

Regional comparisons of *Codium* (Chlorophyta) assemblages in the northern versus southern English Channel

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*The cryptic invasion of the Asian macroalga *Codium fragile* (Suringar) Hariot ssp. *fragile* on north-eastern Atlantic shores has been long considered a classical example of a successful invader that has competitively displaced native congeners. Yet, the lack of quantitative information about morphologically similar native congeners, namely *Codium tomentosum* Stackhouse and *C. vermilara* (Olivi) Delle Chiaje, has hindered interspecific comparisons and ecological predictions. From September 2002 to 2005, we made extensive intertidal surveys on 12 northern and 26 southern rocky shores of the English Channel, specifically documenting the abundance, distribution and identity of *Codium* assemblages. On the north side of the English Channel, the native *C. tomentosum* and alien *C. fragile* were both sparsely distributed in intertidal pools in Devon. In contrast, the natives were absent from and the alien was locally abundant in shallow lagoons and rocky reefs around Bembridge and Whitecliff on the Isle of Wight. Finally, in the Channel Islands off the Atlantic coast of France, *Codium* spp. were abundant in pools and on low-shore emergent substrata with native species predominating. Patterns of distribution varied substantially among Guernsey, Jersey and Alderney, despite their close proximity. The regional variation in the algal distributions merits further investigation to determine whether the pattern is produced by: (1) anthropogenic activities and effects; (2) ecological interactions; (3) oceanographic factors; or (4) some combination of these.*

Keywords: *Codium*, non-indigenous species, native species, north-eastern Atlantic, English Channel

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INTRODUCTION

Alien species of marine invertebrates, algae and fish are increasingly becoming important components and interactors in marine communities worldwide. In many cases, the non-indigenous species (NIS) have been noted due to increases in habitat-formers and ecosystem engineers such as invasive mussels, seagrass (Posey, 1988), or large seaweeds (Trowbridge, 2006). In other cases, NIS are detected by functional changes in species interactions. Quite often, the establishment and spread of NIS is visually obvious such as with the brown alga *Sargassum muticum* (Yendo) Fensholt. However, if the NIS are morphologically similar to resident species, the incursion can be cryptic, requiring detailed morphological and/or molecular analyses to elucidate the invasion dynamics. An important example of such cryptic invaders is the Asian green macroalga *Codium fragile*, originating from Japan, which has appeared on temperate shores throughout the world (Trowbridge, 1998; Provan *et al.*, 2005, 2007). In most invaded regions (except north-western Atlantic shores), this seaweed entered communities with native congeners and even non-weedy conspecifics. Untangling the invasion dynamics and documenting contemporary patterns of species occurrence is crucial to

making realistic risk assessments of invasives and establishing a quantitative baseline from which future changes can be assessed.

Codium fragile ssp. *tomentosoides* was first collected on mainland European shores (Holland) ~1900 and described in 1955 (Silva, 1955). Until recently, the first recognized record for the British Isles was 1939 in Devon, south-west England (River Yealm estuary) (Silva, 1955). Exciting new molecular work, however, has demonstrated that the alga was in Ireland by 1845, Scotland by 1891 and England by 1894 (Provan *et al.*, 2007). Furthermore, based on the botanical code, the correct name of this invasive alga is *Codium fragile* ssp. *fragile* (Maggs & Kelley, 2007; Provan *et al.*, 2007). For simplicity, we refer herein to this introduced alga as *C. fragile*. The conspecific *C. fragile* ssp. *atlanticum* (A.D. Cotton) P.C. Silva does not occur in the English Channel (see Trowbridge *et al.*, 2004 and Provan *et al.*, 2007, regarding incorrect records) so our use of the binomial *C. fragile* herein is restricted to the invasive *C. fragile* ssp. *fragile*.

During the 20th Century, *Codium fragile* proliferated on north-western and north-eastern Atlantic shores (and elsewhere). In the former region, the invasive alga now dominates many shallow areas, competing with kelp in the Gulf of Maine (Levin *et al.*, 2002; Scheibling & Gagnon, 2006; Schmidt & Scheibling, 2006, 2007) and seagrass in shallow lagoons (Garbary *et al.*, 1997, 2004). It has also invaded rocky intertidal pools (Bégin & Scheibling, 2003;

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Schmidt & Scheibling, 2005). The interspecific interactions have been well demonstrated (e.g. Levin *et al.*, 2002; Scheibling & Gagnon, 2006; Schmidt & Scheibling, 2005, 2007).

In contrast, work on north-eastern Atlantic shores has been more observational as field experiments with *Codium fragile* are illegal in the UK (Countryside Act of 1981). In the 1970s and 1980s, *C. fragile* formed extensive beds at Bembridge on the Isle of Wight (Benson *et al.*, 1983; Farnham, personal observation) and in Lough Hyne Marine Reserve, County Cork, Ireland (Norton, 1991). Some phycologists expressed concern that the introduced *C. fragile* would outcompete and displace native congeners (e.g. Farnham, 1980) (e.g. Figure 1A). Because there was no direct evidence of interspecific competition or a limiting resource, Trowbridge (1998, 2001) suggested the dynamics may involve temporal species' replacement (a decline of the native and unrelated increase of the introduced species; Figure 1B). A third hypothesis is that the cryptic invasion of *C. fragile* led to historical over-estimation of the native population abundance and geographical range with a subsequent readjusted perception after the description of *C. fragile* ssp. *tomentosoides* by Silva (1955) (Figure 1C). These three models are based on scientific observation and hypothesis formulation.

The species identification problem has been further exacerbated with *Codium tomentosum* and *C. vermilara* being considered warm-water species whose northward range expansion would be expected during climate warming (e.g. Hiscock *et al.*, 2004; Mieszkowska *et al.*, 2005) and *C. fragile* being considered constrained in the UK by cool summer temperatures (Elliott, 2006) when the alga already occurs north of Trondheim, Norway, and in the Canadian

Maritimes. Climatic change predictions could lead to a fourth hypothetical pattern: southward range contraction of the native *Codium* species in cold winters (e.g. 1962/1963) and northward proliferation in warm periods (Figure 1D).

Without a detailed understanding of contemporary distributions of the native and introduced congeners, predictions about future ecological and environmental changes cannot be realistically evaluated. Our primary objective was to test the displacement hypothesis by comparing the frequency of the high-profile native and introduced *Codium* species at three locations within the English Channel where historical records and some historical specimens exist.

MATERIALS AND METHODS

Study regions

We surveyed three major regions of north-eastern Atlantic shores: the mainland of southern England, the Isle of Wight (IoW) and the Channel Islands. Of the six counties bordering the north side of the English Channel, we concentrated most of our surveys in Devon (Figure 2A, B; Table 1) due to the predominance of historical *Codium* records there; however, we did visit a site in Sussex (Pagham Harbour) and another in Dorset (Chesil Fleet) where the second author had previously collected *Codium* (Farnham, 1975). The IoW, directly south of Portsmouth, is separated from the Hampshire mainland by the Solent, a deep strait. We surveyed four sites on the east coast of the IoW: Bembridge/Lifeboat Station North, Whitecliff Ledge, Horse Ledge and Yellow Ledge (Figure 2C). These sites were selected based primarily on the

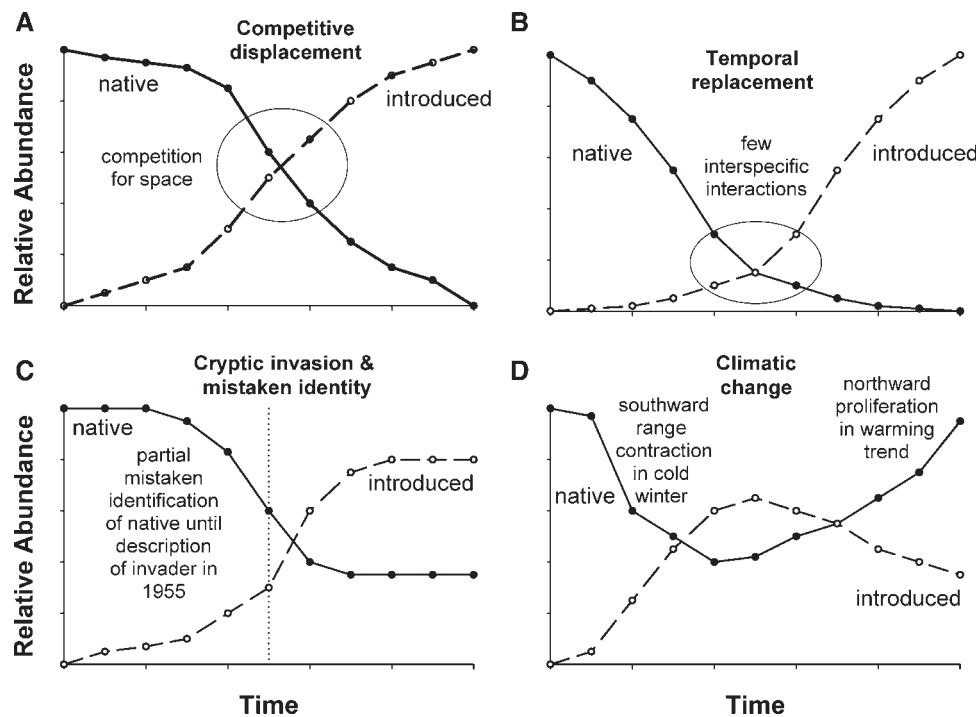


Fig. 1. Four schematic models of temporal change of native and introduced *Codium* spp. in the British Isles. Distinguishing between (A) competitive displacement and (B) temporal replacement depends on quantifying limiting resources (e.g. space) and interspecific interactions (or lack thereof). Phycological re-examination of historical specimens supports (C) but accurate population surveys are lacking. (D) Temporal changes of introduced and native species with known cold winters, warm summers and general climate changes have not been extensively evaluated.

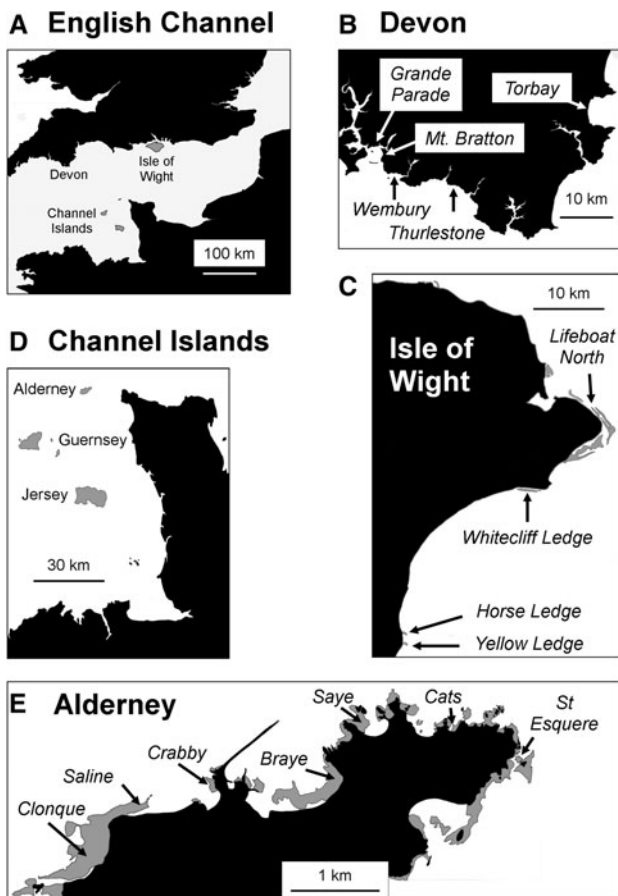


Fig. 2. Map of the survey regions and sites in the English Channel (north-eastern Atlantic shores). Scale bars are shown in each figure panel.

phycological literature (Foslie, 1893; Morey, 1909; Delf & Grubb, 1923; Norkett, 1947; Jones, 2000).

The Channel Islands in the Bay of St Malo off the Atlantic coast of France have received considerable pycological scrutiny (e.g. Lyle, 1920, 1923, 1937; Marquand, 1901, 1902, 1908; Van Heurck, 1908; Dixon, 1961; Feldmann, 1961; Culley *et al.*, 1983). We surveyed the three larger islands (Jersey, Guernsey and Alderney; Figure 2D) for logistical reasons of accessibility. Some results from Jersey and Guernsey have already been published (Trowbridge & Farnham, 2004; Trowbridge *et al.*, 2004). We concentrate herein on Alderney results, previously unpublished Guernsey and Jersey data and regional comparisons. We sampled ~1600 pools from seven Alderney sites from Clonque on the north-west shore to St Esquere on the north-east tip of the island (Figure 2E). Sites were selected based on accessibility and availability of rocky substrata.

Surveys

From September 2002–2005, we surveyed intertidal rock pools plus emergent substrata for the invasive *Codium fragile* and native congeners at 38 sites in the target regions (Table 1). Pools ranged in size from ~15 cm to 2–5 m in diameter with most being <1 m (notable exceptions were the vast intertidal lagoons at Bembridge, IoW). We quantified the percentage of rock pools occupied by *Codium* spp. at each site and, where possible, within each of four tidal levels,

defined by the following zone-forming furoids: the rockweeds *Pelvetia canaliculata* (Linnaeus) Decaisne & Thuret, *Fucus spiralis* Linnaeus, *F. vesiculosus* Linnaeus and *F. serratus* Linnaeus. At the upper end of the shore, *Codium* spp. were never on emergent substrata; on the lower half of the shore, *Codium* thalli were occasionally on emergent surfaces; we present results of pool-dwelling thalli (% occurrence) and densities of emergent thalli (based on 0.25 m² quadrats). Our sampling methodology was, for the most part, non-destructive. Branch tips of *Codium* thalli (2 cm from each thallus) were harvested to identify species, based on the morphology of the utricles, using an Olympus Mic-D digital compound microscope. Digital voucher specimens were captured for large numbers of thalli. Thallus length was measured to the nearest millimetre; size–frequency distributions of thallus length were compared for different regions, using pairwise Kolmogorov–Smirnov tests.

RESULTS

Sites in southern England (mainland)

At six sites in Devon (964 pools surveyed in April–May 2004), 1.2% of pools overall contained *Codium* spp. (Figure 3A), and <4% contained *Codium* spp. at any site. *Codium* spp. were sparsely distributed but did occur in pools at all tidal levels. We found 16 thalli of *C. fragile* and 23 of *C. tomentosum* in six days of low-tide surveys. The algal species averaged 3.2 and 3.3 thalli per occupied pool, respectively (Student's *t*-test, $t = 0.068$, $P = 0.947$). We found no *Codium* at Pagham Harbour in Sussex, despite the available habitats being characteristic of *C. fragile* (small rocks, large shells and other hard substrate) and the species having been collected there previously. Within the Chesil Fleet lagoon in Dorset, we found abundant native *C. vermilara* and introduced *C. fragile* on low-shore rocks below the *Ascophyllum* zone. Of the 31 thalli collected on 14 September 2002, 45% were *C. fragile* and 55% were *C. vermilara*.

On Devon intertidal shores, a wide range of sizes was present for *Codium tomentosum* and *C. fragile*. However, what was striking was the paucity of juvenile thalli of the latter (Figure 4A): there were no thalli or fronds <8 cm long. The size-structure of *C. fragile* ($N = 16$) on Devon shores was significantly different than that on Jersey shores (Giffard, Grève de Lecq, Castle and Sauchet) sampled during the same 2-week period (Kolmogorov–Smirnov, $d_{\max} = 0.458$, $P = 0.002$) but not that on Lihou and La Jaonneuse on Guernsey shores ($d_{\max} = 0.283$, $P = 0.241$) (Figure 4; Table 2). Jersey pools had large numbers of incipient fronds (0–2 cm) as well as large ones (to 35 cm). When we repeated the analysis across all collections from 2002 to 2005, we found significant differences in populations of *C. fragile* in the different regions (Table 2). The paucity of juveniles in Devon relative to other areas was the main distinguishing factor.

Sites on the Isle of Wight

Only one species of *Codium* was recorded: *C. fragile* was widely distributed on the east side of the IoW. Thalli were large, abundant and heavily covered with epiphytes at Bembridge in the shallow lagoons and at Whitecliff on low

Table 1. Spatial and temporal details of *Codium* surveys in the English Channel, north-eastern Atlantic Ocean.

| Region | County or island | Sampling dates | Site locations | Latitude and longitude | Pools surveyed | |
|------------------|------------------|-----------------------------|----------------------------------|----------------------------------|----------------------|-----|
| Northern Channel | Devon | April 2004 | Grande Parade, Plymouth | 50°24.1' N 4°09.0' W | 50 | |
| | | April 2004 | Mt Batten, Jennycliff Bay | 50°21.5' N 4°07.8' W | 125 | |
| | | April 2004 | Wembury | 50°19.0' N 4°05.6' W | 284 | |
| | | April 2004 | Thurlestone, Bigbury Bay | 50°15.8' N 3°51.7' W | 330 | |
| | | April 2004 | Torquay, Torbay | 50°27.6' N 3°32.4' W | 25 | |
| | Dorset | Isle of Wight | April 2004 | Brixham, Bigbury Bay | 50°24.1' N 3°30.0' W | 150 |
| | | | May 2003 | Chesil Fleet | 50°35.6' N 2°29.6' W | na |
| | | | May 2003, May and September 2005 | Bembridge/Lifeboat Station North | 50°41.6' N 1°03.9' W | na |
| | Sussex | | May 2005 | Whitecliff Ledge | 50°39.9' N 1°06.5' W | na |
| | | | May and September 2005 | Horse Ledge | 50°37.8' N 1°10.4' W | 200 |
| | | | May 2005 | Yellow Ledge | 50°36.7' N 1°10.3' W | 100 |
| | | | September 2002 | Pagham Harbour | 50°35.6' N 2°29.6' W | na |
| | | | Southern Channel | | | |
| Southern Channel | Guernsey | May 2004 | La Jaonneuse | 49°30.4' N 2°32.0' W | 100 | |
| | | June 2005 | Site 3 near Albecq | 49°29.8' N 2°34.2' W | 240 | |
| | | September 2002 | Bordeaux | 49°29.4' N 2°30.0' W | nd | |
| | | September 2002 | Port Soif | 49°29.3' N 2°34.9' W | nd | |
| | | September 2002 | Grande Rocques | 49°29.2' N 2°35.4' W | nd | |
| | | June 2005 | Albecq Bay | 49°28.5' N 2°36.4' W | 231 | |
| | | June 2005 | Le Catiorec, Perelle Bay | 49°27.6' N 2°38.6' W | 123 | |
| | | September 2002 and May 2004 | Lihou Causeway | 49°27.5' N 2°39.6' W | 85 | |
| | | September 2002 | Moulin Huet | 49°25.5' N 2°32.9' W | nd | |
| | | Jersey | May 2003 | La Rocque | 49°10.0' N 2°01.5' W | nd |
| | May 2003 | | Portelet Harbour | 49°10.3' N 2°10.6' W | nd | |
| | May 2004 | | Giffard Bay | 49°15.0' N 2°06.6' W | 130 | |
| | May 2004 | | Grève de Lecq | 49°14.9' N 2°12.0' W | 59 | |
| | May 2004 | | Sauchet Bay | 49°14.5' N 2°03.0' W | 67 | |
| | May 2004 | | Queen Elizabeth Castle | 49°10.5' N 2°07.4' W | 226 | |
| | May 2004 | | Rozel Harbour | 49°14.2' N 2°02.6' W | 147 | |
| | May 2004 | | Fliquet Bay | 49°13.6' N 2°01.1' W | 59 | |
| | May 2003 | | L'Archirondel | 49°12.9' N 2°01.1' W | nd | |
| | May 2004 | | Corbière | 49°10.9' N 2°14.6' W | 133 | |
| | Alderney | June 2005 | Saye Bay South | 49°43.8' N 2°11.0' W | 101 | |
| | | June 2005 | Cat's Bay | 49°43.7' N 2°09.7' W | 324 | |
| | | June 2005 | St Esquere Bay | 49°43.6' N 2°09.6' W | 290 | |
| | | June 2005 | Crabby Bay | 49°43.4' N 2°12.3' W | 119 | |
| | | June 2005 | Braye Bay | 49°43.4' N 2°11.3' W | 311 | |
| | | June 2005 | Saline Bay | 49°43.3' N 2°13.0' W | 149 | |
| | | June 2005 | Clonque Bay | 49°43.0' N 2°13.3' W | 325 | |

na, pool habitat not available; nd, no data because pool occupancy was not measured (algal thalli were surveyed rather than pools *per se*).

intertidal, emergent substratum; high densities of juveniles were found under the adult canopies (Figure 5A, B). For example, at Whitecliff Bay, we recorded a mean of 3.0 large thalli and 19.0 small ones per 0.25-m² quadrat. In contrast, thalli were small and sparsely distributed in high pools at Horse and Yellow Ledges, Shanklin (Figure 5C, D). Of 300 pools surveyed, only nine contained *C. fragile* (3%); this value was significantly greater than the pool occupancy in Devon (Figure 3A). The derbesioid or vaucheroid stage of *Codium* carpeted the bottom of the pools; utricle examination of the small fronds (Figure 5C, D), arising from the mats, confirmed that the species was *C. fragile*. This size-structure and the vaucheroid stage were seen again when we re-visited the sites in June and August 2007. The size-frequency distributions differed significantly between each pairwise combination of sites (Kolmogorov-Smirnov, $P < 0.001$ for all except one (Horse Ledge versus Whitecliff), which was $P = 0.009$).

Sites on the Channel Islands

The situation on shores in the southern English Channel differed fundamentally. Three species of upright, branching *Codium* were recorded, and all were relatively abundant. By far the most abundant taxon was *C. tomentosum* that formed patchily distributed, high-density beds at many low-shore sites around Alderney. In addition, pool occupancy by *Codium* was greater in the Channel Islands than on Devon shores and IoW shores (except for the intertidal lagoons at Bembridge) (Figure 3A; Likelihood ratio Chi-square, $G = 210.2$, $N = 4483$, 2 df, $P < 0.001$). Furthermore, there was clear tidal-level variation in pool occupancy in the Channel Islands. In June 2005, *Codium* was present in 7% of pools surveyed on Guernsey in the high-shore, *Pelvetia* zone and 27% of the pools in the mid-shore, *Fucus spiralis* zone. This pattern was highly significant (Likelihood ratio Chi-square test, $N = 594$ pools, $P < 0.001$).

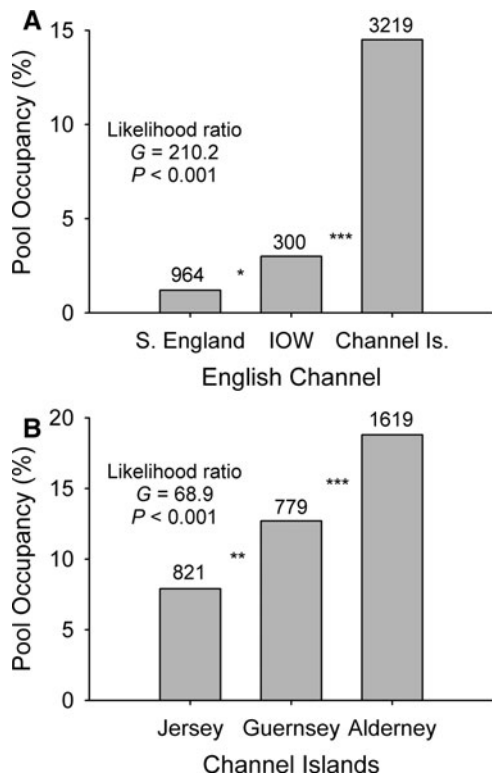


Fig. 3. Pool occupancy of *Codium* spp. in different areas of the English Channel: southern England, Isle of Wight (IoW) and the Channel Islands (CI). Survey sites and dates are listed in Table 1. Sample sizes over each bar indicate the number of pools surveyed that were used in the analyses.

At sites on Alderney with extensive freshwater seepage from land and, thus, many ulvoid-dominated pools, *Codium* spp. were significantly less frequent in ulvoid than non-ulvoid pools (Likelihood ratio Chi-square test, $G = 10.5$, 1 df, $N = 426$ pools, $P = 0.002$). On Alderney, *Codium* spp. occurred in pools at all levels and on emergent substrata on lower parts of the shore. Of the ~1600 pools surveyed in June 2005, 19% contained *Codium* spp. (Figure 3B). There was considerable variation among sites and tidal levels. In the *Pelvetia* zone, the frequency of pool occupancy was high at Clonque but low elsewhere (overall: 8.2% of pools, $N = 574$). In the *Fucus spiralis* zone, *Codium* spp. occurred in 10–40% of the pools at all sites (overall: 24.3%, $N = 609$). In the *F. vesiculosus* zone, *Codium* spp. occurred in 20–60% of the pools at the four sites with pools at this tidal level (overall: 31.2%, $N = 311$). In the *F. serratus* zone, there were few pools at most sites but *Codium* occurred in 17.5% of the 80 surveyed. *Codium tomentosum* predominated in pools. On low-shore emergent substrata, we recorded a mean of 1.2 thalli per 0.25-m² quadrat ($N = 48$ quadrats at Clonque Bay).

Both *Codium vermilara* and *C. tomentosum* thrived throughout the Channel Islands surveyed (Guernsey, Jersey and Alderney). On Alderney shores, *C. tomentosum* predominated within the *Codium* assemblages at most sites and tidal levels (Figure 6). *Codium fragile* was present at the wave-protected Braye Bay (protected by the elongate breakwater; Figure 1E) but not at many other, more exposed sites (Figure 6). *Codium vermilara* was frequent at two sites on the north side of the island, Cats Bay and St Esquere Bay, high on the shore.

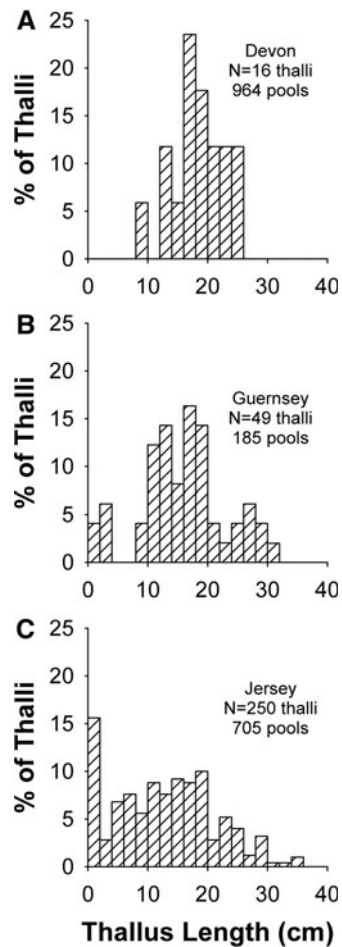


Fig. 4. Comparison of populations of *Codium fragile* in three different regions of the English Channel. All data were collected between 25 April and 9 May 2004. Devon data were from Wembury and Thurlestone; Guernsey data were from Lihou Causeway and La Jaonneuse; and Jersey data were from Sauchet Bay.

DISCUSSION

At most sites surveyed in the UK and the Channel Islands, native *Codium* spp. were equally or more abundant and widely distributed than the invasive *C. fragile*; displacement or elimination of native congeners has not occurred. *Codium tomentosum* dominated the *Codium* assemblage on Alderney (Figure 6) and Guernsey (White, 2003; Trowbridge *et al.*, 2004), and was the most frequent species in Devon. This alga is also frequent on rocky intertidal shores of western Ireland (Trowbridge, 2001 and unpublished data). *Codium vermilara* was abundant intertidally on Alderney and Jersey, and present in small amounts elsewhere in the Channel Islands; this species is abundant subtidally on the western shore of Lough Hyne, County Cork, south-western Ireland (Trowbridge *et al.*, unpublished data). In summary: (1) the displacement and replacement scenarios (Figures 1A, B & 7) have not occurred; and (2) pool occupancy in southern England does not support interspecific competition among native and introduced congeners as a causal mechanism.

Sites on the Isle of Wight

The only area we surveyed that had dense populations of *Codium fragile* was the IoW, particularly the Bembridge to

Table 2. Results of pairwise comparisons between regions of size distribution of *Codium fragile*, using Kolmogorov–Smirnov two-sample tests. Left part of table includes all data collected (N = 1159 thalli from 2002 to 2005); right part of table includes data collected from 25 April to 9 May 2004 (N = 315 thalli).

| Regions | d_{\max} 2002–2005 | | | | d_{\max} 2004 | | |
|----------|----------------------|----------|---------------------|--------|---------------------|----------|--------|
| | Devon | IoW | Guernsey | Jersey | Devon | Guernsey | Jersey |
| Devon | 0.000 | – | – | – | 0.000 | – | – |
| IoW | 0.670*** | 0.000 | – | – | – | – | – |
| Guernsey | 0.550*** | 0.486*** | 0.000 | – | 0.283 ^{ns} | 0.000 | – |
| Jersey | 0.433** | 0.445*** | 0.142 ^{ns} | 0.000 | 0.458** | 0.278** | 0.000 |

***, $P < 0.001$; **, $P < 0.010$; ^{ns}, $P > 0.050$.

Whitecliff Bay series of intertidal lagoons. Here the invasion history of the species is unclear. Based on molecular analyses of herbarium specimens (Provan *et al.*, 2007), the first verified record of *C. fragile* in England was in 1894 in Swanage, Dorset. Three sets of early records on the IoW may also indicate the unrecognized incursion of the introduced alga: (1) Delf & Grubb (1923) reported *Vaucheria velutina* C. Agardh as common on upper rocks on Shanklin Ledge (where we recorded *C. fragile*); (2) *Vaucheria litorea* C. Agardh was ‘very common on the sand-covered rocks at Horse Ledge and growing as a mat at the bottom of some rockpools’ (Norkett, 1947: 59); and (3) Foslie (1893) collected *C. tomentosum* at Steephill, IoW; Morey (1909) referred to *C. tomentosum* at Steephill and West Cowes (based on the Batters ‘Catalogue’ (Batters, 1902)); Delf & Grubb (1923) reported *C. tomentosum* from Puckaster Cove and Shanklin Ledge in tidepools; and Kain (1958) reported *C. tomentosum* from Bembridge. The first two records may refer to the derbesioid or vaucheroid stage of *Codium fragile* (see Blunden *et al.*, 1989; Fletcher *et al.*, 1989; Yang *et al.*, 1997; Trowbridge, 1998). This conjecture is based on the absence of genuine

Vaucheria during our surveys, the presence of *C. fragile* and the known historical confusion between the two taxa. This juvenile vegetative stage can persist for years on docks or the shore. While it is possible that the vaucheroid stage could have belonged to *C. tomentosum*, there are no verified records of the alga at the IoW. Either *Codium tomentosum* (*sensu stricto*) occurred on the IoW historically but not now and/or *C. fragile* was present on the IoW by the 1890s (Foslie, 1893). This conclusion is based on the current absence of *C. tomentosum* during our surveys, the absence of verified historical specimens of the native from the IoW, the known historical confusion about *C. tomentosum* (*sensu lato*) and the contemporary occurrence of *C. fragile* at most of these locations. The first author has searched historical herbaria in the British Isles; the requisite specimens were not available to distinguish between the two hypotheses. Given that the invasive *C. fragile* was present in the British Isles by the mid to late 1800s (Provan *et al.*, 2007), there is circumstantial support for the second hypothesis.

What factors would explain the contemporary, dense populations of *Codium fragile* at Bembridge and Whitecliff Bay (IoW) and the comparative scarcity of the alga elsewhere? At least four possible explanations have been suggested.

(1) The habitat *per se* may contribute to alga’s abundance at the IoW: the shallow, expansive, intertidal lagoons are ideal environments for the *Codium fragile*. Although the Wildlife and Countryside Act (1981) of the UK forbids experiments ‘planting’ listed species (such as *C. fragile*), experimental work in Korea demonstrated that growth of *Codium fragile* was greatest at shallow sublittoral depths (Hwang *et al.*, 2007).

(2) The tidal patterns around the IoW are unusual (Kain, 1958; Collins *et al.*, 1990) with double or prolonged high and low waters; spring high tides occur around midday. Thus, desiccation and light stress during early morning and early evening low tides would be comparatively less than in other areas that experience daytime low tides. Recent papers on desiccation tolerance (Schaffelke & Deane, 2005; Kim & Garbary, 2007) indicate that *C. fragile* is susceptible to desiccation but exhibits impressive recovery. On European shores, there is no direct experimental evidence that desiccation is a major problem. However, sunny weather does contribute to periodic thallus destruction and disintegration of *Codium fragile* (and congeners) in the British Isles and Japan (Trowbridge, personal observation). Similarly, cold damage to apical tips exposed to cold air in winter may influence littoral *Codium* distribution.

(3) Local diversity within a community might influence invasion success of some NIS (White & Shurin, 2007). Certainly, *C. fragile* has proliferated on north-western

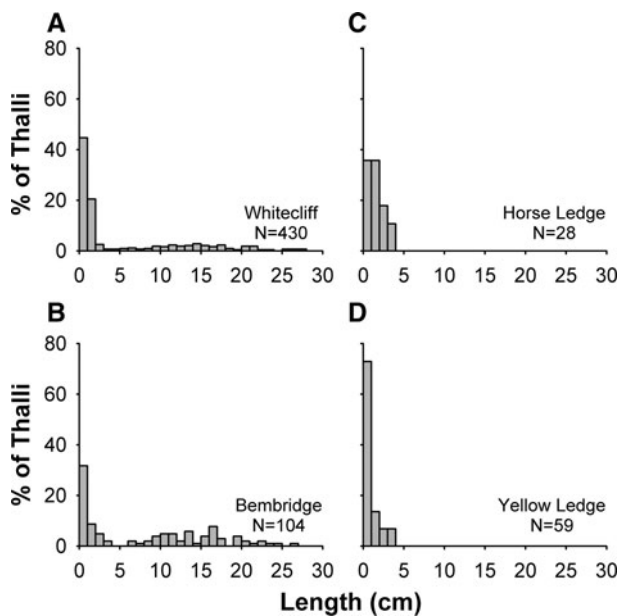


Fig. 5. Population size–frequency distributions of *Codium fragile* at four sites on the Isle of Wight. Data were collected in May 2005 with subsequent verification in September 2005 and June and August 2007. The number of thalli of *Codium fragile* sampled from each site is indicated. Identifications were confirmed by utricle examinations.

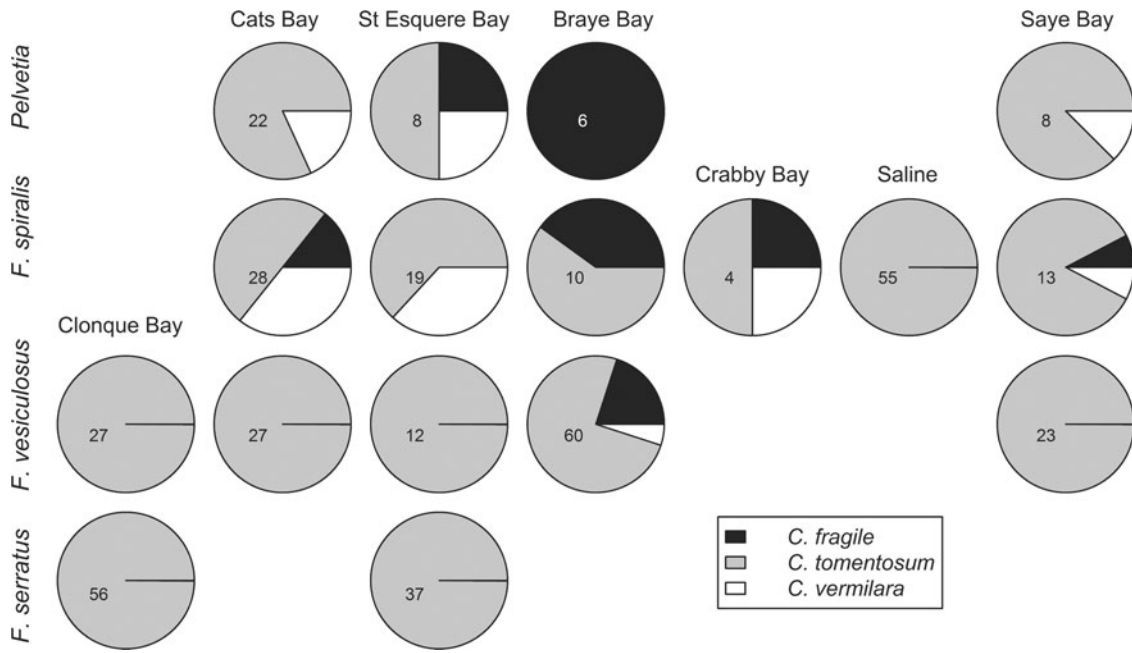


Fig. 6. Species composition of the *Codium* assemblages at seven Alderney sites and different tidal levels, based on utricle examination of thalli. Sample sizes indicate number of thalli examined microscopically. Data collected in June 2005.

Atlantic shores, a region with comparatively low species richness, compared to its response on the north-eastern Atlantic shores. The local-diversity hypothesis is not supported at Bembridge as the IoW is regarded as floristically rich (Farnham, 1982, 1994), particularly on the east coast (Figure 1C) in the intertidal lagoons. The entire coastline is listed as SMAs (Sensitive Marine Areas) by English Nature and large sections of the coastline are also SSSIs (Sites of Special Scientific Interest) due to high biodiversity.

(4) The Solent appears to be a UK hotspot of introduced macroalgae (Farnham, 1980; Collins *et al.*, 1990), in part due to extensive shipping and ferry traffic among Southampton and Portsmouth ports and continental ones. Bembridge could, thus, receive a comparatively high influx of NIS propagules relative to the Devon open-coast shoreline or to the complex, Channel Island shores.

Sites on the Channel Islands

Codium vermilara and *C. fragile* were first recorded on the Channel Islands by Dixon (1961) and Feldmann (1961). Lyle's voucher specimen of *C. tomentosum* (collected on 17 November 1921 and lodged in the Guernsey Museum and Galleries) is correctly identified (verified by Trowbridge). Although Kain (1961) made a subtidal algal collection at Braye Harbour at the breakwater, she did not record *Codium* spp., presumably due its intertidal location. Thus, the arrival date of *C. fragile* to the Channel Islands is not known.

Jersey and Guernsey both have a history of Pacific oyster cultivation, whereas Alderney does not. Given that the translocation of Pacific oysters is considered a primary vector in the spread of *C. fragile*, oyster mariculture activities may have contributed to the algal distributions. If this were the sole contributory factor, Alderney should have little or no *C. fragile*. Although Alderney does have appreciably less *C. fragile* than Jersey and Guernsey, it would be hard to accept that many decades of time (and generations of *C. fragile*) did not enable the invasive alga to disperse, establish and proliferate to most local areas.

Alternative explanations include current levels of disturbance. Jersey, as the most commercially developed of the three largest Channel Islands, presumably has the highest level of anthropogenic disturbance and it does have appreciable populations of *C. fragile*. Alderney, the least developed of the three islands, has substantial, thriving populations of *C. tomentosum*, *C. vermilara* and some *C. fragile*. However, direct experimental evidence of the stimulatory effects of anthropogenic effects (human trampling, nitrogen effluent, other pollution, mariculture and ship hull transport) has not been demonstrated anywhere in the alga's invaded range. Furthermore, the native congeners have not been experimentally manipulated to evaluate their presumed sensitivities. Anthropogenic effects do not provide a simple explanation

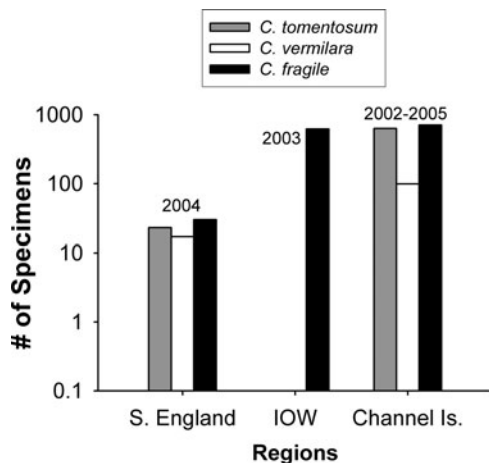


Fig. 7. Contemporary specimens of native and introduced *Codium* in southern England, Isle of Wight (IoW) and the Channel Islands. Data based on surveys made at sites and locations listed in Table 1.

that could account for the spatial variation in *Codium* populations between sides of the English Channel.

Other regional issues

Differential oceanographic conditions might account for cross-channel differences in *Codium* assemblages. The IoW is at a biogeographical boundary between western oceanic and eastern continental communities (Collins *et al.*, 1990; Herbert *et al.*, 2007). The water masses carrying *Codium* propagules to Devon shores might be hydrodynamically distinct from those moving into and around the Bay of St Malo where the Channel Islands occur. Yet, it is hard to see how this mechanism could account for the flourishing population of *C. fragile* at Bembridge.

The three Channel Islands do have dramatically different environmental situations from southern England, including large tidal ranges (9–11 m) and rapid tidal currents (e.g. Alderney Race). Native and introduced *Codium* species occur intertidally in the southern English Channel where tidal ranges are huge; the species are often in lower intertidal pools and/or in subtidal areas at the northern channel sites (Devon and IoW) where tidal ranges are much smaller (3–4 m). However, western Ireland also has a small tidal range as well as a diverse and abundant assemblage of *Codium* on many rocky shores (e.g. Trowbridge, 2001). Thus, tidal range *per se* is not a sufficient regional explanation.

Regional differences in wave action may contribute to differential patterns of littoral *Codium* distributions. *Codium fragile* can propagate vegetatively via fragmentation and subsequent attachment of fragments. Perhaps the Channel Islands have comparatively higher wave exposure and *Codium* fragmentation than Devon or IoW shores. This testable hypothesis could account for the broad-scale patterns documented in this study.

In summary, simplistic explanations of algal distributional constraints are fraught with peril. Past accounts of temporal *Codium* changes have postulated anthropogenic effects and/or ecological interactions (Figure 1); while one or more of these mechanisms may be true, the hypotheses have not been well supported scientifically. Alternative oceanographic explanations such as tidal patterns, tidal range and tidal streams have never been seriously evaluated for spatial variation in the distribution of NIS on north-eastern Atlantic shores. The classical case of an introduced species displacing a native congener is still speculative after three decades of scientific study and alternative mechanisms have been generally ignored.

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