

Ligula intestinalis (Cestoda: Pseudophyllidea): an ideal fish-metazoan parasite model?

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SUMMARY

Since its use as a model to study metazoan parasite culture and *in vitro* development, the plerocercoid of the tapeworm, *Ligula intestinalis*, has served as a useful scientific tool to study a range of biological factors, particularly within its fish intermediate host. From the extensive long-term ecological studies on the interactions between the parasite and cyprinid hosts, to the recent advances made using molecular technology on parasite diversity and speciation, studies on the parasite have, over the last 60 years, led to significant advances in knowledge on host-parasite interactions. The parasite has served as a useful model to study pollution, immunology and parasite ecology and genetics, as well as being the archetypal endocrine disruptor.

Key words: *L. intestinalis*, endocrine disruption, pollution, immunology, parasite ecology, parasite speciation.

THE PARASITE

There is perhaps no other metazoan parasite that evokes such awe and revulsion within the fishing fraternity world-wide than the plerocercoid stage of the pseudophyllidean cestode, *Ligula intestinalis*. This is primarily due to the impressive size which this parasite can obtain in the body cavity of its fish intermediate host, and yet it is the size of the parasite and its relationship with its fish host which has made this parasite such a valuable model to study parasite/host interactions at the molecular, cellular, organismal and population levels. As a highlight of the size which the infection can attain, Barus and Prokes (1994) noted that the weight of parasite tissue can be greater than that of the fish tissue (Fig. 1). This relationship has been studied utilising the parasitisation index (parasite weight/fish weight \times 100) which normally lies in the range of 1–20% (Claridge *et al.* 1985) or up to 40% as recorded by Morrison (1977). It is perhaps not surprising therefore that the infection has been associated with a distension of the body wall which leads to separation of the scales which may allow entry of pathogens (e.g. Sweeting, 1977), and effects on the body wall and musculature (Richards and Arme, 1981; Loot *et al.* 2001*c*). In some rare instances, perforation of the body wall and intrusion of the parasite into the

aquatic environment has been noted (e.g. Barus *et al.* 1997). The authors proposed that this may represent a possible means for a free-living phase of the plerocercoid or, at least, prolong the window for transmission into a definitive host for several days after death of the host.

The plerocercoid of *L. intestinalis* occurs in a wide range of fish hosts. In Europe it has been found in several species of cyprinids, e.g. roach (*Rutilus rutilus*), rudd (*Scardinius erythrophthalmus*), dace (*Leuciscus leuciscus*), gudgeon (*Gobio gobio*), bream (*Abramis brama*), bleak (*Alburnus alburnus*), minnow (*Phoxinus phoxinus*), chub (*Leuciscus cephalus*), tench (*Tinca tinca*) and silver bream (*Blicca bjoerkna*) (Orr, 1968; Arme and Owen, 1968; Harris and Wheeler, 1974; Adamek *et al.* 1996; Museth, 2001; Loot *et al.* 2001*a, b, c, d*, 2002*a, b*, 2006; Hecker *et al.* 2007). Dubinina (1966) noted 49 species of fish in the former USSR were hosts of the parasite, whilst there are several authors who have indicated the presence of the worm in the North American continent. For example, in Canada, Szalai *et al.* (1989) noted the parasite in white suckers (*Catostomus commersoni*), yellow perch (*Perca flavescens*), quillback (*Carpoides cyprinus*) and spottail shiners (*Notropis hudsonius*), and Groves and Shields (2001) and Shields *et al.* (2002) found the parasite in the Crooked River system and Haystack Reservoir in central Oregon. *Ligula* has also been found in the Middle East (Ergonul and Altindag, 2005*a, b*; Sasi, 2005; Kir and Tekin-Ozan, 2005; Hatice *et al.* 2006; Shargh *et al.* 2008; Hajirostamloo, 2008; Aydogdu *et al.* 2008; Tekin-Ozan and Kir, 2008),

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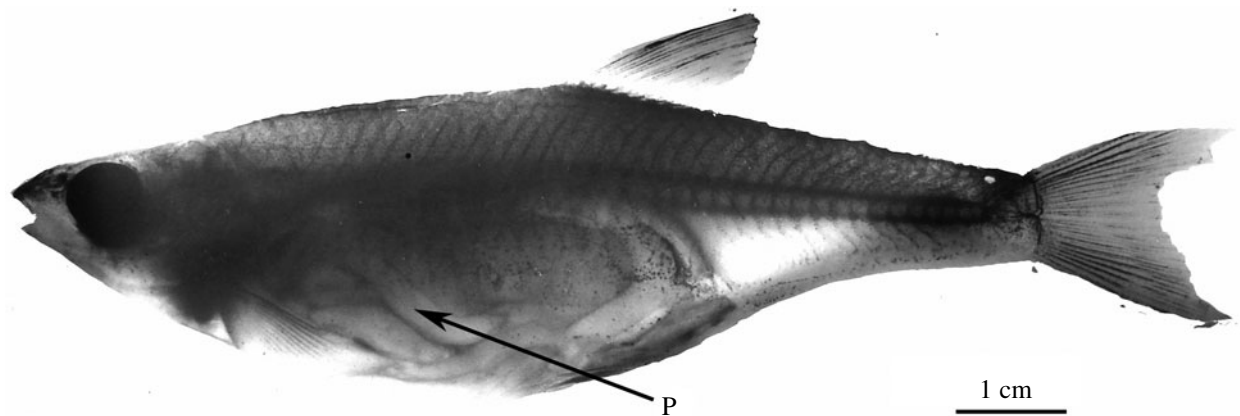


Fig. 1. Ligulosed roach, *Rutilus rutilus*, fixed in Bouins fixative and cleared in xylene. Note plerocercoid burden (P) within the body cavity of the fish host. (Picture reproduced by kind permission of Professor C. Arme, Keele University, UK.)

Africa (Dejen *et al.* 2006; Cowx *et al.* 2008; Britton *et al.* 2009) and Australasia (Pollard, 1974; Weekes and Penlington, 1986; Morgan, 2003). In China, the parasite can cause problems in culture systems (Xianghua and Zhixin, 1987; Li and Liao, 2003). The large range of fish species that the parasite has been recorded from is also extended to the association between the definitive host and the adult parasite, for example, Dubinina (1966) claimed 72 species of bird can serve as final hosts for the Ligulidae. Dubinina has also given an account of the development cycle of *Ligula* in which parasite eggs, which first appear in the uterus 45–50 hours after infection, pass out in the birds' faeces and hatch after 5–8 days to release the free-swimming coracidia. These are ingested by the copepod first intermediate host in which the proceroid develops, and the fish becomes infected by consuming the infected copepod. The ubiquitous nature of *Ligula* in terms of host fish and geographical range has meant that this parasite has been a good model to study speciation and diversity in fish parasites.

LIGULA AS AN ECOLOGICAL MODEL

Although there have been numerous studies on the ecological interactions between fish parasites and their hosts, the majority have concentrated on limited time spans with, on some occasions, speculation on the long-term implications of infection. None of the studies undertaken, except some of those carried out on *Ligula*, have considered what happens to the epizootic and the implications for the aquatic community over an extended time period. Although some studies (Bauer and Stolyarov, 1961; Black and Fraser, 1984; Izyumova, 1987) have revealed that *Ligula* may persist for several years within a single water body, the majority of studies, for example those carried out in the UK (Wilson, 1971; Sweeting, 1976; Morrison, 1977; Tobin, 1986; Bean and Winfield, 1992), have noted that a decline of fish host

populations leads to reduction in transmission rates of the parasite. This reduction is thought to be due primarily to an increase in fish mortality in infected fish caused by either a reduction in their ability to survive over winter (Wyatt and Kennedy, 1988) and/or by making the infected fish more susceptible to predation both by birds and fish (Van Dobben, 1952; Holmes and Bethel, 1972; Sweeting, 1976; Hoole, 1994). Studies on metazoan parasite/host systems in natural environments are invariably, due to funding restrictions or available opportunities, restricted to short time periods. The effects of *Ligula* on host populations noted above make this parasite a unique model to study long-term effects of parasitisation on host dynamics. This has been achieved by several eloquent studies carried out by Kennedy and co-workers over a period of 31 years utilising *Ligula* infections in Chew Lakes, Slapton Ley, Devon, UK (Kennedy *et al.* 2001). The events that pre-disposed the lake for introduction of *Ligula* began in the 1960s when increasing eutrophic conditions and an expansion of the roach population led to intra-specific competition and the presence of numerous small, stunted roach and a decline in the dominant rudd population. This change in the fish population dynamics and availability of an ample small sized food source attracted Great Crested Grebes, *Podiceps cristatus*, one of the definitive hosts for *L. intestinalis* and it was not surprising to note the appearance of ligulosed roach in 1973. The prevalence peaked at 28% in 1975 and resulted in a decline in the roach population and a subsequent recovery in the rudd numbers. This was the first cycle of *Ligula* infection whose details were described by Burrough and Kennedy (1979), Kennedy and Burrough (1981) and Wyatt and Kennedy (1988, 1989). The decrease in roach numbers led to increased growth rates, but adverse winter conditions during 1984–1985 caused substantial fish deaths and the parasite re-appeared in 1991 in the 1989, 1990 and 1991 year classes. In this second wave of infection, prevalence levels were

very high i.e. 75% peaking in 1992. A decline in infection occurred between 1993–1994. It would appear that a third cycle also occurred from 1999 and although infection occurred in 0, 1 and 2 year old fish, prevalence never exceeded 14%. The studies on Slapton Ley are therefore unique, not only in their duration, but also in the fact that 3 distinct infection cycles were clearly observed although this does not imply any regulation or stability within the system as infection cycles varied. In previous short-term studies (Arme and Owen, 1968; Harris and Wheeler, 1974; Sweeting, 1976, 1977; Bean and Winfield, 1992), biotic factors such as fish feeding behaviour and copepod levels were suggested to affect *Ligula* predominance in smaller fish. In contrast, in the second cycle at Slapton Ley, fish of all ages may have been infected. Kennedy *et al.* (2001) proposed that whilst the long-term study may suggest controlled population regulation, the first and second cycles were in fact independent events and the cycles reflect changes in habitat. Recently *Ligula intestinalis* has proved an invaluable model to elucidate how parasites may modify their host phenotypic appearance, and the impact that this has on the ecological interactions between parasites and their hosts, and transmission success. In 1999, Lafferty defined a phenomenon which resulted in an increase in parasite fitness mediated by evolutionary processes, the so-called 'Parasite Increased Trophic Transmission – PITT'. Extensive studies carried out since 2000 by Loot and co-workers on ligulosed roach populations in France have made a significant contribution to testing the PITT hypothesis in natural systems. Detailed morphological investigations on roach collected from the Lavernose-Lacasse gravel pit complex in Toulouse revealed that extensive morphological changes occurred in ligulosed roach which were dependent on total parasite biomass (Loot *et al.* 2001c). Further studies by Loot *et al.* (2001d) showed that in the Lavernose, Muret and Pareloup lake systems the distribution of the plerocercoid of *L. intestinalis* was spatially and temporally clumped within their fish hosts although this aggregation was far more pronounced in the Lavernose Lake suggesting the presence of site-related effects. Loot and co-workers (2001a) also noted that ligulosed roach were more highly parasitised closer to the bank and that parasite occurrence and abundance were both highly significant parameters in accounting for this spatial distribution. They also showed that these two parameters decreased in fish older than 3 years of age, suggesting that *Ligula* infection resulted in host death. Three hypotheses were considered to explain why parasite infection was greatest near the bank of the lake. Whilst the possibilities that this infection distribution was due to an accidental side-effect of parasitisation or perhaps was correlated to the localisation of infected first intermediate copepod hosts were

considered, the authors proposed that *Ligula* increased feeding motivation into the highly productive littoral areas of the lake and the reduced swimming efficiency increased predation by the bird definitive host. These results thus support the PITT hypothesis. Detailed studies on the interaction between *L. intestinalis* and its fish host have primarily concentrated on the Euro-Mediterranean clades and it is only recently that the PITT hypothesis has been investigated in other geographical localities. Britton *et al.* (2009) investigated *Ligula* host specificity in the fish population of Lakes Baringo and Naivasha in Kenya's Rift Valley. It was noted that the parasite had a restricted second intermediate host range occurring in two cyprinid fish species, *Barbus lineomaculatus* and *B. paludinosus*, and, in comparison to the Euro-Mediterranean clade, multiple infections were not observed frequently. The authors suggested that since parasite prevalence was correlated to habitat the results obtained could be interpreted as supportive of the PITT hypothesis. The complex life cycle of *L. intestinalis* has also recently been utilised to establish the contributing role that host ecological dynamics and physiological compatibility has on parasite transmission and evolution. Loot *et al.* (2006), using measurements of host abundances over time, noted that in the Lavernose-Lacasse gravel pit system used in earlier studies, the favoured *Ligula* hosts were the copepod, *Eudiaptomus gracilis*, the roach *Rutilus rutilus* and the great crested grebe (*Podiceps cristatus*). It was suggested that the similar temporal dynamics and frequent associations between hosts and parasite created a stable system that promoted the successful completion of the *Ligula* life cycle. Such an association, it was proposed, affected the evolution of *L. intestinalis* specificity. It was hypothesised that the selection of hosts was driven by the probability of successive hosts encountering each other which aided parasite transmission and whilst, biochemical compatibility between host and parasite was important, this may be secondary to spatial and temporal dynamics.

There have been extensive studies on the pathological and biochemical interactions between *Ligula* and its fish hosts (for example see Arme *et al.* 1982; Hoole, 1994) which have primarily been stimulated by the large parasite burdens that can occur in some hosts and the subsequent extensive size that infected hosts can attain. The possibility that host gigantism occurs in *Ligula* infection has thus been recently investigated by Loot *et al.* (2002b) who revealed that the parasite induced an enhanced growth rate during the first 2 years of the life of the roach host. Whilst the possibility that the parasite produced growth enhancers or may divert energy from gonadal development to somatic growth were considered, the authors proposed that since cestode-associated growth in the heaviest infected fish was only found in one out of the three localities studied, that the

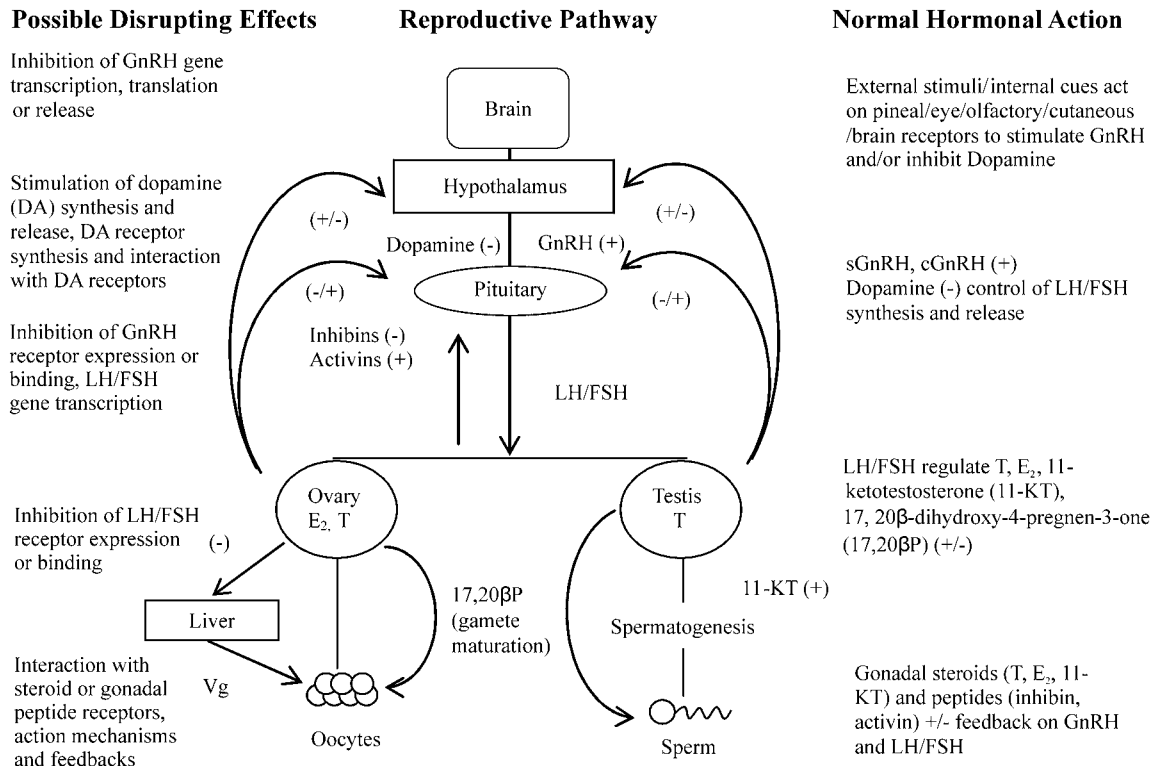


Fig. 2. Possible actions of *Ligula* on the roach reproductive pathway, + = stimulation, - = inhibition.

effect may result from a change in foraging behaviour of the fish. What these and other studies have highlighted however is the intricate association between *Ligula* and the reproductive potential of its fish host.

LIGULA: THE MODEL OF ENDOCRINE DISRUPTION

Perhaps the most significant impact of *Ligula* on the fish host population is its ability to inhibit reproductive function of its host. Whilst it has been established previously that parasite effects on host reproduction occur in several host/parasite interactions, for example the snail, *Lymnaea stagnalis* infected with *Trichobilharzia ocellata* (Joose and van Elk, 1986); *Schistosoma mansoni* infections of *Biomphalaria glabrata* (Crews and Yoshino, 1989) and in vertebrate hosts – the tapeworms *Taenia taeniaeformis* affect the testis in rats and *T. crassiceps* induces feminisation in mice (Lin *et al.* 1990; Larralde *et al.* 1995, respectively) – it is the intricate relationship between *Ligula* and the reproductive endocrine status of the fish that has fascinated and perplexed biologists. Studies by Kerr (1948) and Arme (1968) noted that in ligulosed roach, putative gonadotrophs were reduced in number and had a reduced granular content than their uninfected counterparts. This effect on reproductive potential has been reported subsequently by Arme and Owen

(1968), Mahon (1976), Sweeting (1977), Bean and Winfield (1989) and Cowx *et al.* (2008) and has driven the idea that the effect of *Ligula* on reproductive inhibition is mediated through the host's pituitary gland. The effects of the parasite on the pituitary gland were observed 6 weeks post-implantation of a small worm into a mature female roach and also occurred in non-host species such as *Xenopus laevis* (Arme, 1968, 1975). These observations not only indicate that the effects on the pituitary gland may not be dependent on worm burden, but suggest a general endocrine effect rather than an inhibition which is species specific. The mechanism by which this effect is mediated is unknown, but is thought to be via the hypothalamus/pituitary gland/gonadal axis. This axis is a complex array of hormones and feedback mechanisms, any number of which could be affected by *Ligula* (Fig. 2). To understand the possible intricate effects of *Ligula* on fish reproduction thus requires a brief consideration of the complexity that is the reproductive endocrine control mechanism in fish. Gonadotrophin Releasing Hormone (GnRH) is regarded as the first key hormone in the cascade controlling reproduction. GnRH occurs in a variety of forms which are thought to have arisen from gene duplication and mutations (O'Neill *et al.* 1998; Okubu and Aida, 2001) and interacts with receptors present on the pituitary gonadotrophs. The expression of these receptors dictates the sensitivity of the pituitary to

the GnRH, as increased receptor numbers at spawning enhance the preovulatory gonadotrophin surge (Yu *et al.* 1998). During the reproductive cycle, the GnRH content in the hypothalamus and pituitary gland of goldfish increases to a maximum on the day of aggregation for spawning and is lowest in pre-ovulatory and immature fish.

As revealed by the pioneer studies of R.E. Peter and coworkers in the goldfish, pituitary gonadotropic cells of teleosts may be subjected to regulation, not only by a stimulatory control of GnRH as in mammals, but also as direct inhibition by dopamine (for review see Peter *et al.* 1986). GnRH and dopamine are, respectively, the principal stimulatory and inhibitory neurohormones controlling gonadotrophin release (e.g. Hernandez-Rauda *et al.* 1996; Trudeau, 1997). In the goldfish, dopamine was shown to be involved in the inhibition of the final steps of gametogenesis in mature fish (final oocyte maturation and ovulation in females, spermiation in males), a role that was also found in the other cyprinids studied, as well in other, but not all, teleost species (for review see Dufour *et al.* 2005). The possible inhibitory role of dopamine in the inhibition of earlier steps of gametogenesis was investigated in the European eel, in which a strong dopaminergic inhibition was shown to be involved in the pre-pubertal blockade of sexual maturation, before the oceanic reproductive migration (Vidal *et al.* 2004; Dufour *et al.* 2005). A dopaminergic role in the inhibition of puberty was also recently found in the mullet (Nocillado and Elizur, 2008) but not in some other teleost species such as the striped bass or seabream (for review: Dufour *et al.* 2005). This complexity thus gives *Ligula* a range of potential strategies to affect the endocrine status and reproduction of its fish host.

The gonadotrophic hormones, synthesised and released from the anterior pituitary gland have a controlling role on reproduction. GtHI (FSH) mediates gonadal growth (Tyler *et al.* 1999), whilst GtHII (LH) regulates the final stages of maturation and ovulation/spermiation. These gonadotrophins act primarily on the ovary and testes to promote gametogenesis. Receptors for GtH are located in both cell layers surrounding the oocytes (thecal and granulosa cells) and in the testis (Kanamori *et al.* 1987; Yan *et al.* 1992). Fish gonads have the capacity to synthesise steroid hormones such as 17 β -estradiol, testosterone, 11-KT and 17-20 β P (Nagahama, 1999), as well as gonadal peptides (inhibin, activin) which exert feedback control onto the hypothalamus and pituitary gland in a classical feedback loop, and their regulation is dependent on the maturational status of the fish (Feist and Schreck, 1996). *Ligula* could therefore be affecting the secretion of GtH hormones, their receptors and/or the feedback system between the gonads and the brain. Indeed, the fact that fish usually become infected when young may

suggest that the parasite is preventing puberty in its host. *Ligula* could therefore be considered as a useful tool to gain an understanding of the endocrine control of the development of the reproductive system as it could mediate its effect at various levels in the hypothalamus/pituitary gland/gonadal axis (Fig. 2).

Implantation of *Ligula* plerocercoids into uninfected roach results in atresia of developed follicles, or if implanted after ovulation, inhibited recrudescence in fish examined 3 months after implantation (Arme, 1968). The effects are reproducible with a small plerocercoid implanted into a large mature fish, which precludes pressure effects on fish organs or general debilitation from parasite metabolic demands (Arme, 1975). Arme postulated that *Ligula* produces a substance with an inhibitory effect on the pituitary gland, and noted that the nature of the substance was similar in action to testosterone treatment and may possibly be a steroid with anti-gonadal effects on the pituitary gland. Arme *et al.* (1982) later published evidence against this hypothesis as three different assay attempts failed to identify a sex steroid in extracts of worm and in media in which the parasites had been cultured, although cholesterol (sex steroid precursors) was identified in abundance (Arme, 1997). It is plausible that *Ligula* targets GnRH as two forms of gonadotrophin releasing hormone (sGnRH and cGnRH-II) have been localised in brains of ligulosed and uninfected roach by Penlington *et al.* (1997). Although no differences in cell distribution, cell number or staining intensity were detected that could be attributed to *Ligula*, both GnRH forms were present in uninfected and infected fish, although cGnRH positive neurones were more numerous than sGnRH (Williams *et al.* 1998).

Endocrine studies by Carter *et al.* (2005) have revealed a more intricate effect of *Ligula* on the endocrine status of the pituitary gland in roach. The effect of parasitisation on the LH content of this gland was studied using heterologous radio-immunoassay for the LH β subunit of *Cyprinus carpio* which revealed that the pituitary gland of infected roach contained 50% less LH than non-ligulosed fish. In addition, partial cloning of roach LH β subunit allowed Carter and co-workers to show that there was a 50% reduction in LH β mRNA in the pituitary gland of ligulosed roach. These results support the hypothesis that the pituitary gland plays a significant role in the interaction between the parasite and the reproductive status of the fish host. There is evidence however that the interaction between *Ligula* and the brain of fish may be multifaceted. Testosterone can have a two-fold effect on the regulation of LH transcript levels either in its own right (Huang *et al.* 1997) or after aromatisation in the pituitary to E₂ in order to stimulate LH (Antonopoulou *et al.* 1999). Aromatase activity,

which is seasonally present in the pituitary of the goldfish (Melamed *et al.* 1998), as in all vertebrates, is a member of an enzyme complex including P450 aromatase and reductase which carries out the conversion of androgens to estrogens (Gonzalez and Piferrer, 1999; Carreau *et al.* 2002). This essential enzyme complex for estrogen biosynthesis has been demonstrated in the brain and ovary of teleosts (Cruz and Canario, 1999). Stimulatory effects of small quantities of testosterone are consistent with a positive feedback mechanism, which stimulates accumulation and secretion of LH. Large testosterone doses presumably exert negative effects via inhibition of LH (Berglund *et al.* 1995). Recent studies by Hecker *et al.* (2005, 2007) have further investigated the role of aromatase activity in the control of endocrine status in *Abramis brama* from the river Elbe in Germany. Brain aromatase activity was significantly positively correlated with plasma estradiol in females and 11-ketotestosterone in males which the authors suggested led to the disruption of reproductive parameters such as the maturation of germ cells and secondary sex characteristics. The interesting observation was that the prevalence of *Ligula* in the fish was correlated with a suppression of the aromatase activity.

Further studies are also required to investigate the possible impact of the parasite on the expression and production of roach FSH. Indeed, according to data in other teleosts, FSH is supposed to be mostly involved in the control of the first steps of gametogenesis (vitellogenesis in the female, spermatogenesis in the male) while LH would control the last ones (oocyte final maturation and ovulation in the female, spermiation in the male). Recently, Trubiroha *et al.* (2009) developed quantitative real time PCR to evaluate the impact of *Ligula* on the expression of LH β and FSH β subunits in infected roach. In agreement with the authors' previous studies (Carter *et al.* 2005), field studies revealed a significant decrease in both LH β and FSH β pituitary expression in infected roach as compared to non-infected ones. However, under controlled laboratory condition of infection, only FSH β mRNA levels were lowered. This suggested that FSH may be a prime target of *Ligula* inhibitory effect on roach sexual maturation, in agreement with the early role of FSH in the induction of gametogenesis as discussed above. Considering the early blockade of sexual maturation by *Ligula* in the roach, development of new tools such as the use of reproductive cell lines to investigate the regulation of FSH expression and release would be highly relevant.

The possibility that *Ligula* is producing a substance that directly affects pituitary gland activity has been investigated in unpublished studies carried out by the authors using a pituitary primary cell culture system (PPCC) according to the method previously developed (Montero *et al.* 1996; Huang *et al.* 1997).

Pituitary glands cells obtained from female silver (prepubertal) eels (*Anguilla anguilla*) were exposed to secretions (WCM) from *Ligula intestinalis* and LH content measured by radioimmunoassay (Dufour *et al.* 1983). *Ligula* WCM added to eel pituitary cells induced a significant increase in cellular LH content which occurred in a dose-dependent manner (0.658 μg –1.8 μg of *Ligula* protein). Whilst the presence of protease inhibitors did not affect the amount of LH produced when pituitary cells were exposed to whole parasite WCM, it did increase the LH content of the pituitary cells exposed to parasite fractions (e.g. 30+kDa fraction with protease inhibitors = 412.8 \pm 29 ng LH/ μl cell extract, without protease inhibitors = 20.25 \pm 1 ng LH/ μl cell extract, $T = 13.6$, $P < 0.001$), which suggests that this endocrine active substance is susceptible to proteolytic digestion. Indeed, there are previous reports which indicate that *Ligula* plerocercoids not only produce proteolytic enzymes, but also protease inhibitors (Matskasi and Juhasz, 1977; Juhasz, 1979; Matskasi and Nemeth, 1979). The protective effect of protease inhibitors was also observed on the effects of other fractions of parasite secretions (e.g. 10–30 kDa, 3–10 kDa or < 3 kDa) on the LH content of pituitary cells, which suggests that multiple factors secreted by *Ligula* may contribute to the endocrine disruption of the host reproduction.

Although the eel pituitary cell culture model adopted indicated a possible direct effect of *Ligula* on pituitary gland cells, a method of maintaining pituitary cells from the natural host of *Ligula*, the roach, failed to give any conclusive results. The increase in LH content measured on addition of parasite products to the eel pituitary cell model may be due to several possible mechanisms, such as an increase in production of LH and a decrease in the release of LH from the pituitary gland, or no change in LH production, but a decrease in LH release. The effect of *Ligula* on eel pituitary cells also supports the previous suggestion by Arme (1968) that effects of this parasite on host reproduction may not be limited to its natural host. Further studies should aim at investigating the possible direct effects of *Ligula* products on the expression of roach LH β and FSH β subunits.

Whilst there appears to be much debate on the mechanism by which *Ligula* affects the reproductive status of its fish host there appears to be a consensus, at least in the speculation, as to why this should occur and how the parasite might benefit from this effect. Besides the possible effects on the energy balance in the infected fish several authors e.g. Barber and Huntingford (1996), Loot *et al.* (2001a, 2002a, b), Morgan (2003), Dejen *et al.* (2006) have proposed that infection in fish leads to altered behaviours such as movement to shallow waters, occurrence of fish on the surface of the water, swimming impediments, absence of shoaling, and

delayed response to a stimulus, all of which lead to the infected fish being more prone to predation by the avian definitive host or another fish (Museth, 2001). This host/parasite interaction thus has an important role to play in the studies on the parasite manipulation and predator foraging behaviour (Brown *et al.* 2001, 2002; Fenton and Rands, 2006).

Parasitism, and in particular with *Ligula*, is not the only biological factor which can affect the endocrine system and over the last 25 years there have been many studies (e.g. McMasters *et al.* 1996; Harries *et al.* 1997, 1999) which have shown that chemical pollutants can affect the endocrine status of fish. Surprisingly, there are very few studies that have considered water quality in association with the *Ligula*/fish interaction. Recent investigations by Hecker and co-workers (Hecker and Karbe, 2005; Hecker *et al.* 2007) on *Abramis brama* infected with *L. intestinalis* collected from the river Elbe have attempted to relate endocrine status of the fish to parasitisation and the presence of a range of chemicals from industrial, agricultural and domestic sources. Regional differences in infection of the fish were noted along the length of the Elbe studied with highest prevalence occurring in heavily polluted areas. When the authors used a linear model to compare these regional differences in prevalence of infection with biological parameters, not all the differences observed could be attributed to *Ligula*. They proposed that pollution may have contributed to the observed altered reproductive and endocrine status observed in the fish. Whilst these studies in no way refute the observed effects on *Ligula* on the hypothalamus/pituitary gland/gonadal axis they do highlight the possible importance of other biological parameters such as water quality.

LIGULA: A MODEL FOR POLLUTION MONITORING

The association of water quality with the interaction between parasites and their fish hosts is not surprising as there have been recent studies which have highlighted the role of pollutants in host-parasite interactions (Hoole, 1997; Morley *et al.* 2006, Sures, 2008a, b). Several studies have investigated the effects, not only of pollutants on the immune response of the fish hosts to the parasitic fauna (Hoole, 1997), but also on helminth life cycles. These latter studies have primarily concentrated on trematode stages such as eggs, miracidia, cercariae and metacercariae (Abd Allah *et al.* 1997; Morley *et al.* 2001a, b, c; 2002, 2003; Pietrock *et al.* 2002) and the bioaccumulation of heavy metals in adult acanthocephalans and cestodes in fish (Sures *et al.* 1997; Sures and Siddall, 1999; Sures, 2003). The relatively few studies which have been carried out on the interaction between fish tapeworms and heavy metals have indicated that these metazoan parasites possess,

on some occasions, the ability to bio-accumulate pollutants at greater levels than their fish hosts. For example, the monozoic cestode, *Monobothrium wagneri*, had higher concentrations of both cadmium (Cd) and lead (Pb) than its host tench, *Tinca tinca*, whilst there was no difference between Pb burdens detected in the adult cestode, *Bothriocephalus scorpii*, and those present in the intestinal wall of its host turbot, *Scophthalmus maximus* (Sures *et al.* 1997). In addition, the concentration of heavy metals is dependent on the different body parts of tapeworms analysed. For example, posterior gravid proglottids of the cestode, *Bothriocephalus scorpii* accumulate higher contents of Pb and Cd than the anterior immature ones (Sures *et al.* 1997). This may be due to the ability of tapeworm egg shells to bio-accumulate heavy metals (Khalil *et al.* 2009).

The size and ubiquitous nature of *Ligula* has led to the large plerocercoid stage being considered as a possible model in the study of heavy metal contamination in water bodies. Tenora *et al.* (2000) utilised atomic absorption spectrometry to monitor lead, chromium and cadmium levels in *L. intestinalis* and *Philometra ovata* in the body cavity of three cyprinid fish species (*A. brama*, *R. rutilus*, *B. bjoerkna*). All heavy metals studied bio-accumulated within the plerocercoid to a greater level than in the fish muscle. Further studies carried out by Tenora *et al.* (2002) also revealed that this ability to bio-accumulate heavy metals (Pb, Cd) was not restricted to the plerocercoid, as the adult *L. intestinalis* contained higher levels of these two metals than the definitive avian hosts. It would appear however, that the accumulation of heavy metals by the plerocercoid stage may be dependent on the age of the parasite. Barus *et al.* (2001) noted that whilst nickel levels were higher in young plerocercoids, the levels of the majority of heavy metals monitored, i.e. Pb, Cr and Cd, were greater in the older parasites. They hypothesised that this indicated that accumulation of heavy metals in the parasite was a gradual and long-term process which probably occurred during the growing phase of the parasite in the fish host. More recent studies by Tekin-Ozan and co-workers (Tekin-Ozan and Kir, 2005, 2008; Tekin-Ozan and Barlas, 2008), utilising *Ligula* and tench (*Tinca tinca*), have extended the greater bio-accumulation by the plerocercoid to a more extensive range of heavy metals and speculate that if the parasite reflects the amount of heavy metal contaminants in the water and sediments, it may provide reliable data about pollution in water bodies. Whilst this is perhaps debatable, the value of the association of the plerocercoid with heavy metals probably lies in the size of the parasite which may assist in the elucidation of how parasites can bio-accumulate pollutants with no apparent ill-effects on their biology.

LIGULA: A MODEL OF GENETIC BIODIVERSITY

One important area that underpins all of the studies mentioned above relates to the taxonomic status of the parasites used, which are from a range of fish species and geographical locations. The classification of ligulid tapeworms into the genus *Ligula* Bloch, 1782 and genus *Digamma* Cholodkovsky, 1914 has proved controversial for many years, and yet *L. intestinalis* with its global distribution and complex life cycle may be a valuable model to study speciation and the evolution of parasite genetic diversity. Since the turn of the century there have been several investigations which have applied molecular techniques to studies on speciation within ligulid cestodes. Li and Liao (2003), using sequences for the 5' end of 4 genes, i.e. 28S ribosomal ribonucleic acid (28rRNA), mitochondrial cytochrome c oxidase subunit 1 (CO1), nicotinamide adenine dinucleotide dehydrogenase subunit 1 (ND1) and the internal transcribed spacer of the nuclear ribosomal deoxyribonucleic acid (ITS1), noted that there was low genetic divergence in these in *Ligula* and *Digamma* and suggested that the two parasites should be considered as different species within the genus *Ligula*. These authors had previously utilised 28rRNA and CO1 isolated from formalin-fixed specimens of *Ligula* obtained from Qinghai-Tibet Plateau, Russia and England and proposed that the Chinese *Ligula* was within the same species as that occurring in Europe. Importantly, they proposed that geographic location, host affinity and host habitat were not reliable criteria to use in the classification of *Ligula* (Li *et al.* 2000). Olson *et al.* (2002) provided molecular evidence that *Ligula* occurring in gudgeon, and roach in Lough Neagh, Northern Ireland, were separate strains which may be reflected in their effect on host gonadal development. In addition, *Ligula* from minnow (*Phoxinus phoxinus*) from Wales resembled those from the Lough Neagh roach. The authors suggested that the existence of separate strains in Lough Neagh probably resulted from the introduction of roach and an increase in the number of the definitive host, the great crested grebe. Logan *et al.* (2004) highlighted that *L. intestinalis* from Turkey were genetically isolated from their European and Chinese counterparts. In a recent extensive study carried out by Bouzid *et al.* (2008a), the genetic variation within *L. intestinalis* obtained from 13 fish species originating from different localities i.e. Algeria, Australia, Canada, China, Czech Republic, Estonia, Ethiopia, France, Germany, UK, Mexico, Poland, Russia, Tunisia and the Ukraine, was analysed using two mitochondrial genes, cytochrome oxidase 1 and cytochrome B, and the nuclear sequence of intergenic transcribed 2 (ITS2). The authors proposed that the evolutionary patterns observed were determined at the local and global levels. At the local aspect, the migrating avian

definitive host was thought to be important in preventing the establishment of genetic barriers, whilst on the global scale, genetically distinct clusters were observed. In addition, the authors noted that *Ligula* was split into two clades, termed A and B. Clade A contained samples from Tunisia and Europe and were obtained from 'derived cyprinids' (*Abramis*, *Alburnus*, *Phoxinus*, *Rutilus* and *Scardinius*), whilst Clade B was restricted to European, Algerian, Chinese and Australasian samples from 'basal cyprinid fish' (*Barbus*, *Gobio* and *Rhodeus*). Bouzid *et al.* (2008b), utilising *L. intestinalis*, also proposed that inter-simple sequence repeat markers was a rapid and inexpensive technique to define markers that could be used to assess genetic diversity. Recently, isolation and characterization of microsatellite loci have been used to understand the genetic complexity and diversity that occurs within the ligulids. After this technology was established as a useful tool in the case of *L. intestinalis* (Stefka *et al.* 2007), Stefka and co-workers in 2009 used 15 microsatellite loci to monitor the genetic differences in populations of *L. intestinalis* from a range of distant geographical locations in North America, Europe, Asia, Africa and Australasia. They noted a very high level of polymorphism and strong genetic structure in *Ligula* from these localities and proposed some very interesting reasons for this. For example, the existence of parasite subdivisions between Europe and Tunisia was due to the Mediterranean Sea effect which, although did not prevent migration of the avian definitive host, the fact that the adult tapeworm persists in the bird for only one week (Dubinina, 1966) would probably mean that the tapeworm was not transported between the two localities. In addition, they also suggested that the fish immune response may be a factor in determining host-specificity of the *Ligula* genotypic lineages. Unfortunately, most of the studies on the immune response to the plerocercoid stage have been carried out in roach. In a series of publications by Hoole and co-workers it was established that there is an intense cellular response to the parasite (Hoole and Arme, 1982, 1983a,b; Taylor and Hoole, 1989a,b, 1993, 1994, 1995) which involved several specific and non-specific humoral components (Hoole and Arme, 1986, 1988; Williams and Hoole, 1992, 1995). However, even with the host response, substantial differences to *Ligula* infection occur in roach and gudgeon (Arme, 1997), again indicating possible genetic diversity of the parasite.

In conclusion, since the classical experiments carried out by J. D. Smyth in 1947, which revealed *L. intestinalis* as a model to study tapeworm development *in vitro*, this large metazoan parasite has proved an invaluable model not only for parasitologists in general, but endocrinologists, ecologists, geneticists, immunologists and in pollution studies.

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