

Patterns of seed rain and seedling regeneration in abandoned agricultural clearings in a seasonally dry tropical forest in India

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Abstract: Forest recovery in abandoned pastures and agricultural fields is often impeded, therefore it is important to understand the factors limiting regeneration. Patterns of seed arrival and regeneration in five abandoned agricultural clearings nested within a seasonally dry tropical forest in India were examined along five transects radiating from the forest edge into the clearings. Wind-dispersed seeds dominated the seed arrival in clearings compared with vertebrate-dispersed seeds: 5563 wind-dispersed seeds and 1094 vertebrate-dispersed seeds of 14 and 13 tree species, respectively, were recorded. Numbers of the former declined steeply with increasing distance from the forest, whereas the latter showed no evident pattern with distance. Seeds of the invasive herb, *Chromolaena odorata*, were abundant in clearings. Although wind-dispersed seeds greatly outnumbered vertebrate-dispersed seeds, seedlings and saplings of vertebrate-dispersed species were three times more abundant than those of wind-dispersed species, indicating distinct differences in patterns of actual and effective seed dispersal. This points to recruitment limitation, and suggests that seed arrival may not be the principal barrier to regeneration in these clearings. Nonetheless, the clearings are likely to revert to forest over time.

Key Words: agricultural clearings, *Chromolaena odorata*, regeneration, seasonally dry tropical forests, seed dispersal

INTRODUCTION

Across the tropics, vast expanses of forest have been converted into pastures and agricultural fields and subsequently abandoned (Brown & Lugo 1990, Chapman & Chapman 1999). In such tropical forest clearings, seed arrival has been shown to be a critical limitation to forest recovery, with most seeds dispersed close to the forest edge; post-dispersal seed predation and seedling herbivory after germination also inhibit regeneration (Aide & Cavelier 1994, Cubiña & Aide 2001, Holl 1999, McClanahan & Wolfe 1993, Uhl 1987, Willson & Crome 1989). Existing vegetation such as remnant trees and shrubs may facilitate regeneration (Duncan & Chapman 1999, Galindo-Gonzalez *et al.* 2000, Guevara *et al.* 2004, Holl 2002, Willson & Crome 1989) whereas grasses and ferns may impede tree regeneration (Aide *et al.* 1995, Holl 1998, Slocum *et al.* 2004). Abiotic factors such

as low concentrations of soil nutrients due to several years of cultivation, and harsh microclimatic conditions compared to adjoining forest, can further impede tree regeneration (Aide & Cavelier 1994, Holl 1999, Uhl 1987).

In tropical moist and wet forests, over 80% of tree species are vertebrate-dispersed (Ganesh & Davidar 2001, Howe & Smallwood 1982, Vieira & Scariot 2006), and their dispersal is adversely affected by forest fragmentation, since animals are less likely to traverse open habitats (Aide & Cavelier 1994, Cubiña & Aide 2001, Holl 1999). In seasonally dry tropical environments, however, more than a third of all tree species are wind-dispersed (Justiniano & Fredericksen 2000, Kahn & Lawrie 1987, Parthasarathy & Karthikeyan 1997), and seed dispersal is likely to be somewhat less affected by fragmentation and opening-up of habitats (Gillespie 1999, Vieira & Scariot 2006).

Studies of forest recovery in abandoned pastures and agricultural fields have been conducted mostly in the humid tropics, with few studies carried out in the dry

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and seasonally dry tropics (Janzen 1988, Khurana & Singh 2001, Sanchez-Azofeifa *et al.* 2005, Vieira & Scariot 2006). In fact, only 3% of all tropical forest restoration studies over the last 20 y have been on dry forests (Meli 2003). We investigated patterns of seed arrival and regeneration of forest tree species in abandoned agricultural clearings within a seasonally dry tropical forest in South India. We also investigated seed arrival of *Chromolaena odorata* (Linn.) King & Robinson (Asteraceae), an invasive herbaceous perennial. We hypothesized that: (1) seed rain declines exponentially with distance from the forest-field edge for both vertebrate- and wind-dispersed species, with the pattern of decline in seed numbers for the latter being determined by seed size and weight, (2) regeneration in clearings declines with distance from forest-field edge, and (3) patterns of species composition and abundance in regenerating vegetation correspond to these patterns in seed arrival.

METHODS

Study area

Bhadra Tiger Reserve (hereafter, Bhadra, 13°22'–13°47'N, 75°29'–75°47'S) is a 493-km² reserve located in the Western Ghats biodiversity hotspot (Myers *et al.* 2000). See Teegalapalli *et al.* (2009) for a map of the study area.

The monthly mean temperature in Bhadra ranges from 10 °C to 32 °C and the relatively high annual rainfall of 2000–2540 mm is largely concentrated between June and September, followed by a pronounced dry period from October to May (Karanth 1982). The reserve is predominantly covered by southern tropical moist deciduous forest, which gradually merges with southern tropical dry deciduous forest towards its north-eastern edge (forest types *sensu* Champion & Seth 1968). The upper canopy of the moist deciduous forest is composed of species such as *Dalbergia latifolia*, *Terminalia bellerica* and *Pterocarpus marsupium*, whereas the mid-canopy is composed of species such as *Cassia fistula* Linn. (Leguminosae), *Ziziphus xylopyrus* and *Phyllanthus emblica* Linn. (Euphorbiaceae). *Hydnocarpus pentandra* Gaertn. (Flacourtiaceae), *Elaeocarpus tuberculatus* and *Trewia polycarpa* characterize the riverine semi-evergreen patches within the reserve (Karthik 2006).

Bhadra has a diverse mammalian fauna with a herbivore assemblage that includes species such as the Asian elephant (*Elephas maximus* Zimmerman), gaur (*Bos gaurus* Smith), sambar (*Cervus unicolor* Linn.), chital (*Axis axis* Erxleben), muntjac (*Muntiacus muntjak* Zimmerman) and wild pig (*Sus scrofa* Zimmerman) (Jathanna *et al.* 2003). These herbivore species were frequently observed

in the clearings formed after villages were relocated from within the reserve (K. Teegalapalli, *pers. obs.*).

In the year 2002, 457 families from 11 villages were voluntarily relocated from within Bhadra. This resulted in about 130 ha of abandoned agricultural clearings, which had formerly been under rice (*Oryza sativa* Linn.) cultivation (Karanth 2007). The present study was carried out in the abandoned village sites of Madla, Vadihaddi, Hipla, Kesavé and Karvani.

Seed rain

In each of the five clearings, five points, separated by at least 50 m, were marked at random along the forest-field edge. At each point, a transect was established perpendicular to the forest edge for recording seed rain. Seven traps were set up along these transects at distances of 0, 2, 4, 8, 16, 32 and 64 m from the forest edge, and one trap was set up at a distance of 20 m inside the forest. Since raised traps were destroyed by elephants within days of being set up, we used pits (1 × 1 m and 0.1 m deep) as seed traps, instead. The pits were lined with cloth to facilitate detection of small seeds. Seeds of tree species were collected from the seed traps in numbered bags every fortnight for a total of seven visits between 1 March and 31 May 2006. This period coincided with the season during which most trees in southern Indian deciduous forests flower and fruit (Sundarapandian *et al.* 2005). We distinguished between wind- and bird-dispersed seeds and separately recorded numbers of *C. odorata* seeds collected in these traps.

Since mammals were likely to avoid approaching the pit-traps, mammal-dispersed seeds were collected from 1 × 1-m plots offset by 5 m from each of the pit-traps. Seeds were collected monthly, for three collections between 1 March and 31 May 2006. Chances of underestimation of seed numbers due to seed predation or secondary dispersal is likely to be low since seeds observed at the first collection were present in these clearings throughout the study period (K. Teegalapalli, *pers. obs.*).

We examined patterns of seed arrival in the clearings as a function of distance from the forest-field edge. Sites were considered replicates and seed rain data from transects within each site were pooled for analysis. Seed rain density was log-transformed to remove heteroscedasticity in the data. Using distance to the forest-field edge as the predictor variable and log-transformed seed rain as the response variable we fitted linear, quadratic, cubic and exponential functions to the data and used R^2 to assess the fit of the models to the data. The model that best described the process underlying the data was determined using Akaike's Information Criterion, corrected for small samples (AIC_c, Burnham & Anderson 1998). The effect of distance from the forest edge on mammal-dispersed seeds was investigated by fitting a linear model and examining

the slope parameter and R^2 . We also used linear models to compare seed rain density and species richness values between seed traps in clearings and in the forest.

Vegetation composition and cover

Vegetation composition was recorded from nested plots along five belt transects that were laid adjacent to seed-rain transects in each clearing. Each belt transect was divided into 10×10 -m quadrats, with eight quadrats located in the clearing, and two reference quadrats located in the adjoining forest. Seedlings (individuals < 25 cm tall) and saplings (individuals 25–100 cm tall) were recorded from 250 5×5 -m plots nested within these quadrats. In addition, per cent grass cover, shrub cover, weed (*C. odorata*) cover and bare-ground cover were visually estimated in 1×1 -m plots nested within each of the 5×5 -m plots.

We expected available soil moisture to vary with distance from the forest edge, and to have a strong effect on regeneration. Soil samples (cores 5 cm in diameter and 10 cm deep) were collected from each of the 10×10 -m quadrats along two vegetation transects per site for a total of 100 cores and soil moisture was estimated gravimetrically.

We used generalized linear models (GLMs; McCullagh & Nelder 1989) to examine factors that potentially affect regeneration in the clearings. The response variable was seedling and sapling density, and the predictor variables were distance from the forest-field edge, per cent native shrub cover, per cent grass cover, per cent exotic weed cover, and per cent soil moisture, singly and in combination. We tested for multicollinearity among the potential predictor variables and ensured that variables with significant correlations were not used together as predictors in the same model. The model that best described the observed data was selected based on the lowest AIC_C value, and summed AIC_C weights were used to assess the role of individual predictors (Burnham & Anderson 1998).

RESULTS

Wind-dispersed seeds

The seed rain density in the traps in the forest was four times higher than that in clearings, averaged over all distances. However, when we fitted a linear model to the data, seed rain density did not differ considerably between forest and clearings (slope parameter = 62.4; SE = 49.3). The density of *C. odorata* seeds arriving in the traps was over seven times higher than that of native tree

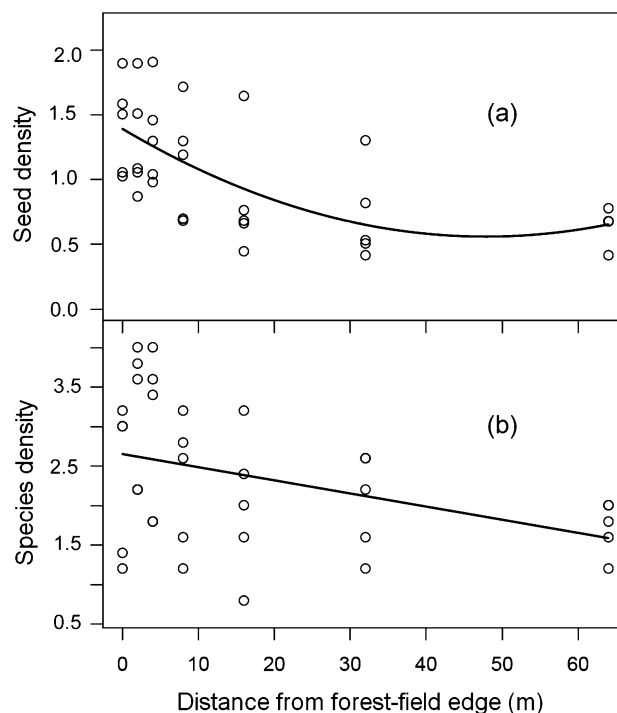


Figure 1. Decline in seed numbers and species richness with distance from the forest edge in five abandoned clearings within Bhadra, showing seed density ($\log((\text{seeds per } 14 \text{ wk m}^{-2}) + 1)$) (a), and species density (species per 14 wk m^{-2}) (b), of wind-dispersed seeds. Lines indicate functions that best describe the patterns of decline in seed density (quadratic function: $y = 0.0004x^2 - 0.035x + 1.39$, $R^2 = 0.41$) and species density (linear function: $y = -0.02x + 2.65$, $R^2 = 0.16$). Each point denotes the average seed rain across five transects within each site, over a period of 14 wk.

species (146 ± 35.5 vs. 19.2 ± 19.4 seeds per 14 wk m^{-2} , averaged over all distances).

The natural logarithm of seed rain declined with increasing distance from the forest edge and this pattern was best described by a quadratic function, based on comparison of AIC_C values for different models (Figure 1a, AIC_C weight = 0.41). Of the five most abundant species in seed rain, the seed-rain pattern of the larger-seeded species (*Pterocarpus marsupium*, *Terminalia paniculata* and *T. alata*) was best described by quadratic functions, which mirrored the overall pattern of seed rain. For the smaller-seeded species (*Lagerstroemia lanceolata* and *Dalbergia latifolia*) the seed-rain pattern was best described by linear functions (Figure 2).

The mean species density of seeds arriving in the clearing, averaged over all distances, in species per 14 wk m^{-2} of wind-dispersed seeds was considerably higher in the forest than in the clearings (slope parameter = 1.4, SE = 0.2). The decline in species density with distance from the forest edge was best described by a linear function (Figure 1b, AIC_C weight = 0.45).

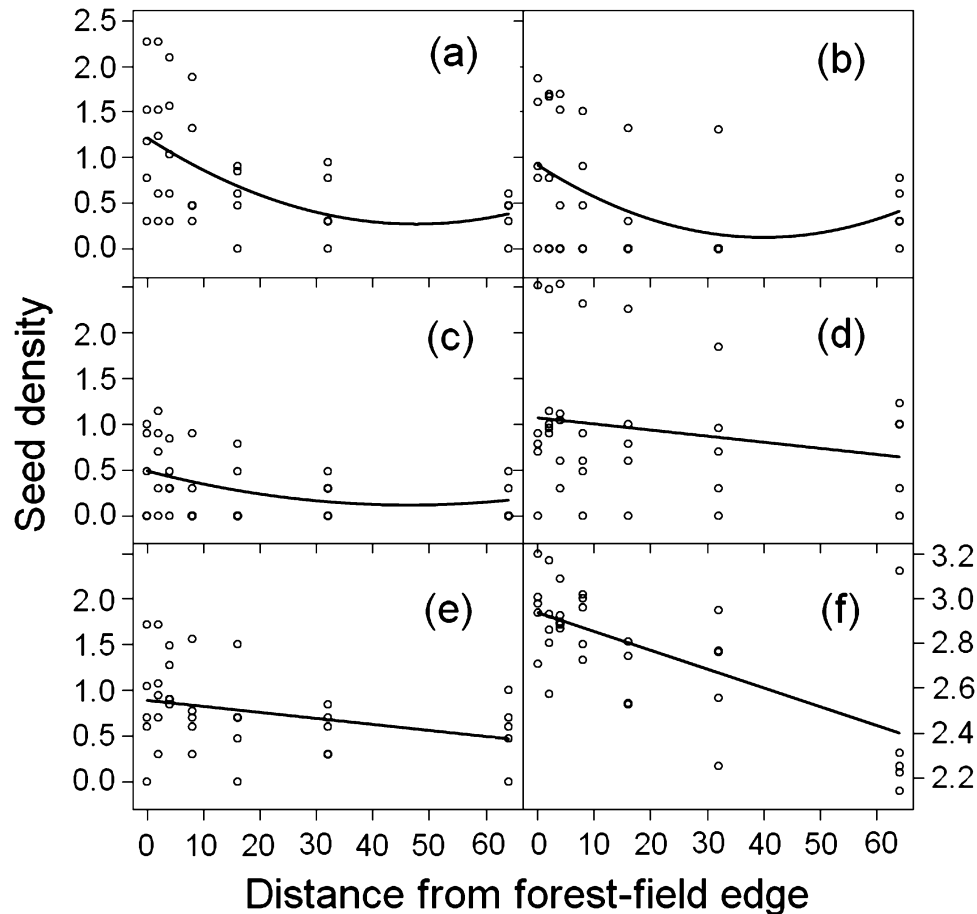


Figure 2. Seed density ($\log((\text{seeds per } 14 \text{ wk m}^{-2}) + 1)$) of the five most abundant wind-dispersed tree species as a function of distance from the forest edge in five abandoned clearings within Bhadra. Lines indicate functions that best describe the pattern of decline of seed numbers ((a) *Terminalia paniculata*: $y = 0.0004x^2 - 0.04x + 1.22$, $R^2 = 0.27$; (b) *Terminalia alata*: $y = 0.0005x^2 - 0.04x + 0.92$, $R^2 = 0.15$; (c) *Pterocarpus marsupium*: $y = 0.0002x^2 - 0.02x + 0.49$, $R^2 = 0.13$; (d) *Lagerstroemia lanceolata*: $y = -0.007x + 1.07$, $R^2 = 0.04$; (e) *Dalbergia latifolia*: $y = -0.0065x + 0.89$, $R^2 = 0.1$). Also included is the pattern of seed arrival of the invasive herb, *Chromolaena odorata* ($y = -0.008x + 2.94$, $R^2 = 0.39$) (f). Note that the ordinate is scaled differently in the *C. odorata* panel to accommodate the greater seed rain of that species.

Vertebrate-dispersed seeds

Mammal-dispersed seeds that arrived in the clearings were mostly (about 95%) of two species: *Melia dubia* and *T. bellerica* (Appendix 1). Although these seeds arrived in low densities (about 4 seeds per 14 wk m^{-2}), we observed no pattern in seed numbers with distance from the forest. This is indicated by the extremely low R^2 (0.031, $n = 35$) and slope parameter (0.003) obtained when a linear model was fitted to the data. Densities of mammal-dispersed seeds, averaged over all distances, did not show clear differences between the clearing and the forest (slope parameter = 1.7, $SE = 4.3$). This was also the case with species density of mammal-dispersed seeds (slope parameter = 0.2, $SE = 0.2$). The few bird-dispersed seeds collected (about 1 seed per 14 wk m^{-2}) were largely from a single genus *Ficus*. Thus, we use vertebrate-dispersed seeds only to refer to mammal-dispersed seeds, hereafter.

Vegetation composition and cover

A total of 447 seedlings and saplings of 47 tree species were recorded in the $250 \text{ } 5 \times 5\text{-m}$ plots. Across all sites, the mean number of individuals recorded per plot in the forest was considerably higher than in the clearings (slope parameter = 24.3, $SE = 5.6$). The number of individuals per plot in the clearings declined with increasing distance from the forest-field edge.

There was a marked increase in grass cover as a function of increasing distance from the forest edge. Weed cover in the clearings, specifically *C. odorata* cover, was highest at the forest-field edge and declined with distance from the edge, both into the adjoining forest and into the clearings. Litter cover ceased within 10 m from the forest-field edge due to the near-absence of trees in the clearings, and native shrub cover was unaffected by distance from the forest edge. Contrary to our expectation, soil moisture was also unaffected by distance

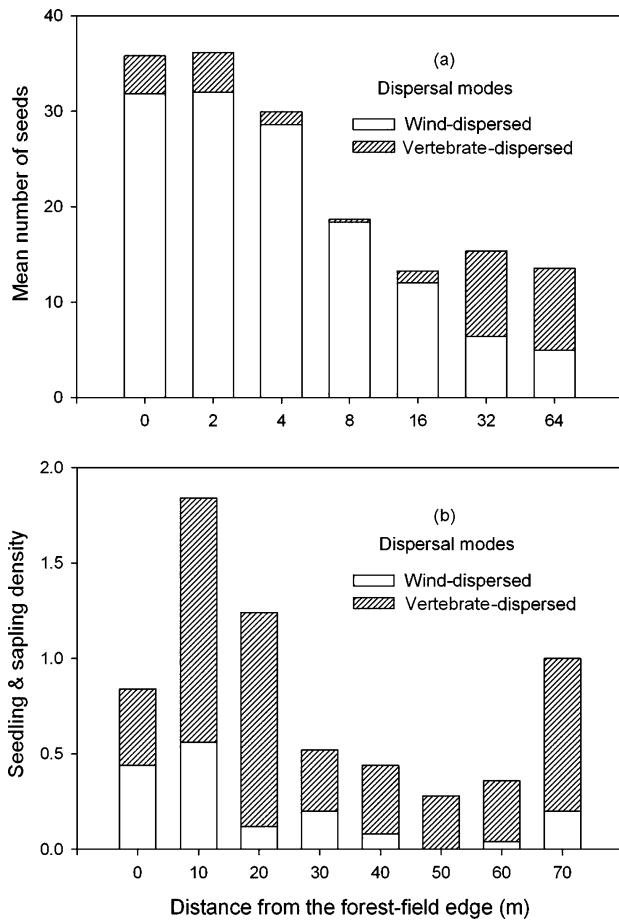


Figure 3. Comparison of seed numbers and seedling and sapling densities at different distances from the forest edge recorded from clearings in Bhadra, showing mean number of wind- and vertebrate-dispersed seeds per 14 wk m^{-2} (a), and seedling and sapling densities (individuals per 25 m^2) of wind-, and vertebrate-dispersed species (b).

from the forest-field edge (Pearson's $r = 0.115$, $P = 0.24$, $n = 40$).

Of the potential predictors of seedling and sapling densities, distance from the forest-field edge and per cent grass cover were correlated (Pearson's $r = 0.52$, $P = 0.001$, $n = 40$) and were therefore never included in the same model. Distance from the forest-field edge (summed AIC_C weight = 0.39) was the strongest determinant of mean seedling and sapling density amongst the predictors examined, followed by per cent shrub cover and per cent weed cover (0.17 in both cases), per cent soil moisture (0.16) and per cent grass cover (0.10).

Species composition of seeds and seedlings

Wind-dispersed seeds, compared to vertebrate-dispersed seeds, dominated the seed rain in clearings. However, vertebrate-dispersed species were predominant among seedlings and saplings (Figure 3) (0.66 ± 0.48 vs. 0.2 ± 0.1).

Overall, we encountered seeds of 14 wind-dispersed species across the clearings, but only five wind-dispersed species in the regenerating vegetation. Of the five most abundant wind-dispersed species in the seed rain, two were also the most abundant species in the regeneration (Figure 4). The one wind-dispersed species found in the regenerating vegetation (*Bauhinia malabarica*) that was not amongst the seeds recorded in seed rain most likely fruited after our study period (Mishra *et al.* 2006).

In the case of vertebrate-dispersed species, we encountered 13 species as seeds, but 15 as seedlings and saplings. Of the five most dominant species in the seed rain, only two also occurred as seedlings, and only one of these was amongst the five most abundant species in the regeneration (Figure 4). Seedlings of some vertebrate-dispersed species we recorded (e.g. *Psidium guajava*, *Mangifera indica*) are likely a legacy of past human settlements in these clearings.

DISCUSSION

Seed rain and regeneration in clearings

As expected, wind-dispersed seeds showed an exponential decline with distance from the forest edge, and a quadratic equation best described the pattern of seed arrival. Most seeds ($\sim 90\%$) arrived within 30 m from the forest, and only very few seeds ($\sim 8\%$) occurred as far as 64 m into the clearings. As expected, seed size and weight influenced the distance to which seeds dispersed, with a more rapid decline in number of heavier seeds compared to lighter seeds, with distance. Other studies also report short-distance dispersal of wind-dispersed seeds. Seeds found farther into clearings tend often to be seeds dispersed from remnant trees in abandoned pastures (Cubiña & Aide 2001, Holl 1999).

Vertebrate-dispersed seeds are also reported to be found close to adjoining forest; in pastures adjoining neotropical forests, most vertebrate-dispersed seeds were restricted to about 10 m from the forest (Aide & Cavelier 1994, Cubiña & Aide 2001, Holl 1999). In contrast to this, and contrary to our expectations, we found no clear effect of distance on the arrival of vertebrate-dispersed seeds. This is most likely due to the difference in faunal composition between our study site and the neotropical forests of most other studies. A large proportion of seeds (about 95%) that we recorded were the relatively large seeds of *M. dubia* and *T. bellerica*. These are eaten by large ungulates like sambar and chital which do not occur in the neotropics. These mammals take refuge in forest clearings, which afford them better visibility to avoid predation by large carnivores such as the tiger (Karanth & Sunquist 1992, Schaller 1967).

Harsh micro-climatic conditions could be a barrier to seed germination and seedling survival and the effect of harsh micro-climatic conditions in clearings may be

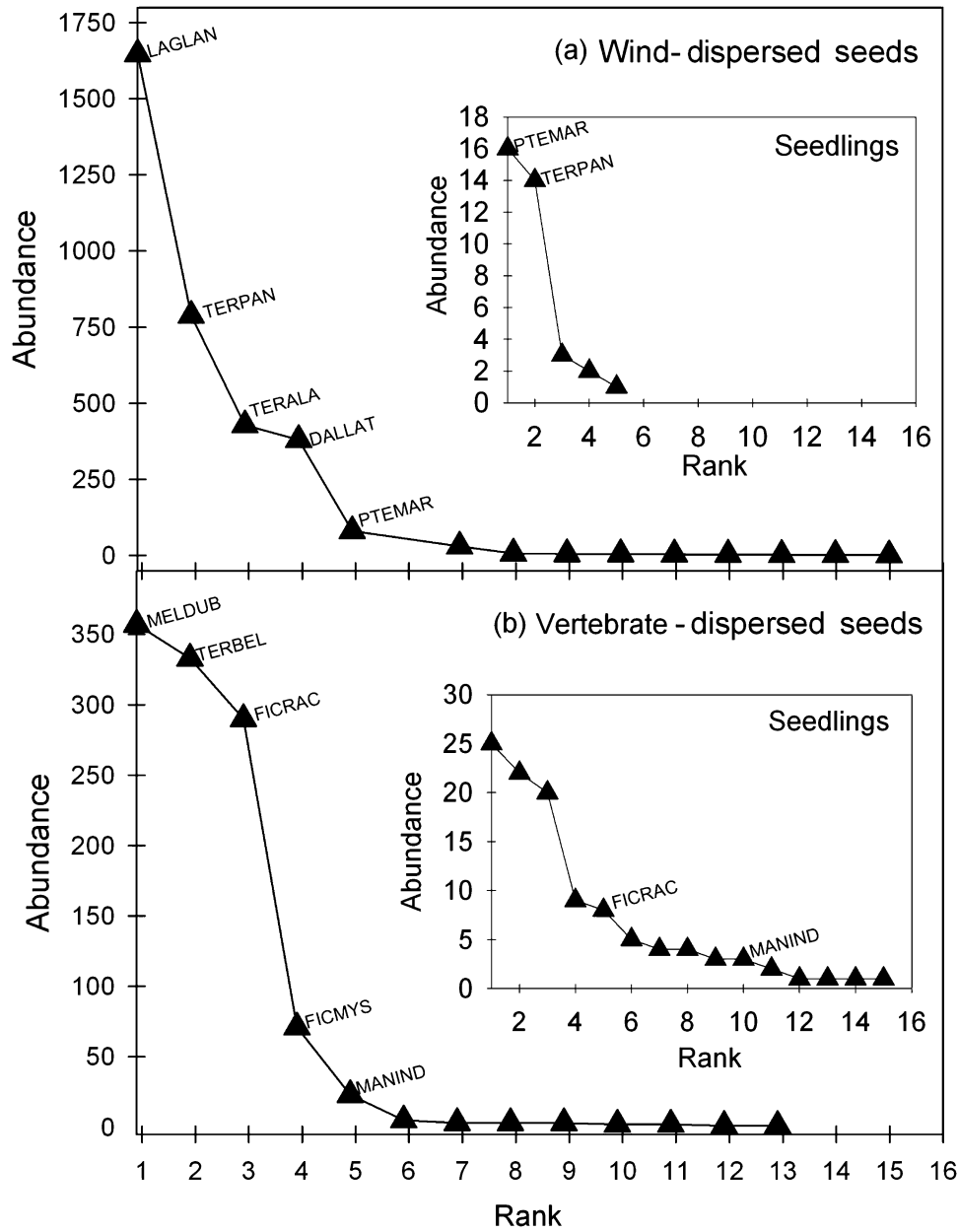


Figure 4. Rank-abundance curves for seeds and seedlings of wind- and vertebrate-dispersed species recorded from clearings in Bhadra over a period of 14 wk. Panels denote seeds of wind-dispersed species (seedlings, inset) (a), and seeds of vertebrate-dispersed species (seedlings, inset) (b). Labels indicate the five most abundant species in the seed rain, and their relative position as seedlings. (LAGLAN = *Lagerstroemia lanceolata*, TERPAN = *Terminalia paniculata*, TERALA = *Terminalia alata*, DALLAT = *Dalbergia latifolia*, PTEMAR = *Pterocarpus marsupium*, MELDUB = *Melia dubia*, TERBEL = *Terminalia bellerica*, FICRAC = *Ficus racemosa*, FICMYS = *Ficus mysorensis*, MANIND = *Mangifera indica*.)

mitigated closer to the forest (Duncan & Duncan 2000). We also found that distance from forest edge was the most important determinant of regeneration in clearings, with numbers of seedlings and saplings declining with distance from the edge. Per cent shrub and weed cover were other variables that were negatively correlated with seedling and sapling density. *Chromolaena odorata*, especially, can form dense stands, preventing regeneration of native species and promoting fires when dry (Yadav & Tripathi

1981). Others have found a similar effect of grasses and ferns in inhibiting regeneration (Aide *et al.* 1995, Holl 1998, Slocum *et al.* 2004).

Actual and effective dispersal

Contrary to our expectation, there were marked differences between patterns of actual seed dispersal (i.e.

seeds arriving in clearings) and effective seed dispersal (seedlings recorded in clearings) (Cain *et al.* 2000), both with distance from forest edge (Figure 3) and in terms of species composition (Figure 4). Especially in the case of wind-dispersed seeds arriving in the clearings, there appears to be strong recruitment limitation, with very few species occurring in regenerating vegetation. This could be due to one of several post-dispersal factors limiting regeneration, e.g. lack of suitable microsites, or seed predation or mortality (Münzbergová & Herben 2005). In our study, wind-dispersed species accounted for only 25% of total seedlings encountered. This is in contrast to findings from abandoned pastures in a dry forest in Costa Rica, where almost 90% of regeneration was composed of wind-dispersed species (Janzen 1988). This difference could be attributed to the difference in spatial extent of the Costa Rican pastures (which ranged from less than 1 ha to several km² in area) compared with the clearings in Bhadra (15–95 ha), as well as to differences in vertebrate frugivore assemblages in Bhadra, compared with Costa Rica.

Our findings are similar to those of Ingle (2003) who also found that in a 5-y-old fallow following slash-and-burn agriculture in montane rain forest of the Philippines, vertebrate-dispersed species exceeded wind-dispersed species in regenerating vegetation, even though wind-dispersed species far outnumbered vertebrate-dispersed species in terms of seed arrival. The predominance of seedlings of vertebrate-dispersed species could be due to one of several factors, for example, directed dispersal (Wenny 2001), through increased dormancy (Willson 1993), or through the protection afforded to seeds by their being deposited in faecal matter (Traveset 1998) or in clumps (Howe 1989). We did not, however, observe seedlings of *M. dubia* or *T. bellerica*, the two species that were dispersed in clumps in sambar and chital dung in our study.

Implications for vegetation recovery

In Bhadra, although seed arrival is limited farther into clearings, regeneration is proceeding relatively rapidly in comparison with regeneration in abandoned or degraded land in other parts of the world. Our findings also indicate that, overall, seed arrival may not be the principal limitation to vegetation recovery and that other factors, such as existing vegetation, may play a significant role in limiting recruitment. Experiments involving seed or seedling addition and habitat manipulation are required to clearly ascertain dispersal and recruitment limitation (Münzbergová & Herben 2005, Nathan & Muller-Landau 2000).

In the future, regeneration in the clearings we studied is also likely to be increasingly inhibited by the invasion of

C. odorata. The large number of *C. odorata* seeds collected in seed traps, regardless of distance from the edge, and the negative effect of weed cover on regeneration that we observed, are both indicative of this. Invasion by *C. odorata* in other sites across the state of Karnataka in India has been dealt with by uprooting, or cutting and burning (Doddamani *et al.* 2001). Continued monitoring of these changes in the clearings, with or without management interventions, will prove to be invaluable in understanding the dynamics of forest recovery in seasonally dry tropical forests.

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Appendix 1. List of native tree species, seeds of which were collected from seeds traps in this study in Bhadra. Also given are the locations where seeds of each species were encountered (C = in clearings; F = in the adjoining forests), and the species dispersal mode (B/M = bird/mammal dispersed, W/G = wind/gravity dispersed).

Tree species		Clearing/Forest	Mode of dispersal
<i>Mangifera indica</i> Linn.	Anacardiaceae	C/F	M
<i>Spondias mangifera</i> Willd.	Anacardiaceae	C	M
<i>Alstonia scholaris</i> Linn.	Apocynaceae	C/F	W
<i>Stereospermum personatum</i> (Hassk.) Chatterjee	Bignoniaceae	C	W
<i>Terminalia bellerica</i> (Gaertn.) Roxb.	Combretaceae	C/F	M
<i>Terminalia alata</i> Heyne ex Roth.	Combretaceae	C/F	W
<i>Terminalia paniculata</i> Roth.	Combretaceae	C/F	W
<i>Elaeocarpus tuberculatus</i> Roxb.	Elaeocarpaceae	F	M
<i>Flueggea leucopyros</i> Willd.	Euphorbiaceae	C	B/M
<i>Trewia polycarpa</i> Benth.	Euphorbiaceae	F	M
<i>Albizia lebbek</i> Linn.	Leguminosae	C/F	W
<i>Albizia odoratissima</i> (L.f.) Benth.	Leguminosae	C/F	W
<i>Butea monosperma</i> (Lam.) Kuntze	Leguminosae	C/F	W
<i>Cassia</i> spp.	Leguminosae	C/F	W
<i>Dalbergia latifolia</i> Roxb.	Leguminosae	C/F	W
<i>Erythrina indica</i> Lam.	Leguminosae	C/F	W
<i>Erythrina</i> spp.	Leguminosae	C/F	W
<i>Pongamia pinnata</i> Linn.	Leguminosae	C/F	W/G
<i>Pterocarpus marsupium</i> Roxb.	Leguminosae	C/F	W
<i>Lagerstroemia lanceolata</i> Wall.	Lythraceae	C/F	W
<i>Melia dubia</i> Cav.	Meliaceae	C/F	M
<i>Artocarpus hirsuta</i> Lamk.	Moraceae	C	M
<i>Ficus mysorensis</i> Heyne ex Roth.	Moraceae	C/F	B/M
<i>Ficus racemosa</i> Linn.	Moraceae	C/F	B/M
<i>Ficus religiosa</i> Linn.	Moraceae	C/F	B/M
<i>Syzygium cumini</i> Linn.	Myrtaceae	C	B/M
<i>Ziziphus xylopyrus</i> (Gaertn.) Roxb.	Rhamnaceae	C/F	B/M
<i>Atalantia monophylla</i> (Roxb.) Corr.	Rutaceae	C	B/M
<i>Grewia tiliifolia</i> Vahl	Tiliaceae	C	B/M
<i>Gmelina arborea</i> Roxb.	Verbenaceae	C	B/M
<i>Tectona grandis</i> Linn.	Verbenaceae	F	W