Ecological Applications* 7, 1278–1298.
- Rakocinski C.F., Brown S.S., Gaston G.R., Heard R.W., Walker W.W. and Summers J.K. (2000) Species-abundance-biomass responses to sediment chemical contamination. *Journal of Aquatic Ecosystem Stress and Recovery* 7, 201–214.
- Rhoads D.C. and Young D.K. (1970) The influence of deposit-feeding organisms on sediment stability and community trophic structure. *Journal of Marine Research* 28, 150–177.
- Rodrigues A.M. and Quintinho V. (1985) Estudo granulométrico e cartografía dos sedimentos superficiais da Lagoa de Obidos (Portugal). *Comunicações dos Serviços Geológicos de Portugal* 71, 231-242.
- Sanders H.L. (1958) Benthic studies in Buzzards Bay. I. animal-sediment relationships. *Limnology and Oceanography* 3, 245-258.
- Sanders H.L. (1968) Marine benthic diversity: a comparative study. American Naturalist 102, 243–282.
- **Sokal R.R. and Rohlf F.J.** (1980) *Introducción a la bioestadística*. Barcelona: Ed. Reverté, S.A.
- ter Braak C.J.F. and Prentice C. (1988) A theory of gradient analysis. Advances in Ecological Research 18, 271–317.
- Thorson G. (1957). Bottom communities (sublittoral or shallow shelf). Memoirs of the Geological Society of America 67, 461-534.
- **Trask P.D.** (1932) Origin and environment of source sediments of petroleum. Houston, TX: Houston Gulf Publications Co.
- **Troncoso J., Urgorri V. and Olabarria C.** (1996) Estructura trófica de los moluscos de sustratos duros infralitorales de la Ría de Ares y Betanzos (Galicia, NO España). *Iberus* 14, 131–141.
- Vilas F., Nombela M.A., García-Gil E., García-Gil S., Alejo I., Rubio B. and Pazos O. (1995) Cartografía de sedimentos submarinos: Ría de Vigo. Memoria del Departamento de Recursos Naturales y Medio Ambiente (Área de Estratigrafía) de la Universidad de Vigo. Ed. Xunta de Galicia.

and

Wilson J.G. and Jeffrey D.W. (1994) Benthic biological pollution indices in estuaries. In Kramer K.J.M. (ed.) *Biomonitoring of coastal waters and estuaries*. Boca Raton, FL: CRC Press, pp. 311-327.

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