Estimating the assimilation of mangrove detritus by fiddler crabs in Laguna Joyuda, Puerto Rico, using dual stable isotopes

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ABSTRACT. Dual stable isotope analyses (δ^{13} C and δ^{15} N) of fiddler crabs from a forest-fringed, land-locked lagoon in Puerto Rico indicated the differential assimilation of material from ingested sediments. Fiddler crabs preferentially selected for nitrogen-fixing benthic microalgae (cyanobacteria) over vascular plant detritus. These results question the assumption that mangrove detritus is always the principal source of energy to estuarine consumers. Previous research from this lagoon as well as from Amazonia suggests that the magnitude of land-water ecotonal coupling may be low for these particular tropical systems where benthic algal productivity is presumably high.

KEY WORDS: benthic microalgae, differential assimilation, fiddler crabs, mangroves, stable carbon and nitrogen isotopes

INTRODUCTION

Mangrove forests in coastal estuaries provide an invaluable habitat resource for numerous organisms including those supporting important sustainable fisheries (e.g. Lugo & Snedaker 1974). Conventional wisdom has suggested that mangroves provide the primary carbon source fueling detrital-based foodwebs, and as such, represent the major route of energy flow in such systems (e.g. Austin & Austin 1971, Flores-Verdugo *et al.* 1987, Odum & Heald 1975, Qasim & Sankaranarayana 1972, Wells 1984). Mangroves, once a prevalent feature of Puerto Rican coastlines and estuaries (e.g. Gleason & Cook 1927), are now seriously threatened: loss of 86% of original surface area cover (Lugo & Cintron 1975). Research on mangroves in Puerto Rico has centred more on understanding basic biological attributes (e.g. Cintron *et al.* 1978, Golley *et al.* 1962, Levine 1981, Pool *et al.* 1977, Wadsworth 1959) than on predicting the effects of deforestation. Quantifying the incorporation of terrestrial forest detritus into estuarine foodwebs would provide insight into the magnitude of ecotonal coupling and might be used in the managerial design of protective buffer strips around sensitive areas (France 1996a).

Although there is a long tradition heralding the importance of vascular plant detritus as a direct food source for estuarine consumers (e.g. Darnell 1961, Marples 1966), such assumptions have been criticized by others (e.g. Walker 1973). Odum & Heald (1975), however, presented evidence from an estuarine mangrove system in Florida which they believed 'refutes these [critical] arguments'. Because they found that more than 20% of material contained in the digestive tracts of all herbivores and omnivores contained mangrove detritus, Odum & Heald (1975) considered that 'the principal flow of energy is along the route: mangrove leaf detritus \rightarrow bacteria and fungi \rightarrow detritus consumers \rightarrow lower carnivores \rightarrow higher carnivores'. However, simple ingestion of detritus does not necessarily imply any direct assimilation of that material. Therein lies the advantage in using stable isotope analysis over traditional gut content analysis in determining real pathways of energy flow in mangrove ecosystems (Marguillier et al. 1997, Newell et al. 1995, Stoner & Zimmerman 1988, Zieman et al. 1984). For example, Rodelli et al. (1984) recorded mangrove detritus in the guts of offshore animals in Malaysia for which there was no isotopic evidence for any mangrove assimilation.

Because consumer δ^{13} C values became less negative with distance from mangrove swamps to offshore waters, Rodelli *et al.* (1984) interpreted this effect as being due to the progressive reduction in assimilation of mangrove detritus. By focusing on only fiddler crabs, and extending the analysis to include more recent global data, a similar trend can be seen whereby δ^{13} C values are more negative in forested estuaries than in open marine-influenced areas (Figure 1).

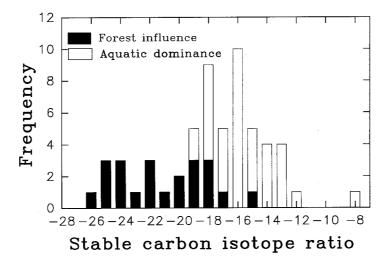


Figure 1. Frequency distributions of stable carbon isotope ratios ($\delta^{13}C_{\infty 0}$) from the literature for various species of fiddler crabs collected from regions of either aquatic dominance (e.g. seagrasses, saltmarshes, n = 22) or riparian forest influence (n = 36). Data sources available upon request.

Interpreting such findings as unequivocal evidence for the direct assimilation of mangrove detritus may be incorrect. This occurs because the seaward enrichment in ¹³C could also result from the diminishing outwelling of respiratory CO_2 that is produced by heterotrophic activity associated with decomposing mangrove detritus; i.e. all organisms, regardless whether they are autotrophs or consumers, with or without any likelihood of ingesting mangroves directly, are ¹³C-depleted in mangrove swamps (Fleming *et al.* 1990, France & Holmquist 1997, Hemminga *et al.* 1994, Lin *et al.* 1991, Marguillier *et al.* 1997, Zieman *et al.* 1984), just as they are in other areas when exposed to high rates of decomposition, microbial activity and respiration (e.g. del Giorgio & France 1996, France 1996b, France *et al.* 1997). The purpose of the present study is to use a dual isotope approach ($\delta^{13}C$ and $\delta^{15}N$ analyses) in an attempt to determine whether mangrove detritus is actually incorporated into a population of fiddler crabs.

Ocypodid (fiddler) crabs are ideal subject organisms in which to measure the relative importance of allochthonous and autochthonous dietary contributions. Fiddler crabs are conspicuous residents of the land-water ecotone in many mangrove swamps (Macintosh 1982, Montague 1980, Warner 1969), being behaviourally terrestrial but physiologically aquatic. Fiddler crabs feed by scooping sediment into their buccal cavities where water from the branchial chamber is used to suspend organic material (Miller 1961). This material adheres to the maxillipeds bordering the mouthparts as water is withdrawn, and is then transferred to the mouth via the maxillae. Heavier inorganic particles settle to the floor of the buccal cavity and are formed into a pellet for removal. Therefore, by having small chelipeds adapted to filtering small particulate matter such as algae, bacteria and detritus from the surface of the substratum, fiddler crabs may be true generalists, facultatively capable of preferentially assimilating either benthic algae or vascular plant detritus under different situations (Crane 1975) as superficially implied by Figure 1.

STUDY AREA

Laguna Joyuda is a tropical lagoon situated on the west coast of Puerto Rico (18° 07' N latitude, 67° 07' W longitude) 8 km south of the city of Mayaguez (see maps in Stoner 1986, 1988) within a Subtropical Moist Forest that receives up to 200 cm of annual rainfall (Levine 1981). Surface area of the lagoon is 1.21 km^2 (122 ha) within a total drainage area of 5.95 km^2 . Average lagoon depth is 1.5 m with two deeper regions of 2.5 m (Armstrong 1978). The lagoon is thought to have been formed about 500 y ago through the accretion of two sandbanks which enclosed the small embayment (Comer 1969). Today, a single small channel (3 m mean width; <0.5 m mean depth) connects the lagoon with the ocean at Passage de la Mona located 0.5 km away. The lagoon is therefore almost completely landlocked and can be referred to as being 'choked' (sensu Kjerfve 1986). Tidal ranges are less than 15 cm (Stoner 1988) with volumes of

 3.5×10^3 m³ over 12 h, a very small amount compared with the total lagoon volume of 1.8×10^6 m³ (Armstrong 1978). The lagoon is brackish with long-term salinities ranging from between 4 to 44‰ (Stoner & Lopez 1985) and averaging between 20 to 30‰ (Armstrong 1979, Pagan & Austin 1970). Water temperature ranges from 24 to 32°C annually. Due to strong winds, the lagoon is fairly homogeneous both longitudinally and vertically with regard to temperature, salinity and dissolved oxygen.

Laguna Joyuda is nearly completely fringed by red mangroves (*Rhizophora* mangle) ranging in riparian width from 5 to 40 m. Individual trees are often over 22 m in height, and thus among the tallest specimens of this species found in Puerto Rico. Annual litterfall approaches 2000 g dw m⁻², of which half (2.6 g dw m⁻² d⁻¹) is from mangroves (Levine 1981). These values are greater than those of many other mangrove systems, and were considered by Levine (1981) to be the highest leaffall rates ever recorded for any forest ecosystem. All litterfall is decomposed within a year's submersion (Levine 1981) and contributes substantially to the high organic content of the sediments (Comer 1969). Seagrasses (*Thalassia testudinum*) cover only a small portion of the lagoon bottom, possibly as a result of secchi depths being consistently below 0.6 m.

Over 40 species of fishes, 12% of which are year-round residents (Stoner 1986), and up to 10 species of large crustaceans (Stoner & Zimmerman 1988), including four penaid shrimps (Stoner 1988), inhabit the lagoon. Some of these species support a local fishery for nearby restaurants. Severe drought or heavy rainfall events can cause sudden changes in salinity and temperature and consequent periodic kills of fish and both pelagic and benthic invertebrates (Pagan & Austin 1970). Stoner & Zimmerman (1988) conducted a preliminary δ^{13} C survey of selected invertebrates of the lagoon.

METHODS

Sampling and analysis

During September 1994, 80 fiddler crabs (*Uca vocator*) of a wide size range (5.5–22.6 cm carapace width) were hand-collected from underneath the intertidal forest along the southern shoreline of Laguna Joyuda. Such a high sampling effort was adopted to account for the possibility, disregarded in most studies (France 1996d) due to specimen pooling (e.g Haines 1976), of ontogenetic diet shifts (France 1996c), as suggested by previous observations of intraspecific variability exceeding 4‰ in δ^{13} C for fiddler crabs (Rodelli *et al.* 1984). Twenty samples of red mangrove leaf litter of two types (freshly senescent, still green; and yellow with a copious microbial biofilm) were sampled from water depths of several cm. Twenty samples of intertidal sediment composed of both small particles of organic detritus and inorganic grains were collected from the water's edge.

All samples were frozen (after allowing time for gut clearance for the crabs), dried, pulverized, acid-washed in dilute hydrochloric acid to remove inorganic carbonates, redried, weighed and packed in tin capsules. Samples were measured for stable isotopes and C/N ratios on a Europa tracermass-spectrometer interfaced with a Roboprep-CN analyzer. An average SD of \pm 0.3‰ was obtained for replicate sub-samples after correction for weight-related bias. Samples were run in 12 sequences of five per tray, each sequence being separated by paired standards (i.e. a blank to clear the combustion tube followed by an identical standard), with duplicates never occurring in the same run sequence. If there were differences of >1‰ between the values of paired duplicates, these samples were disregarded (this occurred for 12 crabs, three mangrove samples and four sediment samples).

Background data for autochthonous food sources

Aquatic autotrophs display considerable variability in their isotopic composition in relation to such environmental factors as light, temperature, season, turbulence and water chemistry (e.g. Boon & Bunn 1994; France *et al.* 1997, Fry *et al.* 1986). Consequently, it may be difficult to speculate about the use of various food sources from the basis of individual studies in which the isotopic variability of aquatic autotrophs may often be underestimated through inadequate sampling replication (France 1995a, 1996e). In such cases, a compilation of literature values can provide some indication of the range of presumed food sources which often bracket the few study-specific measurements that were obtained (e.g. Boon & Bunn 1994, Currin *et al.* 1995). Therefore, for this reason (as well as the logistic limitation that no watercraft were available), the isotopic composition of autochthonous autotrophs in Laguna Joyuda was estimated from the literature. Whenever possible, values for mangrove estuaries were used over simple global values.

RESULTS AND DISCUSSION

Background data for autochthonous food sources

Plankton δ^{13} C in Laguna Joyuda (Stoner & Zimmerman 1988) averaged $-25.8 \pm 1.1\%$. Seagrass δ^{13} C from mangrove swamps averaged $-14.1 \pm 2.7\%$, similar to the single value measured by Stoner & Zimmerman (1988) in Laguna Joyuda of -16.1 for *Thalassia testudinum*. Macro- and microalgal δ^{13} C from mangrove swamps averaged $-19.9 \pm 3.3\%$, similar to the two values of -14.2 and -20.2 determined respectively for *Spirulina* sp. and *Caulerpa sertularoides* in Laguna Joyuda by Stoner & Zimmerman (1988). As expected (France 1995b), due to differences in boundary layer diffusion resistance (Hecky & Hesslein 1995; France & Holmquist 1997), δ^{13} C values of attached algae in Laguna Joyuda are enriched compared with those of plankton. All these δ^{13} C values for autotrophs in mangrove estuaries are 13 C-depleted relative to similar autotrophs (France 1995b, 1996f) sampled away from the heterotrophic activity associated with mangrove detritus decomposition (cf. Marguillier *et al.* 1997). Plankton and macroalgae were presumed to have δ^{15} N values of about 6.3 ±

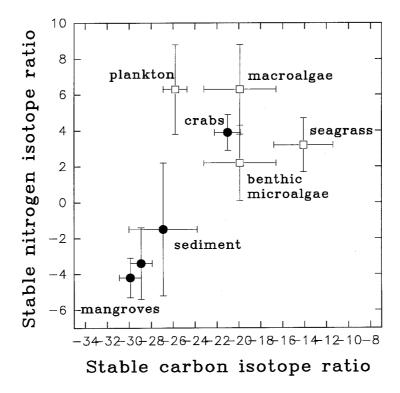


Figure 2. Stable carbon and nitrogen isotope ratios (\pm SD) of fiddler crabs, sediments and mangroves from Laguna Joyuda, and of presumed autochthonous autotrophs obtained from the literature (see text).

2.5‰ based on worldwide marine data (France 1995c, France *et al.* 1998). Average seagrass δ^{15} N values (3.2 ± 1.5‰) are generally low compared to those of plankton and macroalgae as a result of nitrogen fixation (e.g. Fry *et al.* 1986). Likewise, the presence of nitrogen-fixing cyanobacteria in estuarine mudflats (Pulich & Scalan 1987) also reduces the average δ^{15} N values of benthic microalgae (2.2 ± 2.1) compared to plankton and macroalgae (Currin *et al.* 1995).

Mangroves and sediment

Freshly senescent, still green, mangrove leaves had δ^{13} C values of $-29.9 \pm 1.0 \text{ (SD)}\%$, and when covered by a thick biofilm growth after loosing their green photopigments, were $-28.9 \pm 1.0\%$ (Figure 2). These values are close to those recorded for various mangrove species in other studies (Marguillier *et al.* 1997, Newell *et al.* 1995, Rao *et al.* 1994, Rezende *et al.* 1990, Rodelli *et al.* 1984, Zieman *et al.* 1984), including two measurements previously determined by Stoner & Zimmerman (1988) for green (-29.2‰) and dead (-27.3‰) red mangrove leaves from Laguna Joyuda. Freshly senescent leaves displayed δ^{15} N values of $-4.2 \pm 1.1\%$, whereas microbially colonized leaves averaged $-3.4 \pm 1.6\%$. These values are representative of those found for about a quarter of all terrestrial plants (France 1995c) as well as the few values

determined by Zieman *et al.* (1984) for decomposing red mangroves in Florida, and are only slightly lower than those determined by Marguillier *et al.* (1997) for fresh leaves from nine mangrove species in Kenya. Zieman *et al.*'s (1984) δ^{15} N values of 6 to 8‰ for fresh mangroves are very high compared to the synthesis of worldwide data for terrestrial plants presented in France (1995c). Newell *et al.*'s (1995) δ^{15} N values of 5.40 ± 0.66 for mangroves are also higher than those found in the present study and may be a consequence of the mangroves in Laguna Joyuda deriving their inorganic nitrogen from the brackish water.

Littoral sediment in Laguna Joyuda averaged $-26.9 \pm 3.1\% \delta^{13}$ C and $-1.5 \pm 3.7\% \delta^{15}$ N (Figure 2). These values are enriched for both isotopes compared with those for the major source material of mangroves, as Rao *et al.* (1994) also found for δ^{13} C in an African coastal lagoon. This is to be expected, as Haines (1976) similarly observed, since all sources of autochthonous organic matter are substantially enriched relative to those of terrestrial plants (Figure 2). The rarity of seagrasses in Laguna Joyuda is indicated by the differences in δ^{13} C values for these sediments compared with those found in other lagoons characterized by luxurious seagrass growth (e.g. Fry *et al.* 1977). Littoral sediment in small water bodies such as Laguna Joyuda often assume δ^{13} C values representative of fringing riparian trees (e.g. LaZerte 1983). Stoner & Zimmerman's (1988) δ^{13} C values of '-22.9 to -25.0' for 'detritus particles from the sediment' (no further details provided) are slightly higher than those recorded here for the complete organic sediment complex.

Fiddler crabs

Fiddler crabs from Laguna Joyuda averaged $-21.0 \pm 1.2 \delta^{13}$ C (Figure 2), a value consistent with the single previous measurement by Stoner & Zimmerman (1988) of -23.0% as well as for the global average for various species collected from other areas of forest influence (Figure 1). What is most interesting from these findings is the large disparity of 6‰ between δ^{13} C for crabs and that for the sediment upon which they feed. Obviously, a considerable selective assimilation of specific particles is taking place following ingestion of the bulk sedimentary organic matter. Other studies have also observed consistent differences of about $4.4 \pm 1.4\%$ between substrate δ^{13} C values and those of fiddler crabs in saltmarshes (Haines 1976, Haines & Montague 1979). Isotopically, fiddler crabs appear to be an exception to the adage that 'you are what you eat'.

Fiddler crabs from Laguna Joyuda averaged $3.9 \pm 1.0\% \delta^{15}N$ (Figure 2), substantially below the global mean of 9‰ for marine zoobenthos (France 1994), thereby raising the possibility of localized influences of nitrogen fixation. Stable nitrogen isotopic data for fiddler crabs are few and limited to saltmarshes. In one such system with no evidence of nitrogen fixation, a single $\delta^{15}N$ value for a fiddler crab was 7.2‰ (Sullivan & Moncreiff 1990), whereas in two other systems with extensive microbial activity in the sediments, fiddler crabs were found to average $4.6 \pm 1.2\%$ (Peterson & Howarth 1987) and $3.6 \pm 0.5\%$ (Currin *et al.* 1995) δ^{15} N, similar to values from Laguna Joyuda.

Food source provenance

Opinions differ as to the relative importance of either predominantly benthic algal (Shanholtzer in Crane 1975, Schwartz & Safer 1915) or mangrovebacterial detrital (Crane 1975, Macintosh 1982, Montague 1980, Odum & Heald 1975, Teal 1958, Whiting & Moshiri 1974) energy sustenance for fiddler crabs. As Currin et al. (1995) also found for a North Carolina saltmarsh, the low $\delta^{15}N$ values of fiddler crabs in Laguna Joyuda (Figure 2) virtually rule out any substantial dietary contribution from either plankton or benthic macroalgae given the expected (Cabana & Rasmussen 1994) 3 to 4‰ enrichment in ¹⁵N per trophic fractionation. Dissimilarities in δ^{13} C between benthic animals such as fiddler crabs and those of plankton are not surprising considering the widespread virtual uncoupling that often exists between benthic and pelagic foodwebs in many aquatic systems (France 1996g). Likewise, differences in δ^{13} C between fiddler crabs and seagrasses in Laguna Joyuda (Figure 2) are to be expected as such plants rarely contribute substantially and unilaterally toward the diets of benthic animals (France 1996f). Fiddler crabs in Laguna Joyuda, therefore, can most closely be associated with a presumed benthic microalgal food source (Figure 2).

The bottom substrate in Laguna Joyuda is covered by a thick algal mat comprised predominantly of the blue-green *Sprirulina* sp. which was implicated by Stoner & Zimmerman (1988) as being the major food source for various species of polychaetes (average δ^{13} C value of -21.7%), molluscs (-19.7%) and other crustaceans (-16.5%). Similarly, nitrogen-fixing blue-greens were believed to be the major food source for fiddler crabs in a North Carolina saltmarsh (Currin *et al.* 1995) as well as for various arthropods in a Mexican tidal mudflat (Pulich & Scalan 1987).

Fiddler crabs in Laguna Joyuda varied in δ^{13} C from -24.3 to -18.7‰ and in δ^{15} N from 1.4 to 5.8‰. There were no relationships between the isotopic composition of individual crabs and their C/N ratios (Table 1), indicating that differences in presumed lipid contents cannot be invoked to explain the isotopic variability. This is similar to the situation found for benthic consumers in north-temperate lakes (France 1996h). There was no ontogenetic shift in δ^{13} C

Table 1. Correlation matrix between size, inferred lipid content as C/N ratios, and stable carbon and nitrogen isotopic composition for 68 fiddler crabs from Laguna Joyuda, Puerto Rico. P values are all >0.05 except for $\delta^{15}N$ vs. $\delta^{13}C$ which was <0.01.

	C/N	$\delta^{_{13}}C$	$\delta^{\rm 15}N$
Size C/N δ ¹³ C	0.23	$\begin{array}{c} 0.14 \\ 0.08 \end{array}$	0.28
C/N		0.08	0.10
$\delta^{13}C$			0.81

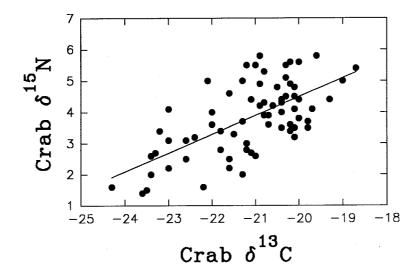


Figure 3. Relationship between carbon and nitrogen stable isotope ratios for fiddler crabs from Laguna Joyuda, Puerto Rico (r = 0.81, P < 0.01, n = 68).

with fiddler crab body size (Table 1) as was the case for crayfish in north-temperate lakes (France 1996c). Measurements of δ^{13} C and δ^{15} N among individual crabs were, however, significantly correlated (Table 1; Figure 3), implying the variable assimilation of algal and mangrove food.

Fiddler crabs with δ^{13} C values above -20% also display higher δ^{15} N (Figure 3) values suggesting the predominant assimilation of benthic microalgae. Individuals with lower δ^{13} C values closer to those of mangroves also exhibit lower δ^{15} N values as would be expected given the proportionally greater assimilation of mangrove detritus. The present results therefore support the contention that for species residing in ecotones, as for example between freshwater and marine environments (Peterson *et al. in* France 1994; Hesslein *et al. in* France 1995d), between freshwater and terrestrial environments (France 1996i) or between terrestrial and marine environments (Peters *et al.* 1978; this study), δ^{15} N values are clearly not inviolate markers of foodweb position alone but rather reflect some combination of both trophic and food source fractionations (France 1995c). For example, crabs from a Japanese estuary (Wada *et al.* 1993) and martens from coastal Alaska (Ben-David *et al.* 1997) also demonstrated correlations between both stable isotopes, presumably reflecting variable marine-freshwater/terrestrial feeding.

There are a few other examples of consumers that indiscriminately ingest detrital matter yet display δ^{13} C values distinct from those of the bulk mixture, thereby indicating the preferential assimilation of a particular food source from among the many available within the total organic matter pool (amphipods avoiding mangrove or macrophyte litter – Schlacher & Wooldridge 1996, Zieman *et al.* 1984; oysters selecting for benthic diatoms despite the small

proportion of the latter in the total pool of organic matter – Riera & Richard 1996; zooplankton avoiding terrestrial detritus for phytoplankton in oligotrophic systems dominated by high allochthonous inputs – del Giorgio & France 1996; and polychaetes selectively utilizing algal components of sediments – Kikuchi & Wada 1996).

Several previous researchers, using gut content analysis, have contended that fiddler crabs rely upon a mixture of both benthic algae and mangrove detritus (Crane 1975, Teal 1962) but that 'no definite statement, however, has been made concerning the relative use of each of these foods' (Montague 1980). The present results (Figure 2) for Uca vocator from Laguna Joyuda, Puerto Rico, indicate a proportionally greater contribution of benthic algae to diets than mangrove detritus. This is counter to Odum & Heald's (1975) belief that animals in mangrove estuaries 'derive their nourishment primarily from a diet of vascular plant detritus and small quantities of fresh algae'. Other isotope studies from seagrass meadows and saltmarshes also indicate that, given the choice, most consumers will rely more substantially upon algae than vascular plants (e.g. Currin et al. 1995, Fry 1984, Kitting et al. 1984, Schlacher & Wooldridge 1996, Sullivan & Moncrieff 1990). The present results support other isotopic work which has suggested that aquatic consumers in or near tropical forests assimilate a proportionally greater contribution of benthic algae than they do forest detritus when the former display a high productivity (Arujo-Lima et al. 1986, Forsberg et al. 1993, Hamilton et al. 1992, Newell et al. 1995, Stoner & Zimmerman 1988).

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