

# Collateral benefits from public and private conservation lands: a comparison of ecosystem service capacities

A. VILLAMAGNA<sup>1\*</sup>, L. SCOTT<sup>2</sup> AND J. GILLESPIE<sup>2</sup>

<sup>1</sup>Department of Fish and Wildlife Conservation, Virginia Tech, Blacksburg, VA, 24061, USA and <sup>2</sup>Environmental Science Program, Department of Crop and Soil Environmental Science, Virginia Tech, Blacksburg, VA 24061, USA

Date submitted: 24 April 2014; Date accepted: 19 October 2014; First published online 8 January 2015

## SUMMARY

Protected areas remain the most commonly used tool for *in situ* conservation; however growth in the USA's system of public lands has stagnated while private land conservation continues to expand. Easements can provide a range of ecosystem services (ESs), but it is unknown whether conservation easements maintain ES capacities equivalent to public protected areas. Evaluation of the capacity of seven ESs on federal and state protected areas and conservation easements in the USA using spatially-explicit ES models and publicly available data indicated that ES capacities in easements were equal to or greater than capacities within state or federal protected areas for six of seven services and, when bundled together, conservation easements protected greater focal ES capacity than other conservation areas. Economic incentive programmes and regulatory mechanisms may be used to stimulate capacity improvements for surface water regulation, riparian filtration, erosion control, and carbon storage on conservation easements, and landscape-level conservation efforts should (1) continue to protect natural and uninhabited areas that provide ecosystem and biological diversity, (2) expand private conservation efforts close to human population centres, and (3) limit future development to areas with high regulating service capacity that can sustain new population growth.

**Keywords:** conservation easements, geographic information systems, positive externalities, protected areas

## INTRODUCTION

Conservationists are reassessing the rationales and outcomes of conventional approaches to conserving biodiversity. Since the 1800s, the USA and much of the world has focused on large-scale government land acquisition as a means to prevent land conversion and protect natural resources for the national

good (Merenlender *et al.* 2004). Protected areas remain the most commonly used tool for *in situ* conservation (Reyers *et al.* 2012) and currently cover 12% of the global land surface (Chape *et al.* 2005). Even so, growth in protected areas and protective management policies has not slowed the rate of global biodiversity loss (Butchart *et al.* 2010).

Many of the large protected areas in the USA were designed and acquired with little regard for protecting biodiversity, imperilled species or critical ecosystem services (ESs). About 95% of imperilled species habitat in the USA occurs on private land (Turner *et al.* 2006) and most ecosystem types do not occur on public protected areas (PPAs, such as national and state parks and forests) (Scott *et al.* 2001). However, PPAs protect land from conversion and most highly degrading uses, and they therefore also protect a suite of ESs that potentially convey great benefit to society. These benefits vary in scale from local services, like surface water regulation, to global services, like carbon sequestration and climate regulation. For example, the United States (US) National Forest system is thought to generate 14% of the USA's clean water, valued at US\$ 3.7 billion per year (Dissmeyer 2000). Likewise, at least US\$ 26.9 billion worth of ESs are generated annually within the US National Wildlife Refuge system (Ingraham & Foster 2008). A fuller understanding of the magnitude, scale and beneficiaries of these ESs can help bolster public support to maintain existing PPAs and acquire new ones, which, in turn, will contribute further to biodiversity conservation.

Growth in the USA's system of public lands has stagnated since 1970, despite substantial growth at the global scale (USGS [US Geological Survey] 2012). The lack of PPA growth in the USA may be attributed to greater focus on regulatory conservation strategies, such as the Clean Water Act and Endangered Species Act, non-environmental spending priorities, including extensive military and diplomatic involvement overseas, and expanding residential development, especially in the south-eastern USA. For example, national forests in Virginia and North Carolina are expected to experience residential development on 10–42% of adjacent private lands over the next 30 years (Stein *et al.* 2007). Further, the downgrading, downsizing and degazetting of protected areas is becoming an increasing concern in some areas (Mascia & Pailler 2011), which may have residual effects on ESs and human well-being.

Given such trends, conservationists increasingly recognize that the approach to conservation planning may require a shift.

\*Correspondence: Dr A. Villamagna, Department of Environmental Science and Policy, Plymouth State University, Plymouth, NH 03264, USA e-mail: [amvillamagna@plymouth.edu](mailto:amvillamagna@plymouth.edu)

\*Supplementary material can be found online at <http://dx.doi.org/10.1017/S0376892914000393>

Proposed changes include rather extreme measures, such as degazetting conservation areas that are found to be ineffective and replacing them with areas of greater cost effectiveness (Fuller *et al.* 2010), and more subtle enhancements to the network by increasing private land conservation. Conservation easements on private lands are a common and growing land management tool in the USA, as well as in the UK (Jackson & Gaston 2008), Colombia (for example civil society private reserves; Bonnells 2012), Australia (Figgis 2004; Gordon *et al.* 2011), South Africa (Von Hase *et al.* 2010), and Costa Rica (Langholz & Lassoie 2001). Compared to PPAs, easements are distinctive in spatial distribution, parcel size, purpose, and function. Since 1990, the area of easements in the USA has grown from 202 342 ha to > 12 million ha (NCED [National Conservation Easement Database] 2013). For many landowners in the USA, the impetus to establish easements in the USA is typically to protect personal landowner values rather than to conserve biodiversity or public benefits. Nevertheless, easements have great potential to provide a range of benefits for people other than the landowners, especially in areas where PPAs do not exist or are impractical to develop. To date, it is largely unknown whether conservation easements, which are typically smaller and more diffuse, maintain ES capacities (per unit area) equivalent to PPAs.

To compare ES capacity between PPAs and conservation easements and evaluate the potential for conservation easements to provide ES protection, we evaluated the service capacity of federal and state protected areas and conservation easements in the mid-Atlantic and south-eastern USA. Herein, we define capacity as the ability of the landscape to provide an ES. This is considered the maximum potential ES that could be provided, rather than a measure of the flow of the service, which varies with societal demand and ecological pressure (Villamagna *et al.* 2013a). Using spatially-explicit ES models and publicly available data on land conservation, we mapped and compared capacity measures of seven ESs. We focused on a diverse suite of local, regional and global services: surface water regulation, groundwater protection, (surface) water quality regulation, erosion control, recreational fishing, carbon storage, and biodiversity support. This assessment will provide a current assessment of ES conservation in the south-east USA and help identify areas of future conservation investment.

## METHODS

### Study region

We focused our comparison of ES capacity in public and private conservation areas within the mid-Atlantic and south-east USA, namely Virginia (VA) and North Carolina (NC). This region, which includes four ecoregions (namely the Mid-Atlantic Coastal Plain, South-eastern Plains, Piedmont, and Blue Ridge Mountains) (Griffith *et al.* 2007), provides a relevant current synopsis of public and private conservation within the USA, where private land development is increasing

adjacent to public lands and where public land acquisition has slowed. Although the spatially-explicit models we use to quantify ESs are highly transferable across landscapes because they use nationally available public data, they were initially developed for use in the Albemarle-Pamlico Basin which, being located within the same region, comprises the same diversity of geophysical and biological features.

### Mapping conservation areas

Conservation easement boundaries were derived from the National Conservation Easement Database which contains records of more than 95 000 easements covering more than 7 million ha (18 million acres) in the USA (NCED 2013). The database provides information regarding the holders of the easement, the landowner type, the purpose of the easement, and the location. We constrained our study to only include easements with the stated purpose of: environmental systems, recreation and education, open space forest, and open space farm and with gap status code (namely a measure of the management intent for the long-term protection of biodiversity) of status 1 (permanent protection from land conversion and maintenance of a natural disturbance regime), status 2 (protected from land conversion, but allows uses or management practices that may 'degrade the quality of existing natural communities, including suppression of natural disturbance'), and those with unknown status (USGS [US Geological Survey] 2012). We excluded easements with historical preservation, unknown purpose and those preserving open space without specifying forest or farm. The remaining 230 easements are held by non-governmental organizations (NGOs), and state and federal government agencies.

We used geospatial boundary data for state and federal lands provided by the USGS National Inventory Protected Areas Database of the US (PADUS; USGS 2012). This database provides extensive information about each protected area, including land owner, management type, the purpose, and gap status. To focus on lands with an environmental conservation purpose, we restricted our analysis to state and federal lands with a gap status classification code of 1 or 2. We excluded federal land holdings that were classified as memorial parkways, national battlefields, national monuments, and several other classes that did support environmental conservation. Easements managed by state and federal agencies were excluded to avoid overlap with NCED records.

### Capacity

We used geospatial models to quantify the mean capacity of seven ES for each conservation easement and federal and state PPA in NC and VA. Geospatial capacity models were initially developed for ES analyses in the Albemarle-Pamlico basin

(Villamagna & Angermeier 2015), which flows from central and south-west VA to the eastern shore of NC.

Surface water regulation includes the hydrologic and biological processes that prevent precipitation from becoming surface run-off. These processes include interception, infiltration and uptake. The surface water run-off algorithm published by the US Department of Agriculture Natural Resources Conservation Service (NRCS 1972) estimates expected run-off based on precipitation and a curve number that integrates the soil's hydrologic group classification and land cover. We developed and applied a geographic information system (GIS) model that integrated these factors using the curve number approach to map and summarize daily surface water run-off from each conservation area (see supplementary material, Table S1). Since surface water regulation implies the abatement of run-off, we calculated the standardized inverse of summary run-off values for each conservation area so that values closer to 1 suggest greater regulating capacity.

To evaluate groundwater protection, we developed a GIS model that calculates the mean New York nitrate leaching index (Czymmek *et al.* 2003) for each conservation area (see supplementary material, Table S1). The leaching index integrates soil hydrological group classification and monthly precipitation amounts to calculate a percolation and seasonality index. Given a consistent soil type, areas with less rainfall in winter, when evapotranspiration is low, have a lower probability of leaching nitrate to groundwater. These areas are considered to have a high groundwater protection capacity. Like the surface water regulation service, we calculated the standardized inverse of the leaching index for each conservation area, in which values closer to 1 suggest greater regulating capacity.

Riparian filtration was evaluated using a GIS model we developed to assign the land cover-based per cent effectiveness values of Mayer (2007) to land cover parcels within the riparian area of all surface waters. Land cover-based per cent effectiveness values range from 0 for barren land and cropland to 85 for woody wetlands. Conservation areas that lacked surface waters were not included in the riparian filtration assessment. Thus, the riparian filtration comparison was based on a subset of conservation areas that contained or were immediately adjacent to surface water features, including 263 state PPAs, 351 federal PPAs, and 71 private conservation easements.

Mapping erosion control required the development of an intricate model based on the principles and equations of the NRCS revised universal soil loss equation (NRCS 2003) and Lim *et al.* (2005). This equation was designed to integrate five key factors that influence soil loss: soil erodibility, rainfall erosivity, the slope angle and length, land cover, and management practices to keep soil on site (see supplementary material, Table S1). We evaluated the likelihood that erosion would occur by assuming all conservation areas practised the same erosion best-management practices (such as silt fences). Erosion control was measured by calculating the standardized inverse of estimated total soil loss ( $t\ ha^{-1}\ yr^{-1}$ ).

We used predictive measures of soil organic carbon (SOC) and above and below ground carbon (biomass) to calculate total carbon storage per hectare. Land cover-based SOC values were derived from the NRCS Rapid Assessment of US Soil Carbon (West *et al.* 2013). The SOC values in this report are categorized by major land resource areas (MLRAs). Therefore, we developed a model that assigned SOC values to specific land cover classes within each MLRA according to the NRCS report. Forest carbon stocks of the contiguous USA in Wilson *et al.* (2013) provided estimates for above ground and below ground carbon storage capacity within the woody biomass of forests through the contiguous USA. We summed these stocks and multiplied values by the area of each conservation area to produce a total carbon estimate for each conservation area in tonnes (t). However, we must again acknowledge the carbon storage and sequestrations are different metrics. Although carbon sequestration is the active measure of carbon uptake from the atmosphere, protecting areas of high storage may be of equal or greater importance to climate regulation.

Biodiversity support included the predicted species richness of birds, amphibians, reptiles, and mammals throughout NC and VA. This data was derived from the NC and VA Gap Analysis Projects and included mammalian, reptilian, amphibian, and avian diversity (USGS 2013). Species richness was summarized for all taxa and the maximum species richness value within a given conservation area was chosen as the biodiversity support capacity value.

Freshwater recreational fishing capacity was assigned to each conservation area based on the results from a previously conducted assessment at the 12-digit hydrologic unit-scale for all of NC and VA (Villamagna *et al.* 2014) within which capacity was calculated using a multi-indicator framework that included biophysical (such as water availability) and social factors known to contribute to the fishing experience (for example boat ramps). The freshwater recreation fishing index ranges from 0 to 1 and the values assigned to PPAs and easements represent a weighted average based on the value assigned at the 12-digit hydrologic unit watershed scale. We used the data (Table 1) to map ES capacity throughout the region.

We summarized capacity for each conservation area by calculating: (1) the area-weighted mean of index-based metrics (groundwater protection and freshwater recreational fishing), (2) percentage-based metrics (riparian filtration), and (3) model-derived metrics (surface water regulation, biodiversity support, erosion control, and carbon storage). For services in which the capacity metric reflects the composite of several data factors with different units of measure (such as freshwater recreation fishing) and for regulating services (groundwater protection, surface water regulation, and erosion control) for which we infer regulation from the inverse of leaching, surface run-off, and erosion estimates (for example soil loss derived from the revised universal soil loss equation is the inverse of erosion control), we standardized capacity metrics to range from 0 for the lowest relative capacity

**Table 1** Summary of data inputs, models, and equations used to generate ecosystem service capacity maps for Virginia and North Carolina, USA. <sup>1</sup>Surface water regulation is inferred from taking the inverse of surface water run-off; <sup>2</sup>according to Villamagna and Angermeier (2015); <sup>3</sup>calculated from 30-year normal monthly precipitation (30 m) (PRISM Climate Group 2010); <sup>4</sup>data provided by SSURGO (30 m) (Soil Survey Staff 2010); <sup>5</sup>US national land cover dataset (30 m) (Fry *et al.* 2011); <sup>6</sup>US national hydrography dataset (NHD) flowline and waterbody (USDA *et al.* 2010); <sup>7</sup>according to Villamagna *et al.* (2014); <sup>8</sup>USGS national inventory protected areas database of the US (USGS 2012); <sup>9</sup>USGS (2013) Gap analysis programme species data (30 m) (USGS 2013); <sup>10</sup>USGS national elevation dataset (30 m) (Gesh *et al.* 2007); <sup>11</sup>revised universal soil loss equation 1.06: bulletins (USDA Agriculture Research Service 2007).

<i>Ecosystem service</i>	<i>Data inputs</i>	<i>Models and/or equations</i>
Surface water regulation <sup>1,2</sup>	Daily precipitation <sup>3</sup> , soil hydrologic group <sup>4</sup> , land cover <sup>5</sup>	Surface run-off estimator (NCRS 1972)
Groundwater protection <sup>2</sup>	Monthly precipitation <sup>3</sup> , soil hydrologic group <sup>4</sup>	New York nitrate leaching index (Czymmek <i>et al.</i> 2003)
Riparian filtration <sup>2</sup>	Surface water <sup>6</sup> , land cover <sup>5</sup>	Mayer (2007)
Freshwater recreational fishing <sup>2,7</sup>	Surface water <sup>6</sup> , land cover <sup>5</sup> , fish species diversity, water quality impairments, fish stocking, boat ramps, public use areas <sup>8</sup> , agency-supported fishing spots, watershed boundaries (hydrologic unit codes)	Villamagna <i>et al.</i> (2014)
Biodiversity support	Species richness <sup>9</sup>	Gap analysis programme: species data (USGS 2013)
Carbon storage	Soil carbon and below and above ground carbon stocks	NRCS rapid assessment of US soil carbon (Wilson <i>et al.</i> 2013)
Erosion control	Slope <sup>10</sup> , slope length, rainfall erosivity <sup>11</sup> , cover factor <sup>5</sup> , soil erodibility <sup>4</sup>	NRCS revised universal soil loss equation (Lim <i>et al.</i> 2005)

observed for all conservation areas to 1 for the highest relative capacity. The unstandardized statistical summaries for surface water run-off, ground water leaching, and soil loss and the equation used to standardize these values is provided in online supplementary material. We quantified the capacity for each conservation area, calculated a composite metric of all seven ESs by averaging the standardized values, and compared mean ES capacity among conservation area types (federal PPA, state PPA, and easement), owner/holder (for example the US Forest Service), and gap status (1, 2, and unknown) using Kruskal–Wallis analysis of variance and Wilcoxon signed rank tests, both non-parametric approaches since the ES capacity datasets did not meet parametric assumptions. Where a significant difference was detected, we compared means using the Wilcoxon test of pairs.

## RESULTS

### Geographic scope and size of conservation areas

Federal ( $n = 628$ ) and state protected areas ( $n = 493$ ) with a gap status of 1 or 2 were, on average, significantly larger in area than conservation easements ( $n = 230$ ;  $p < 0.0001$ ). We found federal PPAs with gap status of 1 or 2 covered 692 925 ha (1 712 256 acres), state protected PPAs covered 220 714 ha (545 397 acres), and easements covered 31 219 ha (77 144 acres) throughout the study region. There were ten-fold more conservation areas with a gap status of 2 than gap status of 1 at the state level, more than two-fold at the federal level, and seven-fold for easements (Table 2). The majority of federal PPAs were located in the Ridge and

Valley, Blue Ridge Mountains, and Coastal Plain ecoregions. State protected areas were widely distributed throughout VA, with larger parcels in Ridge and Valley, Blue Ridge Mountains, and Coastal Plain ecoregions, but were less common in NC. Conservation easements were much smaller, but evenly dispersed throughout VA (Fig. 1). There were fewer easements that met our study criteria in NC than VA, but they were substantially larger and generally located in the mountainous west and coastal plain. Conservation areas were largely absent near the large population centres of Winston-Salem, Greensboro, eastern Durham, Raleigh, Greenville, and Rocky Mount in NC and Richmond in VA.

### Ecosystem service capacities

#### *Surface water regulation*

State and federal PPAs offered significantly greater surface water regulation than conservation easements (Fig. 2). Surface water regulation did not differ significantly among state lands, but regulation was significantly greater on US National Park Service (USNPS) and US Forest Service (USFS) lands than on US Fish and Wildlife Service (USFWS) managed National Wildlife Refuges (NWRs), although mean capacity was very high for all. Federally-held easements offered the least surface water regulation; about six times more run-off was expected than from state or NGO-held easements. Surface water regulation was significantly greater on easements with gap status of 1, but no other significant difference was detected among the state or federal protected areas considered (Table 3). Geographically, federal PPAs provide the greatest

**Table 2** Mean area (ha)  $\pm$  standard error and the sample size ( $n$ ) for all conservation areas within the study area of North Carolina and Virginia, USA. Federal and state public protected areas were significantly larger than private conservation easements ( $p < 0.0001$ ). Conservation areas with the same superscript letter following the mean and standard error values are not significantly different. Significant differences among conservation area owners-holders are noted with a lower case letter and a capital letter for differences between gap status 1 and 2. Owners/holders of conservation areas are: Virginia Department of Conservation and Recreation (VADCR), Virginia Department of Game and Inland Fisheries (VDGIF), North Carolina Department of Parks and Recreation (NCDPR), US Fish and Wildlife Service National Wildlife Refuges (USFWS-NWR), US Fish and Wildlife Service (USFWS), US National Park Service (USNPS), and non-governmental organization (NGO). SNA = state natural area.

Conservation areas	$n$	Mean area (ha)
<i>State</i>	493	447 $\pm$ 63
VADCR	152	234 $\pm$ 48a
VDGIF	215	706 $\pm$ 130b
NCDPR (SNAs)	31	499 $\pm$ 215ab
NCDENR	65	197 $\pm$ 112
Gap status 1	43	185 $\pm$ 51
Gap status 2	450	472 $\pm$ 68
<i>Federal</i>	628	1102 $\pm$ 303
USFWS-NWR	430	539 $\pm$ 142a
USFS	189	1 423 $\pm$ 687b
USNPS	8	23 900 $\pm$ 14 207ab
Gap status 1	194	1 757 $\pm$ 883
Gap status 2	434	808 $\pm$ 190
<i>Easements</i>	230	135 $\pm$ 34
Federal	135	15 $\pm$ 3a
State	18	323 $\pm$ 206b
NGO	64	242 $\pm$ 82c
Gap status 1	25	455 $\pm$ 197A
Gap status 2	176	9 $\pm$ 631AB

opportunity to regulate run-off in the mountain west, arguably the most important area since these headwaters contribute to all freshwater systems downstream. There is a paucity of gap status 1 or 2 conservation areas in west-central NC (Fig. 1). Although surface water regulation is relatively high throughout the region, pockets of low capacity can be found in and near urban areas where greater impervious surface area contributes to excessive run-off (Fig. 3a).

#### Groundwater quality protection

In general, all conservation areas in the study offered low groundwater protection capacity (Fig. 2), meaning there was high potential for nitrate leaching below the root zone (nitrogen leaching index values  $> 10$ ; Czymbek *et al.* 2003). Groundwater protection capacity was significantly less within federal PPAs than other conservation areas ( $p < 0.0001$ ; Fig. 2); with USNPS lands having the lowest mean capacity. At the state-level, the VA Department of Conservation and Recreation (VADCR), NC Department of Environment and

Natural Resources (NCDENR), and some NC Department of Parks and Recreation (NCDPR) lands had the highest capacity for groundwater protection; however, state natural areas (SNAs) managed by NCDPR had the lowest capacity. There was no significant difference in groundwater quality protection capacity between gap statuses (Table 3). In addition to the lack of conservation areas in the aforementioned Catawba and Yadkin-PeeDee basins of west-central NC, state PPAs generally have greater groundwater protection capacity throughout the study region. In contrast, the geographic clumping of federal PPAs and easements reduces the extent of their service benefits. Within federal conservation areas there was a wide range in groundwater protection capacity, but hotspots of low capacity were detected in the western mountains of VA and NC, as well as along the Atlantic coast.

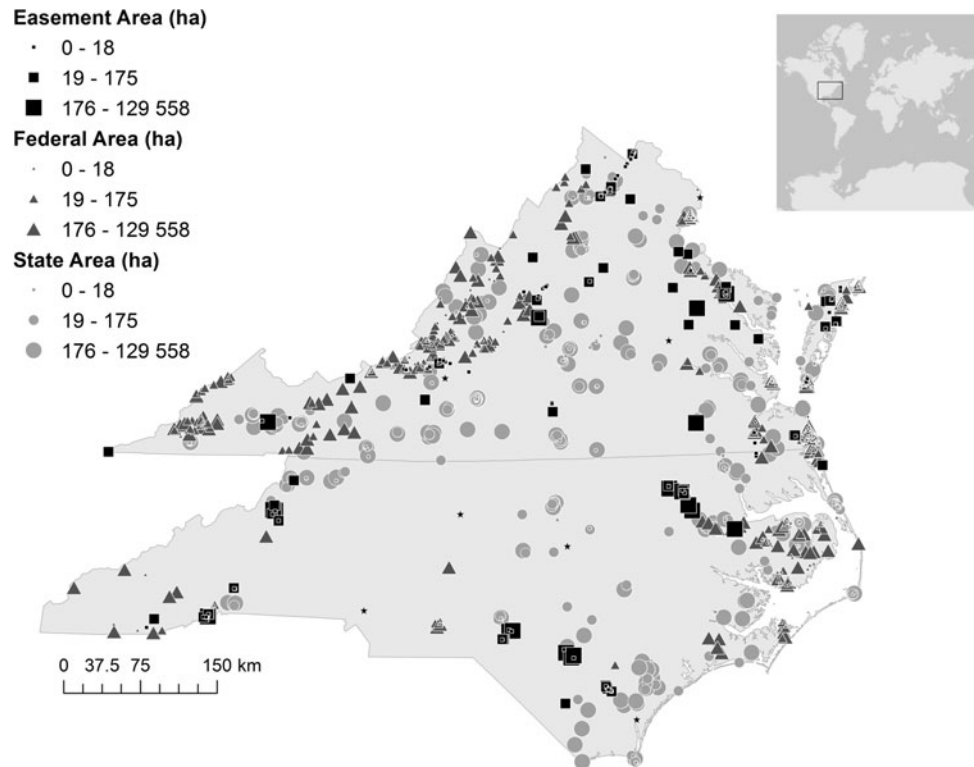
#### Riparian filtration

Among conservation areas with water features, federal, state PPAs and easements had significantly greater mean riparian filtration capacities than federal PPAs ( $p < 0.0001$ ). Riparian filtration capacity was significantly different among state-owned properties; however, there was no difference among federal lands or easements and there was no significant difference in riparian filtration capacity between gap statuses for any of the conservation area types. Mean filtration capacity for all conservation areas was around 66% (Table 3). This means that, on average, about 34% of the nutrients and sediment flowing into the riparian corridor is loaded into surface waters. The map of riparian filtration (Fig. 3b) suggests that there are many state PPAs and easements, particularly in the coastal plain of NC, that could greatly enhance their riparian filtration capacity. It should also be noted that only 51% of all conservation areas in the study contained or were immediately adjacent to surface water features. Overall, this suggests that existing conservation areas could increase their riparian capacity to enhance water quality protection and riparian filtration could be a priority for future conservation area development.

#### Erosion control

Several data sources were needed to estimate soil loss and some of these were not available for all conservation areas in the study. Missing data were largely attributed to areas not covered by the most current US Soil Survey Geographic Database (SSURGO) data, including some barrier islands and aquatic-dominated conservation areas. Only 2% of easements, 6% of federal PPAs, and 7% of state PPAs evaluated (97% of conservation areas in the study has sufficient data to assess soil loss) were expected to fully control erosion. Expected erosion control did not differ among conservation area types, nor did it differ among easement holders. Within the states of NC and VA, NCDPR SNAs and NCDENR lands have a significantly higher capacity to retain soil and prevent erosion. USFS lands have the highest capacity to prevent erosion and there was no difference between USFWS NWRs and USNPS lands. Erosion control was significantly greater in federal PPAs and

**Figure 1** Location and size of conservation easements, and federal and state protected areas (PPAs) that have a gap status of 1, 2, or unknown (easements only) in North Carolina and Virginia. Bubble size corresponds to the relative area of the conservation area measured in hectares.

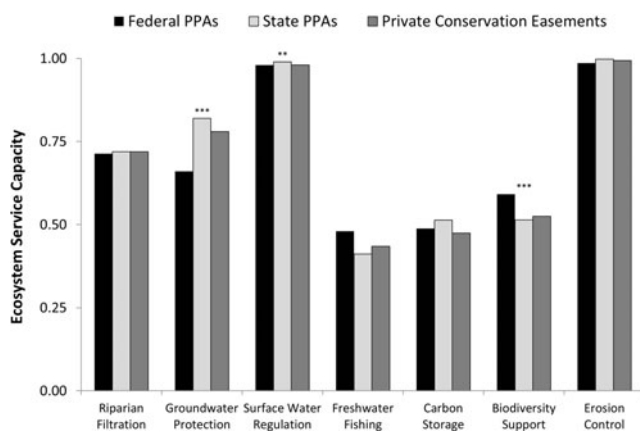


easements with a gap status of 1 than in status 2 lands, but no difference was detected at the state-level (Table 3). We did not detect a clear geographic pattern in erosion control capacity within or among conservation area types (Fig. 3c)

#### Carbon storage

The data sources used did not provide carbon storage data for all of the conservation areas in NC and VA. Most of the conservation areas for which data were not available for all three data sources were located on barrier islands along the Atlantic coast. More than 100 federal conservation areas lacked

carbon data, but only 19 state PPAs and one easement were excluded from the analysis due to insufficient data. Among the conservation areas for which data were available, state PPAs store significantly more carbon than federal PPAs or easements. Within state lands, the VA Department of Game and Inland Fisheries (VADGIF) and NCDENR lands stored the greatest carbon, and VA Department of Conservation and Recreation (VADCR) and NC Divisions of Parks and Recreation (NCDPR) lands in NC and VA stored the least. Surprisingly, USFWS NWRs stored significantly more carbon ( $329.1 \pm 7.7 \text{ tC ha}^{-1}$ ) than USFS lands attributed with greater soil carbon storage ( $158.7 \pm 5.3 \text{ tC ha}^{-1}$ ). Soil carbon comprised 91% of total carbon found on USFWS NWRs and 68% of that on USFS lands. Soil carbon storage was substantially higher in wetlands than forests throughout most of NC and VA (West *et al.* 2013). Also, NGO-held easements stored significantly more carbon per hectare than federal or state PPAs. Higher gap status did not positively affect carbon storage within any of the conservation area types (Table 3). Carbon storage capacity varied the most geographically among all the focal ESs (Fig. 3d). It was relatively low in the mountainous west compared to much greater storage in the coastal plain. This acute disparity is largely attributed to much greater soil carbon storage in coastal wetlands.



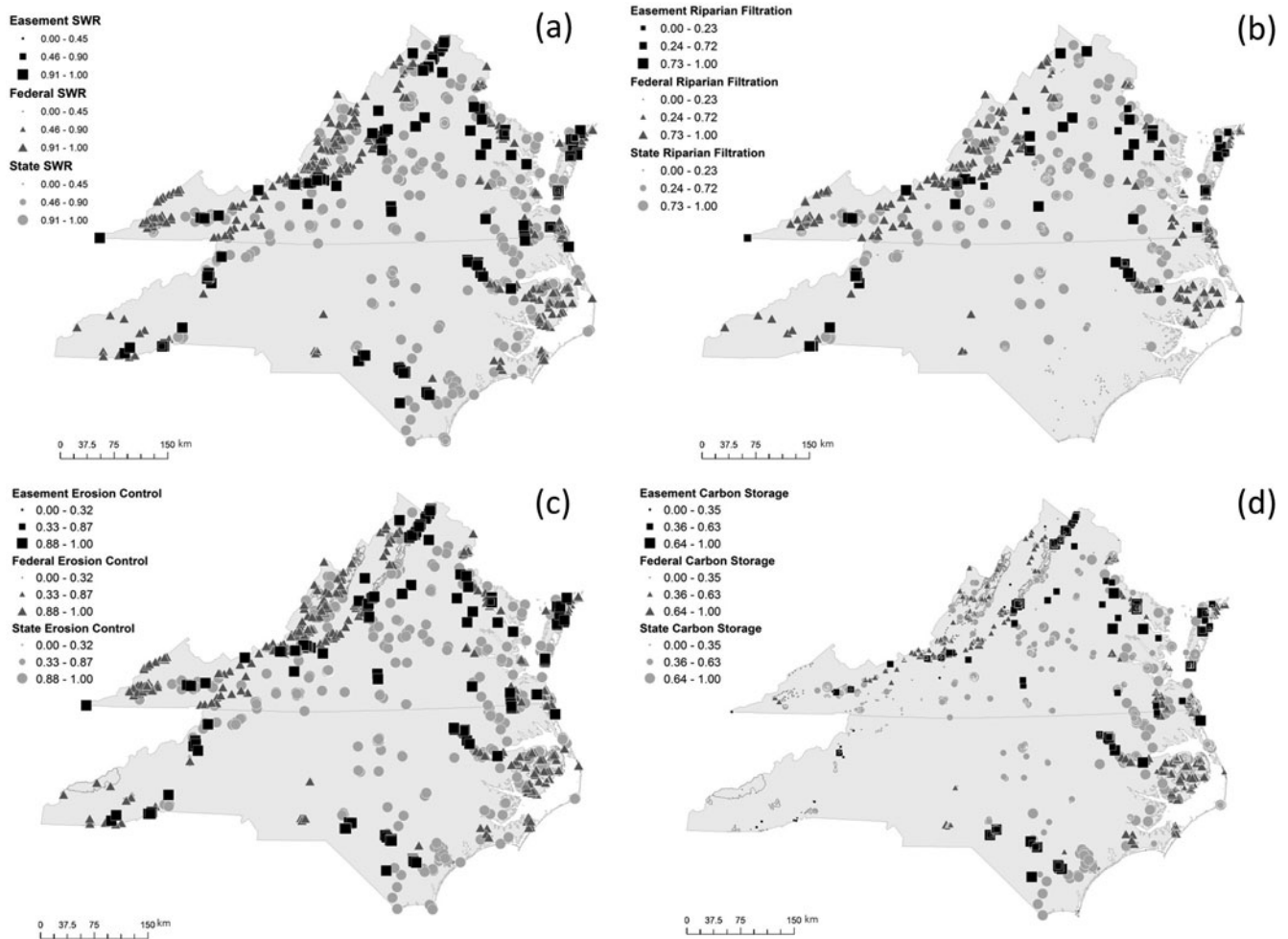
**Figure 2** Mean ( $\pm$  standard error) ecosystem service capacities among public (federal and state) and private conservation lands (easements). Statistical significance noted directly above each comparison: \*\*\* $p < 0.0001$ , \*\* $p < 0.01$ , and \* $p < 0.05$ .

#### Biodiversity support

Data from the NC and VA (terrestrial) gap analysis did not provide sufficient data for 28 federal PPAs, four state PPAs, and 11 easements. This is likely due to the location of conservation areas on barrier islands and conservation

**Table 3** Statistical summary of mean ( $\pm$  standard error) ecosystem service (ES) capacities of conservation areas within North Carolina and Virginia. Riparian filtration, carbon storage, and biodiversity support are reported as absolute values, freshwater recreational fishing, ground water protection, surface water regulation, erosion control are standardized to range from 0 for lowest relative capacity to 1 for highest relative capacity. Statistically significant differences among conservation area owners-holders are noted with a lower case letter and a capital letter for differences between gap status 1 and 2. Owners/holders of conservation areas are: Virginia Department of Conservation and Recreation (VADCR), Virginia Department of Game and Inland Fisheries (VDGIF), North Carolina Division of Parks and Recreation (NCDPR), US Fish and Wildlife Service National Wildlife Refuges (USFWS-NWR), US Fish and Wildlife Service (USFWS), US National Park Service (USNPS), and non-governmental organization (NGO). SNA = state natural area.

<i>Managers</i>	<i>Surface water regulation</i>	<i>Ground water protection</i>	<i>Riparian filtration*</i>	<i>Erosion control</i>	<i>Carbon storage</i>	<i>Biodiversity support</i>	<i>Recreational fishing</i>	<i>All ES</i>
State								
VADCR	0.97 $\pm$ 0.008	0.69 $\pm$ 0.01a	63.1 $\pm$ 1.37a	.982 $\pm$ 0.007a	289.8 $\pm$ 8.3a	222.0 $\pm$ 2.2a	0.23 $\pm$ 0.009a	0.72 $\pm$ 0.011a
VDGIF	0.99 $\pm$ 0.003	0.62 $\pm$ 0.01b	70.5 $\pm$ 1.19b	.991 $\pm$ 0.000a	355.1 $\pm$ 9.5b	193.2 $\pm$ 3.0b	0.17 $\pm$ 0.007b	0.66 $\pm$ 0.012b
NCDPR (SNAs)	0.94 $\pm$ 0.02	0.48 $\pm$ 0.04c	72.1 $\pm$ 1.54c	.999 $\pm$ 0.001b	289.3 $\pm$ 27.4ac	150.6 $\pm$ 8.3c	0.19 $\pm$ 0.02b	0.56 $\pm$ 0.032c
NCDENR	0.99 $\pm$ 0.004	0.72 $\pm$ 0.02a	60.5 $\pm$ 3.09c	.996 $\pm$ 0.292c	332.2 $\pm$ 23.2bc	112.5 $\pm$ 5.9d	0.02 $\pm$ 0.002c	0.50 $\pm$ 0.020c
Gap status 1	0.98 $\pm$ 0.01	0.68 $\pm$ 0.02	66.4 $\pm$ 2.48	0.982 $\pm$ 0.008	287.4 $\pm$ 16.2	222.1 $\pm$ 3.8	0.20 $\pm$ 0.01	0.72 $\pm$ 0.018
Gap status 2	0.98 $\pm$ 0.004	0.65 $\pm$ 0.009	66.4 $\pm$ 0.9	0.990 $\pm$ 0.003	331 $\pm$ 6.6	185.9 $\pm$ 2.5	0.17 $\pm$ 0.006	0.65 $\pm$ 0.008
Federal								
USFWS-NWR	0.98 $\pm$ 0.005a	0.67 $\pm$ 0.01a	64.9 $\pm$ 0.80	0.979 $\pm$ 0.006a	360.5 $\pm$ 8.2a	211.3 $\pm$ 3.0	0.18 $\pm$ 0.006a	0.61 $\pm$ 0.009a
USFS	0.98 $\pm$ 0.005b	0.62 $\pm$ 0.02b	68.0 $\pm$ 0.63	1.00 $\pm$ 0.000b	230.8 $\pm$ 5.5b	212.1 $\pm$ 1.5	0.27 $\pm$ 0.005b	0.66 $\pm$ 0.011b
NPS	0.98 $\pm$ 0.01b	0.58 $\pm$ 0.04b	70.7 $\pm$ 0.98	1.00 $\pm$ 0.000ab	254.5 $\pm$ 34.6b	184 $\pm$ 21.1	0.23 $\pm$ 0.01ab	0.65 $\pm$ 0.71a
Gap status 1	0.98 $\pm$ 0.005	0.66 $\pm$ 0.06	66.7 $\pm$ 0.93	1.00 $\pm$ 0.000	269.5 $\pm$ 8.5	201.1 $\pm$ 2.7	0.22 $\pm$ 0.008	0.64 $\pm$ 0.011
Gap status 2	0.98 $\pm$ 0.004	0.65 $\pm$ 0.01	65.6 $\pm$ 0.72	0.979 $\pm$ 0.005	334.7 $\pm$ 8.1	215.8 $\pm$ 2.8	0.21 $\pm$ 0.005	0.62 $\pm$ 0.009
Easement								
Federal	0.95 $\pm$ 0.01	0.81 $\pm$ 0.008a	63.4 $\pm$ 2.70	0.965 $\pm$ 0.010	296.8 $\pm$ 36.2a	209.1 $\pm$ 3.7a	0.17 $\pm$ 0.007a	0.59 $\pm$ 0.011a
State	0.99 $\pm$ 0.005	0.66 $\pm$ 0.03b	66.0 $\pm$ 4.51	0.994 $\pm$ 0.005	282.1 $\pm$ 27.4a	210.5 $\pm$ 7.9a	0.22 $\pm$ 0.1b	0.69 $\pm$ 0.028b
NGO	0.99 $\pm$ 0.003	0.59 $\pm$ 0.01b	68.5 $\pm$ 3.46	0.998 $\pm$ 0.001	379.1 $\pm$ 24.9b	156.3 $\pm$ 4.1b	0.22 $\pm$ 0.004b	0.59 $\pm$ 0.017a
Gap status 1	0.99 $\pm$ 0.006A	0.72 $\pm$ 0.03	64.5 $\pm$ 4.75	0.992 $\pm$ 0.005	315.9 $\pm$ 24.6	197.8 $\pm$ 10.0	0.18 $\pm$ 0.02	.70 $\pm$ 0.03A
Gap status 2	0.97 $\pm$ 0.01B	0.71 $\pm$ 0.1	67.0 $\pm$ 2.25	0.978 $\pm$ 0.008	302.5 $\pm$ 9.1	185.4 $\pm$ 3.3	0.19 $\pm$ 0.005	0.66 $\pm$ 0.02B



**Figure 3** Location and relative capacity for (a) surface water regulation (SWR), (b) riparian filtration, (c) erosion control, and (d) carbon storage among conservation easements, federal protected areas, and state protected areas.

areas that included mostly aquatic ecosystems. Total species richness was significantly greater in federal PPAs, but there was no difference between state PPAs and easements. The USFWS NWRs and USFS lands were home to significantly more species than USNPS lands. At the state level, VADCR and NCDPR lands contained more species than other types. Species richness on easements was similar to state lands, but within easements federal and state PPAs were home to more species. Species richness was only significantly higher in gap status 1 lands at the state level (Table 3). Overall, biodiversity support differed by taxa and was greatest in the mountainous west and coastal plain. There was substantial geographic variation in biodiversity support capacity across the study region. Greater species richness was supported in the mountains of VA, attributable to the size of federal conservation areas (Table 2, Fig. 1). State PPAs throughout the region maintained relatively high support for biodiversity.

#### Freshwater recreational fishing

Mean freshwater recreational fishing capacity among conservation areas was low, but significantly higher on federal

PPAs ( $p < 0.0001$ ). Among federal PPAs, USFS lands had the highest capacity. At the state-level, VADCR and some NCDPR properties were within 12-digit watersheds with the highest capacity, significantly higher than VADGIF lands and NCDPR SNAs ( $p < 0.001$ ). We did not detect a statistically significant difference in FRF capacity between gap statuses for any conservation areas considered, nor did we find a clear geographic pattern (Table 3).

#### All ecosystem service capacities

At the state level, VADCR PPAs (VA state parks and state natural area preserves) had significantly higher ES capacity than all other state-managed lands. North Carolina state parks were excluded from the analysis based on their gap status rating of 3. NCDENR coastal reserves and NCDPR special natural areas contained the lowest overall ES capacity. Although this may be partially expected because the ES we quantified were less common to coastal areas, it is surprising for the NCDPR SNAs, which are considered to represent some of NC's great diversity of resources and fragile ecological systems (NCDPR 2014). These same lands



protected significantly less species richness than other state systems.

Although total ES capacity was significantly greater on USFS lands compared to USFWS NWRs, the magnitude of these differences was not substantial and total ES capacity on all federal lands was less than we found associated with VA state parks and natural area preserves (Table 3). The total ES capacity was significantly greater within easements managed by the states than those by the federal government or NGOs. This corroborates the aforementioned pattern in which total ES capacity was significantly greater within state PPAs than federal PPAs or private easements. The geography of ESs within all conservation areas suggests that VA is protecting a greater extent of land and that these lands are of greater ES capacity than their NC counterparts. In stark contrast to the overlap of high ES capacity conservation areas in VA, conservation areas in coastal and western NC did not appear to maintain high ES capacity. Also, federal PPAs of high capacity were often adjacent to low capacity PPAs.

## DISCUSSION

### Do conservation easements protect ES as much as PPAs?

Overall, we found that ecosystem service capacity in easements was equal to or greater than capacity within state or federal protected areas for six out of the seven services assessed. On average, conservation easements housed greater focal ES capacity (Table 3); ES capacity was only significantly greater in both state and federal PPAs for surface water regulation. Therefore, we suggest that private land protection is enhancing the regional conservation of ESs and improving the environmental quality in areas where public PPAs cannot. Given that easements are on average significantly smaller than state or federal PPAs, the number of easements needed to conserve the same amount of services as a larger PPA may be higher. With the exception of biodiversity support, which has been shown to have a positive relationship with contiguous area (Rosenzweig 1995), the protection of most of the ESs we evaluated could be achieved through a conservation mosaic approach in which smaller parcels of protected land provide the same amount of an ES per unit area. Moreover, new easements with high erosion control capacity could be sought in and upstream of areas of high biodiversity and high freshwater recreational fishing capacity. Likewise, creating easements that focus on enhancing riparian filtration and erosion control downstream of high surface water run-off areas (such as urban and suburban areas) can help minimize the negative impacts of run-off on biodiversity. At the same time, easements that enhance vegetation to increase erosion control and riparian filtration will likely increase carbon storage potential. With these potential approaches in mind, we believe that ES conservation can be enhanced through private land conservation in NC and VA.

### Enhancing ES on conservation areas

This assessment evaluated differences in the existing capacity to provide ESs among conservation area types; by doing so we can identify groups of conservation areas that could potentially enhance ES protection if and where demand and management objectives aligned. Many of the ESs assessed in this analysis could be enhanced through increased management efforts where appropriate. For example, based on the results of this assessment, state protected properties by agencies like NCDENR, VADCR and NCDPR and federally-held easements could fund riparian restoration projects to enhance riparian filtration. The results of this assessment provide valuable information to land managers that may be used in conjunction with knowledge of management objectives, societal demand, cost effectiveness, as well as other sociopolitical trade-offs and synergies, in order to prioritize funding for ES enhancement. It is important to recognize that it may not be possible to enhance all ES; there may be internal trade-offs. For example, enhancing freshwater recreational fishing capacity with an increase in accessibility (for example boat ramps) may decrease a conservation area's ability to regulate surface water. Likewise, stocking non-native game fish may reduce native biodiversity support. Thus, we highly recommend that the results of this and future assessments be considered in conjunction with an evaluation of potential trade-offs, synergies, ES demand, and conservation budgets.

Conservation easements in NC and VA protect lands with a higher overall ES capacity than PPAs. Thus, increasing ES protection through private land conservation is an important step towards protecting ecological resilience and human well-being throughout the region. We suggest that ES can be enhanced by means of landowner incentives or regulatory mechanisms. Incentive programmes could be modelled after USFWS or NRCS landowner-targeted conservation programmes (such as the landowner incentive programme, environmental quality incentive programme, or wildlife habitat incentive programme) or regulations could be developed as part of a comprehensive watershed plan (see for example the Chesapeake Bay total maximum daily loads).

Among the services considered here, riparian filtration, erosion control, carbon storage, and surface water regulation capacity are the easiest and least expensive services to enhance because land cover and land use play such an important role (Table 1). To enhance these services on easement properties: (1) easement contracts could require the establishment of a minimum vegetated area within the riparian zone within a set period of time (such as two years), (2) riparian vegetation could be incentivized by scaling the tax benefits and breaks to maximize riparian filtration effectiveness (based on state best-management practice guidelines), (3) a cap on allowable impervious surfaces can be set to increase surface water regulation, (4) well-placed and functioning retention ponds and other surface water retention strategies can be included in easement contracts to offset the effects of impervious surface, and (5) upland and riparian reforestation can be

incentivized by allowing the stacking of tax benefits (Banerjee *et al.* 2013). For the most part, these actions could be evaluated using remote imagery and GIS through time to ensure land owners maintain these attributes. By adding a premium to lands that will provide greater public benefit, landscape-level conservation would be invested in. For services that are more difficult to enhance, like groundwater protection, recruiting new easements of low capacity may be even more important than conserving high capacity conservation areas. For example, by offering landowners of low groundwater protection lands incentives to put their land into easement may reduce the ecological pressures on the system (such as introduction of contaminants to a high leaching area) and help prevent contaminants from leaching into the groundwater.

Land can be managed within easements and PPAs to maximize synergies among services and minimize the cost of management. For example, riparian vegetation may enhance water quality regulation, but it also provides habitat to a wide array of aquatic and terrestrial biodiversity that provides the basis for wildlife-based recreation (Villamagna *et al.* 2013*b*). ES-based incentives could be structured such that all ESs are equally valued, or the landowner is rewarded for each additional service generated. The latter is an ES stacking approach that is used in payment for ecosystem service markets (PES) to increase compensation to the providers of ESs and to further incentivize ES conservation (Banerjee *et al.* 2013).

### Wise-ES management

Our assessment of ESs throughout NC and VA provided important information to identify conservation areas for which certain activities should be avoided. Although erosion control capacity did not vary between PPAs and easements, VADCR, VDGIF, and NCDPR lands are more likely to erode soil into their watersheds than SNAs or NCDENR lands. Likewise, USFWS NWRs are more likely to contribute sediment to local waterways than USFS or USNPS lands. Therefore, these conservation areas in particular might seek to reduce highly erosive activities. Likewise, conservation areas with low groundwater protection capacity could alter land use to avoid contaminants from leaching. This would be particularly relevant within easements that continue agricultural practices which may introduce excessive nitrogen. Such considerations should be addressed during the easement contract process, and the assessment of regulating services presented here can help inform this process. In general, areas with low regulating capacities that are difficult to enhance (such as groundwater protection) should have stricter land use requirements written into the contracts. This may include a change in the seasonal application of nitrogen or requiring additional vegetation to enhance the uptake of surface nitrogen, the latter of which may also enhance carbon storage, riparian filtration, and biodiversity support in some areas.

### The future of ES conservation

Large-scale PPAs provide a significant opportunity to protect and enhance ESs and biodiversity. However, the critical need to protect ESs expands far beyond the reaches of traditional parks, reserves, and refuges. Likewise, the current political climate and residential development in the south-east USA (Stein *et al.* 2007) make government-sponsored PPAs increasingly difficult to expand. It is therefore both logical and necessary to consider enhancing the protection of ESs and biodiversity across private landscapes in order to maintain ecological resilience in the face of amplified threats due to growing and shifting human populations and climate change. In order to do so, conservation efforts should: (1) continue to protect natural and uninhabited areas that provide ecosystem and biological diversity, (2) expand conservation efforts into private lands closer to current human population centres, (3) limit future development to resilient areas that can sustain new population growth and incorporate ES-wise landscape planning, and (4) increase protection status from gap status 3 or 4 to 1 or 2 for existing private or public lands in high utility areas. By taking human population density and land use into consideration, we can identify areas where an increase in protection status might provide greater collateral ES benefits. For example, the lack of gap status 1 or 2 conservation areas in west-central NC (Fig. 1) limits the current opportunity to prevent or mitigate the storm water and water quality issues that are increasing in this region (the Catawba River basin) due to rapid population growth and land-use changes (Kramer & Eisen-Hecht 2002; Gage *et al.* 2004). We expect an expansion of ES benefits from conservation areas to more beneficiaries in currently underrepresented areas by shifting conservation efforts to address these aforementioned priorities.

Private conservation easements offer extraordinary potential for expanding biodiversity and ES protection because they include the most productive lands (Scott *et al.* 2001), currently protect areas of relatively high capacity, and could protect diverse habitats from land conversion. When considering public land acquisition as the alternative, conservation easements cost at least 40% less per unit area than the outright purchase of the land (fee simple; Fishburn *et al.* 2009) and current conservation easements have a similar capacity to provide ES to federal and state PPAs. Perhaps more importantly, due to the flexibility of their geographic scope, their relative size, and the economic and cultural appeal to private landowners (WRI [World Resources Institute] 2007), easements can be established in areas where PPAs are limited and where larger human populations are in need of high regulating service capacities. A mosaic of broadly dispersed, albeit relatively small, conservation easements held by a variety of managers could provide a more resilient and synergistic network of biodiversity and ES conservation that is not currently achievable by public conservation efforts alone in the developed and densely populated mid-Atlantic and south-east USA, as well as other developing areas worldwide.

## CONCLUSIONS

This assessment has highlighted several globally important messages. First, private conservation strategies may be the most efficient and cost-effective approach to conserving ESs for the most people. Given their location and relative flexibility, private conservation easements may enhance ES conservation throughout working landscapes thereby providing benefits to a broader array of beneficiaries in areas outside of the reach of federal and state PPAs. Second, there are clear differences in current ES capacities among different public and private entities. Some of the differences may be attributed to landscape-scale management objectives. For example, biodiversity-centric conservation areas such as the USFWS wildlife refuges were likely established in areas of greater species richness than the USFS national forests, which focused more on timber production, aesthetics, recreation and, more recently, water quality. However, despite inherent differences in conservation goals, a baseline assessment of ES capacities across all levels of conservation provides a starting point for evaluating potential trade-offs associated with land-use and policy changes. This assessment also integrates ESs into the on-going conversation of global conservation priorities, including where and how biodiversity and ES conservation can be synergized.

Public and private conservation areas can and should work together to protect a wider array of ESs. Efforts to enhance conservation should explicitly consider: (1) current capacity of ESs, (2) trade-offs or synergies among ESs, (3) the geographic scope and magnitude of benefits, and (4) potential beneficiaries of current and new conservation areas (A. Villamagna *et al.*, unpublished data 2014). Private land conservation offers an exciting opportunity to protect ES in areas outside of PPAs, however such strategies are hinged on the ability to provide sufficient compensation to a landowner. Compensation for ES generation on easement lands can be structured in a variety of ways, including a bundled service or stacked service approach (Banerjee *et al.* 2013) in order to incentivize ES conservation on private lands.

The spatial juxtaposition of conservation areas with high and low total ES capacity suggest fine scale heterogeneity in natural or anthropogenic features (such as land use) across VA and NC. This level of spatial variability supports our suggestion that ES conservation move toward enhancing a broad, landscape-scale network of private conservation areas to achieve ES protection in more areas across the states. Ultimately, a comprehensive conservation network that protects and enhances the flow of critical ESs, including biodiversity, should be achievable by combining existing PPAs with current and future conservation easements on private lands.

## ACKNOWLEDGEMENTS

We thank B. Rollison for his contributions to the GIS analysis and to B. Mogollón and P. Angermeier for their

valuable feedback on earlier drafts. Also thanks to S. Prisley for assistance locating data and to the departments of Fish and Wildlife Conservation and Crop and Soil Environmental Science at Virginia Tech for supporting this research.

## Supplementary material

To view supplementary material for this article, please visit <http://dx.doi.org/10.1017/S0376892914000393>.

## References

- Banerjee, S., Secchi, S., Fargione, J., Polasky, S. & Kraft, S. (2013) How to sell ecosystem services: a guide for designing new markets. *Frontiers in Ecology and the Environment* 11(6): 297–304.
- Bonnells, M. (2012) Private nature reserves: an innovative wetland protection mechanism to fill in the gaps left by the Swancc and Rapanos rulings. *Environs* 36: 1–34.
- Butchart, S.H., Walpole, M., Collen, B., van Strien, A., Scharlemann, J.P., Almond, R.E., Baillie, J.E., Bomhard, B., Brown, C., Bruno, J., *et al.* (2010) Global biodiversity: indicators of recent declines. *Science* 328(5982): 1164–1168.
- Chape, S., Harrison, J., Spalding, M. & Lysenko, I. (2005) Measuring the extent and effectiveness of protected areas as an indicator for meeting global biodiversity targets. *Philosophical Transactions of the Royal Society B: Biological Sciences* 360(1454): 443–455.
- Czymmek, K.J., Ketterings, Q.M., van Es, H.M. & DeGloria, S.D. (2003) The New York nitrate leaching index. CSS Extension Publication E03–2, Cornell University, New York, NY, USA: 34 pp. [www document]. URL <http://nmsp.cals.cornell.edu/publications/extension/nleachingindex.pdf>
- Dissmeyer, G.E. (2000) Drinking water from forests and grasslands: a synthesis of the scientific literature. General Technical Report-Southern Research Station, USDA Forest Service SRS-39, Asheville, NC, USA [www document]. URL [http://www.srs.fs.usda.gov/pubs/gtr/gtr\\_srs039/gtr\\_srs039](http://www.srs.fs.usda.gov/pubs/gtr/gtr_srs039/gtr_srs039)
- Figgis, P. (2004) Conservation on private lands: the Australian experience. Report. IUCN, Gland, Switzerland and Cambridge, UK.
- Fishburn, I.S., Kareiva, P., Gaston, K.J. & Armsworth, P.R. (2009) The growth of easements as a conservation tool. *PLoS One* 4(3): e4996.
- Fry, J., Xian, G., Jin, S., Dewitz, J., Homer, C., Yang, L., Barnes, C., Herold, N. & Wickham, J. (2011) Completion of the 2006 National Land Cover Database for the Conterminous United States. *Photogrammetric Engineering and Remote Sensing* 77(9): 858–864.
- Fuller, R.A., McDonald-Madden, E., Wilson, K.A., Carwardine, J., Grantham, H.S., Watson, J. E., Klein, C.J., Green, D.C. & Possingham, H.P. (2010). Replacing underperforming protected areas achieves better conservation outcomes. *Nature* 466(7304): 365–367.
- Gage, M.S., Spivak, A. & Paradise, C.J. (2004). Effects of land use and disturbance on benthic insects in headwater streams draining small watersheds north of Charlotte, NC. *Southeastern Naturalist* 3(2): 345–358.
- Gesch, D.B. (2007) The National Elevation Dataset. In: *Digital Elevation Model Technologies and Applications: The DEM Users Manual, 2nd Edition*. ed. D. Maune, pp. 99–118. Bethesda,

- Maryland, USA: American Society for Photogrammetry and Remote Sensing.
- Gordon, A., Langford, W.T., White, M.D., Todd, J.A. & Bastin, L. (2011) Modelling trade offs between public and private conservation policies. *Biological Conservation* 144(1): 558–566.
- Griffith, J., Stehman, S. & Loveland, T. (2007) Landscape trends in Mid-Atlantic and Southeastern United States ecoregions. *Environmental Management* 32(5): 572–588.
- Ingraham, M.W. & Foster, S.G. (2008) The value of ecosystem services provided by the US National Wildlife Refuge System in the contiguous US. *Ecological Economics* 67(4): 608–618.
- Jackson, S.F. & Gaston, K.J. (2008) Incorporating private lands in conservation planning: protected areas in Britain. *Ecological Applications* 18: 1050–1060.
- Kramer, R.A. & Eisen-Hecht, J.I. (2002) Estimating the economic value of water quality protection in the Catawba River basin. *Water Resources Research* 38(9): 21–1.
- Langholz, J. & Lassoie, J. (2001) Combining conservation and development on private lands: lessons from Costa Rica. *Environment, Development and Sustainability* 3(4): 309–322.
- Lim, K.J., Sagong, M., Engel, B.A., Tang, Z., Choi, J. & Kim, K. (2005) GIS-based sediment assessment tool. *Catena* 64: 61–80.
- Mascia, M.B. & Pailler, S. (2011) Protected area downgrading, downsizing, and degazettement (PADDD) and its conservation implications. *Conservation Letters* 4(1): 9–20.
- Mayer, P.M., Reynolds, S.K., McCutchen, M.D. & Canfield, T.J. (2007) Meta-analysis of nitrogen removal in riparian buffers. *Journal of Environmental Quality* 36(4): 1172–1180.
- Merenlender, A.M., Huntsinger, L., Guthey, G. & Fairfax, S.K. (2004) Land trusts and conservation easements: Who is conserving what for whom? *Conservation Biology* 18(1): 65–76.
- NC DPR (2014) The park system: overview [www document]. URL [http://www.ncparks.gov/About/system\\_main.php](http://www.ncparks.gov/About/system_main.php)
- NCED (2013) National conservation easement database version 2 [www document]. URL <http://nced.conservationregistry.org/>
- NRCS (1972) *National Engineering Handbook*. Hydrology, Section 4. Chapters 4–10. Washington, DC, USA: USDA.
- NRCS (2003) RUSLE2 (Revised Universal Soil Loss Equation: Version 2) [www document]. URL [http://fargo.nserl.purdue.edu/rusle2\\_dataweb/RUSLE2\\_Index.htm](http://fargo.nserl.purdue.edu/rusle2_dataweb/RUSLE2_Index.htm)
- PRISM Climate Group (2010) PRISM climate data. PRISM Climate Group, Oregon State University, OR, USA [www document]. URL <http://prism.oregonstate.edu>
- Reyers, B., Polasky, S., Tallis, H., Mooney, H.A. & Larigauderie, A. (2012) Finding common ground for biodiversity and ecosystem services. *BioScience* 62(5): 503–507.
- Rosenzweig, M.L. (1995) *Species Diversity in Space and Time*. Cambridge, UK: Cambridge University Press.
- Scott, J.M., Davis, F.W., McGhie, R.G., Wright, R.G., Groves, Turner, C. & Estes, J. (2001) Nature reserves: do they capture the full range of America's biological diversity? *Ecological Applications* 11(4): 999–1007.
- Soil Survey Staff (2010) Web Soil Survey. Soil Survey Staff, Natural Resources Conservation Service, United States Department of Agriculture, USA [www document]. URL <http://websoilsurvey.nrcs.usda.gov/>
- Stein, S.M., Alig, R.J., White, E.M., Comas, S.J., Carr, M., Eley, M., Elverum, K., O'Donnell, M., Theobald, D.M., Cordell, K., Haber, J. & Beauvais, T.W. (2007) National forests on the edge: development pressures on America's national forests and grasslands. General Technical Report. PNWGTR-728. US Department of Agriculture, Forest Service, Pacific Northwest Research Station Portland, OR, USA: 26 pp.
- Turner, W.R., Wilcove, D.S. & Swain, H.M. (2006) Assessing the effectiveness of reserve acquisition programs in protecting rare and threatened species. *Conservation Biology* 20(6): 1657–1669.
- USDA Agriculture Research Service (2007) Revised universal soil loss equation 1.06. Bulletins: rainfall erosivity factor [www document]. URL <http://www.ars.usda.gov/research/docs.htm?docid=5990>
- USDA, NRCS, USGS & EPA (2010) Watershed boundary dataset for North Carolina and Virginia [www document]. URL <http://nhd.usgs.gov/data.html>
- USGS (2012) Gap analysis program: protected areas database of the United States (PADUS), version 1.3 Combined feature class. [www document]. URL <http://gapanalysis.usgs.gov/padus/data/metadata/>
- USGS (2013) Gap analysis program: species data [www document]. URL <http://gapanalysis.usgs.gov/species/>
- Villamagna, A.M. & Angermeier, P.L. (2015) A methodology for measuring and mapping ecosystem services provided by watersheds. In: *Ecosystem Services and River Basin Ecohydrology*, ed. L. Chicara, F. Mueller & N. Fohrer. London, UK: Springer.
- Villamagna, A.M., Angermeier, P.L. & Bennett, E.M. (2013a) Capacity, pressure, demand, and flow: a conceptual framework for analyzing ecosystem service provision and delivery. *Ecological Complexity* 15: 114–121.
- Villamagna, A.M., Angermeier, P.L. & Niazi, N. (2013b) Evaluating opportunities to enhance ecosystem services in public use areas. *Ecosystem Services* 7: 167–176.
- Villamagna, A., Mogollon, B. & Angermeier, P. (2014) A multi-indicator framework for mapping cultural services: the case of freshwater recreational fishing. *Ecological Indicators* 45: 255–265.
- Von Hase, A., Rouget, M. & Cowling, R.M. (2010) Evaluating private land conservation in the Cape Lowlands, South Africa. *Conservation Biology* 24(5): 1182–1189.
- Wilson, B.T., Woodall, C.W. & Griffith, D.M. (2013) Forest carbon stocks of the contiguous United States (2000–2009). USDA Forest Service, Northern Research Station, Newtown Square, PA, USA [www document]. URL <http://dx.doi.org/10.2737/RDS-2013-0004>
- West, L., Wills, S. & Loecke, T. (2013) Rapid assessment of US soil carbon for climate change and conservation planning: summary of soil carbon stocks for the conterminous United States. United States Department of Agriculture, Natural Resources Conservation Service, National Soil Survey Center, USA [www document]. URL [http://www.nrcs.usda.gov/Internet/FSE\\_DOCUMENTS/nrcs142p2\\_050979.pdf](http://www.nrcs.usda.gov/Internet/FSE_DOCUMENTS/nrcs142p2_050979.pdf)
- WRI (2007) Gaining ground: increasing conservation easements in the US South. World Resources Institute Issue Brief. Southern Forests for the Future Incentive Series [www document]. URL <http://wri.org/publication/Gaining-Ground>