

Overlap in avian communities produces unimodal richness peaks on Bornean mountains

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Abstract: Altitudinal gradients provide tractable, replicated systems in which to study changes in species richness and community composition over relatively short distances. Previously, richness was often assumed to follow a monotonic decline with altitude, but recent meta-analyses show that more complex patterns, including mid-altitude richness peaks, are also prevalent in birds. In this study, we used point counts to survey birds at multiple altitudes on three mountains on the island of Borneo in Sundaland, an area for which quantitative analyses of avian altitudinal distribution are unavailable. In total we conducted 1088 point counts and collected associated habitat data at 527 locations to estimate species richness by altitude on Mt Mulu (2376 m), Mt Pueh (1550 m) and Mt Topap Oso (1450 m). On Mulu, the only mountain with an intact habitat gradient, bird species richness peaks at 600 m. Richness appeared to peak at 600 m on Totap Oso as well, but on Pueh it peaked several hundred metres higher. The richness peak on Mulu differs from that predicted by null models and is instead caused by the overlap of distinct lowland and montane avifaunas, supporting the faunal overlap hypothesis. This finding provides further evidence that a lack of coincidence between peak turnover and peak richness is not sufficient evidence to rule out faunal overlap as a causal factor.

Key Words: bird, elevational gradient, faunal overlap, mid-domain effect, MDE, Mulu, NMDS, point count, Pueh, turnover

INTRODUCTION

Altitudinal gradients in species richness on tropical mountains provide tractable systems for studying ecological processes of global importance. These gradients are particularly useful because large changes in both climate and biota take place over relatively short distances (Lomolino 2001, Malhi *et al.* 2010). Moreover, altitudinal gradients are replicated many times (Fjeldså & Rahbek 2012), allowing comparisons of faunal and floristic change at both regional and global levels.

Although species richness was thought originally to decline with altitude, extensive review of global mountain data has shown the actual situation to be much more complex (Rahbek 1995). Several patterns occur regularly among mountains (Figure 1). In vertebrates, global patterns differ among taxa (McCain 2005, 2007, 2009, 2010), but the most common is a hump-shaped

distribution indicating peak richness at low–middle altitudes (Rahbek 1995). Numerous explanations have been proposed for these patterns, including null models, such as the mid-domain effect (MDE) (Colwell *et al.* 2005) and variability in abiotic and biotic factors (Lomolino 2001).

In addition to richness, another important altitudinal-gradient measure is species turnover. Hump-shaped richness peaks are often assumed to correspond to areas of community overlap or turnover (Lomolino 2001). However, McCain & Beck (2015) found that richness and turnover peaks seldom coincide in vertebrate communities. However, a hypothesis based on faunal overlap predicts that on a mountain with distinct lowland and montane communities, altitudinal range midpoints will be bimodally distributed, reflecting the existence of each group, and that richness of the lowland group will decline monotonically with altitude (Beck & Chey 2008). Where the two altitudinal groups overlap, a hump of species richness will occur.

In view of these issues, we set out to survey avian communities using consistent quantitative methods along

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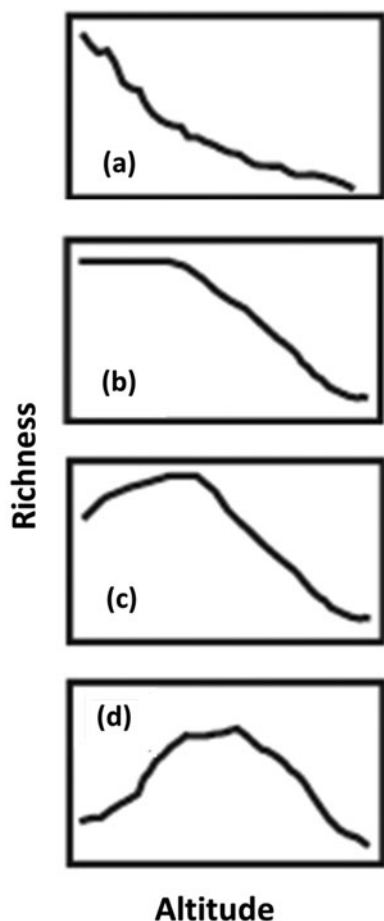


Figure 1. Common patterns of the avian richness-altitude relationship, modified from McCain (2009). These include: monotonic decline in richness (a), low-altitude plateau followed by a monotonic decline (b), hump-shaped decline with a low-mid-altitude peak (c) and symmetrical mid-altitude peak (d).

primary-forest gradients on three mountains on Borneo. Borneo is perhaps the most important centre of rain-forest species diversification in insular South-East Asia (de Bruyn *et al.* 2014, Sheldon *et al.* 2015), and its forests are also changing rapidly due to immense pressure from logging and plantation development (Wilcove *et al.* 2013). Despite Borneo's regional importance and the threats facing its habitats, quantitative surveys of primary-forest birds on the island are relatively few; almost all are focused on lowland forest disturbance at Danum Valley, Sabah (Edwards *et al.* 2011, Lambert 1992), and none examines montane gradients. Indeed, of 78 total avian altitudinal-gradient data sets used in the meta-analysis of McCain (2009), only nine are from South-East Asia, including one each from Sumatra, Java and Borneo. However, the last three derive not from surveys but only from altitudes gleaned from a field guide (MacKinnon & Phillipps 1993). No publications based on quantitative surveys of altitudinal variation in bird species occurrence

exist from any of the Greater Sunda Islands, although surveys have been conducted on Mt Kinabalu (Harris *et al.* 2012).

Here we use avian point counts on three Bornean mountains to quantify patterns of avian species richness, turnover and community composition along altitudinal gradients. With these data we test several hypotheses: (1) bird species richness peaks at an intermediate altitude consistent among mountains; (2) the richness and turnover patterns fit predictions of the MDE; and (3) species composition is similar among the mountains for a given altitude.

METHODS

Study sites

The island of Borneo consists primarily of coastal lowlands surrounding an interior mountain chain that runs from the north-east to the south-west (Figure 2). This chain comprises mountains mainly below 2000 m, with only a few reaching 2400 m and one, Mt Kinabalu, rising to 4096 m. There are also a few isolated mountain ranges and volcanoes of low stature. The central mountain chain, with its larger, more connected mountains hosts a larger complement of montane bird species (Banks 1952). For surveys, we selected three mountains that differ in size, isolation from the main mountain chain, and distance from the coast to compare patterns of species richness and altitudinal distribution (Figure 2).

Mt Topap Oso (0.929°N, 114.206°E) in East Kalimantan, Indonesian Borneo, is a remote 1450-m peak in the central mountain chain in an area where most peaks range from 1200–1450 m. Forest on its lower slopes (below 600 m asl) has been disturbed by shifting agriculture, but above this altitude primary forest is intact. We reached the mountain from the villages of Naha Silat and Long Apari in the headwaters of the Mahakam River and conducted point counts at 600, 800, 1000 and 1200 m asl from June to November 2012 on the western slopes of the mountain, as well as on the southern slopes of a sister peak (referred to as Mt Baring Uning on some maps) connected by a long ridge (0.858°N, 114.150°E). These mountains have never been surveyed for birds and are representative of remaining primary montane forest in Borneo.

Mt Pueh (1.721°N, 109.669°E) is a 1550-m mountain in far western Sarawak, Malaysian Borneo. It sits only a few kilometres from the coast, and is separated from the island's central mountain chain by about 300 km of lowlands and isolated smaller peaks. Several montane bird species that are present on mountains of similar size connected to the central mountain chain are absent from Pueh, probably due to its isolation (Banks 1952,

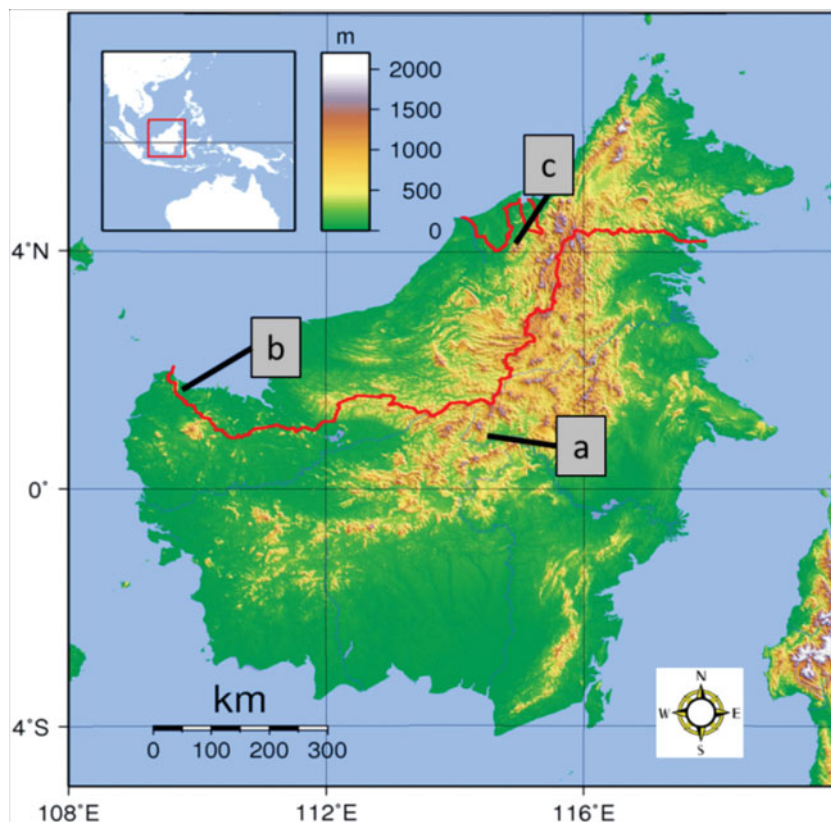


Figure 2. Location of study sites on Borneo. Mt Topap Oso in East Kalimantan, Indonesia (a), Mt Pueh in western Sarawak, Malaysia (b), and Mt Mulu in eastern Sarawak (c).

Chua *et al.* 2017, Manthey *et al.* 2017). Pueh has stunted montane forest at its summit, and thus displays the telescoped vegetational zones sometimes observed on coastal mountains (Bruijnzeel *et al.* 1993). It is also home to one of Borneo's montane endemics, the mountain black-eye (*Chlorocharis emiliae*), a sky-island species that was rediscovered on Pueh only recently (Ramji *et al.* 2012). Its presence on Pueh is probably related to the small patch of ericaceous scrub at Pueh's summit. Parts of our study area on Pueh were selectively logged in the past with tractors up to 900 m asl and by helicopter at higher altitude. We conducted point counts at 600, 800, 1000 and 1200 m asl from June to August 2013. As on Mt Topap Oso, we limited our surveys to altitudes at and above 600 m asl to avoid shifting cultivation plots that have replaced forests.

Mt Mulu (4.045°N, 114.929°E) is Borneo's fifth highest mountain at 2376 m, located in Sarawak near its border with Brunei and the Malaysian state of Sabah. Almost all of Borneo's montane bird species inhabit Mulu, including many endemics (Burner *et al.* 2016). The mountain is the central feature of Mt Mulu National Park and as such is covered by primary forest from near sea level to the summit. Only in the floodplain at the mountain's base (~50 m) has the forest been selectively

logged. Mulu is probably the only site in Borneo outside of Brunei where a complete primary-forest gradient can be found. We accessed Mulu via the summit trail, which is maintained by the Park for tourists, and conducted avian point counts at 50, 300, 600, 900, 1200, 1500 and 1800 m asl from June to September of 2014. Most of our analyses in this paper are focused on Mt Mulu because of its intact forest gradient and the completeness of our survey range.

Survey methods

Avian communities were surveyed at point locations spaced every 150–200 m along transects at each sampled altitude. Counts consisted of 10–12-min (Mt Topap Oso and Mt Pueh) or 6-min (Mt Mulu) audio recordings using a Marantz digital recorder and Sennheiser microphone for later species identification. Count length was shortened on Mulu to allow time for additional replicates at each survey point. A truncated data set that reduced all Topap Oso and Pueh surveys to 6 min produced similar richness estimates, so we retain the full data set in this paper. Using audio recordings allowed more thorough consideration of the many bird sounds in this

Table 1. Habitat and survey parameters measured at each avian point count for inclusion in ordinations. Measurements were taken within plots of three sizes (100-m radius, 20-m radius, 5-m radius).

Parameter
Date and time
Altitude (m)
Weather
Distance to nearest stream (m)
100-m radius plot:
Number of treefall gaps
Average canopy height (m)
Horizontal visibility (m)
20-m radius plots:
Canopy cover %
Count of stems 25–40 cm diameter
Diameter of stems >40 cm diameter
Total basal area
Slope %
5-m-radius plots:
Count of stems <5 cm diameter
Count of stems 5–15 cm diameter
Count of stems 15–25 cm diameter
Shrub height (m) and % cover
Groundcover height (m) and % cover

species-rich environment (Haselmayer & Quinn 2000). Species observed during the count were also noted. Points on Topap Oso were surveyed only once each, but points on Pueh and Mulu were sampled three to four times each, usually within a few days of the first visit (MacKenzie *et al.* 2002). At each altitude on each mountain, 20–70 unique points were surveyed depending on time available and difficulty of access. Unrecognized recorded vocalizations were identified by Andrew Siani, an expert on Malaysian bird songs. At each point, time of day, altitude, weather, latitude and longitude, as well as habitat data were recorded. Counts were conducted from 06h00 to 10h30 solar time, and only when not raining.

Habitat parameters were recorded for each point using the methods of Sheldon *et al.* (2010), and were measured outside the morning survey period (Table 1).

Data analysis

Turnover is a measure of difference in species composition between two altitudes. Nestedness (Baselga 2010) is the extent to which one community is a subset of a larger community (i.e. at an adjacent altitude). To address hypothesis one (patterns of species richness), richness at each location was estimated using the Chao2 estimator in EstimateS, which allows comparison between multiple sites that differ in sampling effort. Turnover was estimated using Simpson's dissimilarity, and turnover and nestedness were calculated using visual basic scripts (available at http://spot.colorado.edu/~mccainc/simulation_programs.htm) from McCain & Beck (2015).

To address hypothesis two (fit to MDE predictions), empirical results from Mt Mulu were compared to the three null models of McCain & Beck (2015): (1) the hard boundaries mid-domain effect (MDE), in which species altitudinal ranges are constrained to lie entirely within the gradient sampled; (2) the partially bounded model, in which ranges are constrained to fit within a gradient that is expanded by 20% on each end; and (3) an unbounded model, in which ranges are placed randomly on an altitudinal gradient twice as large as the sampled gradient. Fit to these models was assessed using R-squared values. The MDE model was not used for Mt Topap Oso and Mt Pueh due to the limited altitudinal sampling range.

To test the faunal overlap hypothesis on Mulu, empirical range midpoints were calculated for each species (max altitude – min altitude) and the results plotted to look for evidence of distinct groups of lowland and montane species (Beck & Chey 2008). Species with five or more detections were then designated as lowland if 75% of observations occurred below the peak turnover point (as calculated above), and montane if 75% of observations occurred above this point. All other species were considered mid-altitude/widespread.

To address hypothesis three (comparisons among the three Bornean mountains), we tested and compared differences in community composition between altitudes and mountains, and tested the correlation between habitat parameters and differences in these communities, using non-metric multidimensional scaling (NMDS) via the metaMDS function in the package Vegan in R. The SIMPER function was used to calculate each species' contribution to dissimilarity between sites. All points from all mountains were ordinated together in a single data set to examine inter- as well as intra-mountain differences. The number of relevant ordination axes was assessed using a measure of stress from the ecodist package (Goslee & Urban 2007) in R.

Multiple Response Permutation Procedure (MRPP) was performed using the function *mrpp* in Vegan to test for differences among altitudinal groups. This procedure tests whether a significant difference occurs between communities at two or more points. It compares both differences of location in ordination space (means) and differences of spread or variation.

RESULTS

Avian surveys

We conducted 1088 point-counts at 527 points over the course of this study, including 238 locations on Topap Oso (one visit per point), 114 on Pueh (\bar{x} = 2.52 visits per point) and 175 on Mulu (\bar{x} = 3.22 visits per

point) based on difficulty of access and time available. Points were divided approximately equally among the four altitudes on Topap Oso and Pueh, and among the seven altitudes on Mulu. From these counts, 11,152 species presence records were obtained, representing 213 species (Appendix 1).

We detected a total of 187 species on Mulu, followed by 155 species on Topap Oso, and 151 on Pueh. Of the 213 total species, 115 were found on all three mountains; 7 occurred on Pueh and Topap Oso only, 16 Pueh and Mulu only, and 27 Mulu and Topap Oso only. Mulu had the largest number of unique species, 29, followed by Pueh with 13 and Topap Oso with 6. However, only 15 of the 29 unique species on Mulu were detected between 600 m and 1200 m, corresponding to the survey range on the other two mountains. Of 52 distinctly montane or submontane species detected, all of which were found on Mulu, 33 were found on Topap Oso but only 24 on Pueh.

Richness, turnover and nestedness

Species richness on Mulu increased with altitude until 600 m, where it peaked and thereafter declined to less than a third of the peak-value at 1800 m (Figure 3). Species turnover on Mulu showed a single peak (0.40) between 900 m and 1200 m (Figure 4), hundreds of metres above the richness peak at 600 m. Nestedness (Figure 4) was highest between 600 m and 900 m (0.12). A lower nestedness value between 900 m and 1200 m (0.06) was consistent with the higher turnover between these altitudes, and helped explain why estimated richness can decline so rapidly from 600 m (138 species) to 900 m (91 species) without a correspondingly high turnover rate; the 900-m community was to some extent just a subset of the community at 600 m. Turnover was high between 900 m and 1200 m, corresponding to the low nestedness value.

Species ranges on Mulu did not occur along the gradient at random. Instead, range midpoints were bimodally distributed according to whether the species were members of the lowland or montane community (Figure 5). Of 132 species on Mulu with five or more detections, the majority (92%) belong to one of the two groups, as defined by having >75% of their detections either below (lowland) or above (montane) the altitude of maximum turnover between 900 and 1200 m (Appendix 1). Those not fitting either group (8%) were considered mid-altitude species. The richness patterns of these three groups combined to form the low-to-mid-altitude hump in species richness (Figure 6).

Richness patterns on the parts of the gradient that were sampled were less obvious on Pueh and Topap Oso (Figure 3). Richness on Topap Oso may also peak at 600 m. Pueh appeared to have a mid-altitude richness peak

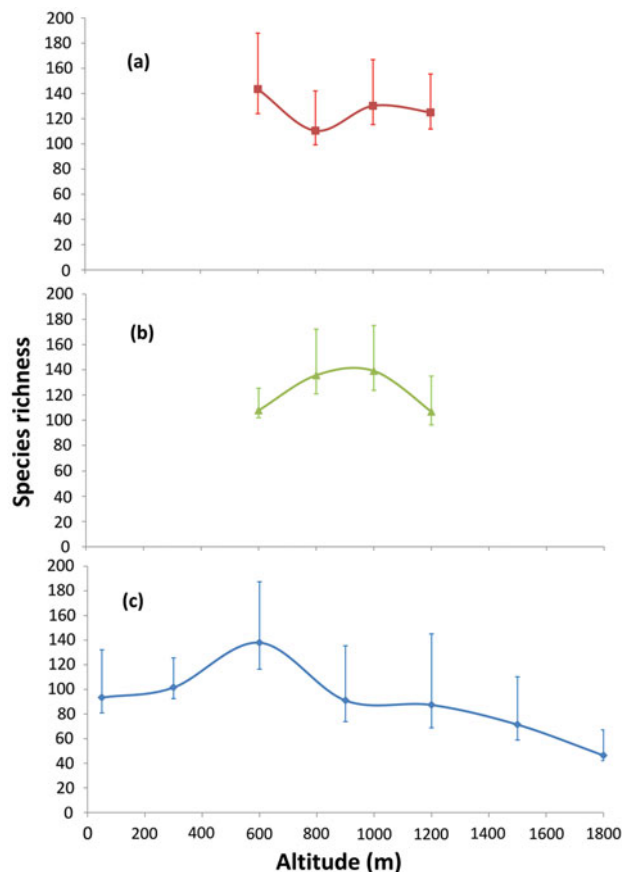


Figure 3. Avian species richness by altitude on three Bornean mountains. Richness based on Chao2 estimator from EstimateS, with 95% confidence intervals: Mt Topap Oso (a), Mt Pueh (b) and Mt Mulu (c).

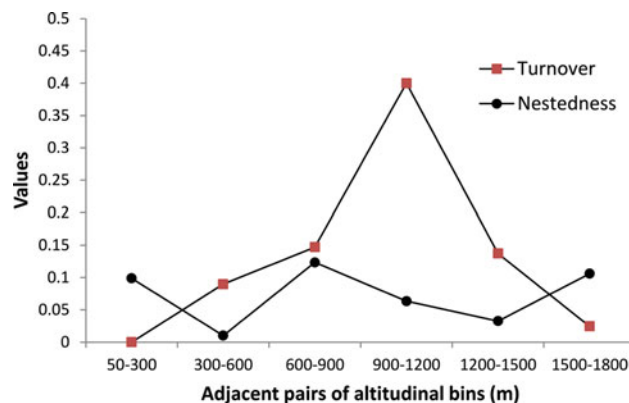


Figure 4. Avian species turnover and nestedness between adjacent pairs of altitudinal bins on Mt Mulu in Sarawak, Malaysian Borneo. Turnover measures differences in species composition between altitudes, while nestedness indicates the extent to which the species at one altitude are a subset of those occurring at an adjacent altitude. The point of maximum turnover is several hundred metres above the point of maximum richness (600 m).

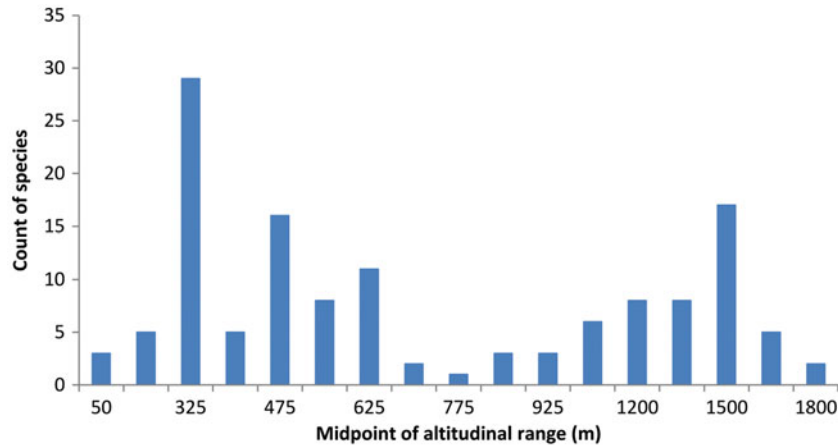


Figure 5. Frequency of midpoints of species' altitudinal ranges on Mt Mulu in Sarawak, Malaysian Borneo. The bimodal distribution provides support for the existence of distinct lowland and montane groups of avifauna.

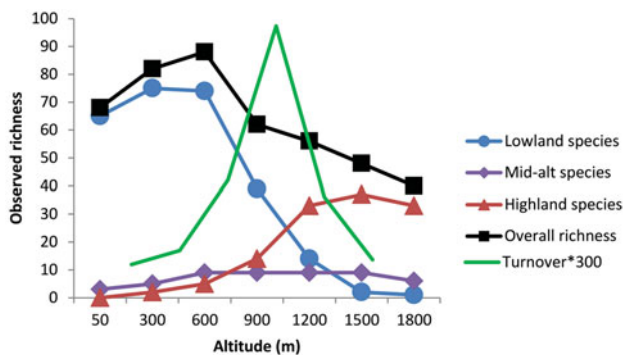


Figure 6. Empirical avian species richness and turnover by altitude on Mt Mulu. Richness of lowland, mid-altitude, and montane communities combine to produce an overall richness pattern with a low-mid-altitude richness peak similar to the estimated total richness curve. Turnover (which is scaled up by a factor of 300 for plotting) peaks in the interval between 900 and 1200 m, the interval in which lowland species richness steeply declines while montane species richness increases.

at a higher altitude than Mulu and Topap Oso, i.e. 800 m to 1000 m, and would certainly fit either the low-mid or mid-altitude peak pattern. Species turnover on Topap Oso and Pueh was on average much lower than on Mulu (peaking at 0.11 and 0.14, respectively, compared with Mulu's maximum of 0.40). Nestedness was highest on Topap Oso and Pueh between 800 m and 1000 m (0.10 and 0.12, respectively).

MDE and other models

For patterns of richness, turnover and nestedness on Mulu fit to expectations of the three null models was generally low. Correlation with predictions of the MDE, soft boundaries, and unbounded models was especially low for richness ($R^2 = 0.11, 0.13$ and 0.21 , respectively)

and nestedness ($R^2 = 0.02, 0.25$ and 0.04 , respectively). Observed turnover fitted the null expectations of the MDE and soft-boundary models somewhat better ($R^2 = 0.35$ and 0.47 , respectively), but did not fit the unbounded model ($R^2 < 0.01$). Over half of the empirical turnover and nestedness values fell outside the 95% confidence intervals of each null model. The incompleteness of the altitudinal range sampled on Topap Oso and Pueh made the null model simulations less informative because data were available from only three altitudinal intervals in the middle of the mountains.

Community composition

The three mountains we sampled shared many species and several of these were common on all three mountains, including the golden-whiskered barbet (*Megalaima chrysopogon*), Bornean barbet (*Megalaima eximia*), blue-eared barbet (*Megalaima duvaucelii*), grey-headed canary-flycatcher (*Culicicapa ceylonensis*), brown fulvetta (*Alcippe brunneicauda*) and chestnut-backed scimitar babbler (*Pomatorhinus montanus*). Additionally, each mountain had a few common species of its own that were not nearly so common on the other mountains. These included the yellow-bellied warbler (*Abrosopus superciliosus*), grey-throated babbler (*Stachyris nigriceps*) and wreathed hornbill (*Rhyticeros undulatus*) on Pueh, and the chestnut-rumped babbler (*Stachyris maculata*), Asian fairy-bluebird (*Irena puella*) and rufous-crowned babbler (*Malacopteron magnum*) on Topap Oso. Most of the examples from Mulu were upper montane species that rarely occur at altitudes sampled on the other mountains (e.g. Blyth's shrike-babbler, *Pteruthius aeralatus*, and chestnut-capped laughingthrush, *Garrulax mitratus*), but also included the low-mid-altitude fluffy-backed tit-babbler (*Macronus ptilosus*).

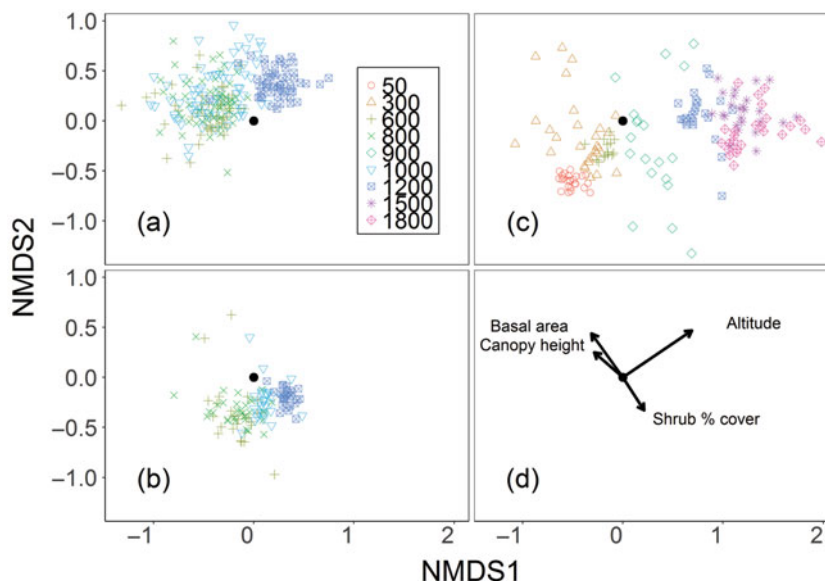


Figure 7. NMDS of combined avian point count data. Each mountain is displayed on a separate graph for clarity: Mt Topap Oso (a), Mt Pueh (b), and Mt Mulu (c). Positions of point clusters from different mountains relative to each other can be compared in reference to the solid dot at the centre. Environmental vectors (d) show the strength and direction of correlations between the labelled habitat parameters and the bird community composition of points. All vectors are significant ($P < 0.001$) based on a perMANOVA in R.

Ordination of combined data from the three mountains via NMDS, and a series of pairwise MRPP tests, showed that all groups on all mountains differed significantly from one another (Figure 7; max $P < 0.01$). Differences in ordination space between altitudes were greatest on Mulu, with clusters moving left to right across the plot with increasing altitude. On Topap Oso and Pueh, altitudinal clusters of points were less visually distinct, although altitudinal groups were still significantly different (mrpp, max $P < 0.01$). Points at 1200 m on both smaller mountains overlap little in ordination space with points from other altitudes, but there were broad zones of overlap between points at 600 m, 800 m and 1000 m. The two smaller mountains occupy parts of the graph distinct from one another, with points from Mulu spread more widely across the plot (reflecting its greater altitudinal range).

The SIMPER function in R was used to calculate individual species' contributions to Bray–Curtis distance between altitudinal groups among and within mountains (Appendix 2). At least 33% of the variation between sites of similar altitude was explained by differences in only 10–15 species. The majority of these most significant species were important across multiple pairwise comparisons among multiple mountains and altitudes. They include the chestnut-backed scimitar babbler, golden-whiskered barbet, brown fulvetta, grey-headed canary-flycatcher and Bornean barbet. Each of these species was detected on all mountains at all altitudes from 600 to 1200 m. Only a few of the most influential species overall were entirely absent from any altitude

within this range on any mountain – the blue-eared barbet was not detected above 600 m on Mulu, while the chestnut-winged babbler (*Stachyris erythroptera*) and red-throated barbet (*Megalaima mystacophanos*) were not detected above 900 m on Mulu, and the spectacled bulbul (*Pycnonotus erythrophthalmos*) was not detected above 1000 m on Pueh.

In contrast to these species that showed up repeatedly in the SIMPER analyses, a few species contributed to differentiating only a single pair of sites. Grey-throated babbler was common at 1200 m on Pueh, but only a few individuals were detected at this altitude on Topap Oso. Bornean bulbul (*Pycnonotus montis*) was common at 1200 m on Topap Oso, but only a few were detected at this altitude on Mulu. Pale blue flycatcher (*Cyornis unicolor*) and short-tailed babbler (*Malacocincla malaccensis*) were common on Pueh at 600 m (and higher), while only a few were detected on Topap Oso at this altitude.

Habitat

Environmental vectors (Figure 7d) highlight habitat features most strongly correlated in the ordination. Altitude, woody plant basal area, canopy height and per cent shrub-cover were all significantly correlated with community composition ($P < 0.001$); this was especially true of altitude ($R^2 = 0.69$). That this relationship held on all three mountains was apparent because points moved left to right on the ordination plots with increasing altitude, although points below 1200 m on the smaller mountains

changed little with altitude. The effects of plant basal area, shrub per cent cover and canopy height were also correlated significantly with species composition ($R^2 = 0.29, 0.15, \text{ and } 0.15$, respectively; all P values < 0.001).

Canopy cover, shrub cover and total woody plant basal area differed between mountains and were significant vectors in the ordination, but none of them was significantly correlated with species richness. On Mulu, average canopy height increased with altitude from an average of 26.0 m at sea level (possibly due to some large tree removal) to a maximum of 36.1 m at 900 m, then declined steadily with altitude above this point to an average of only 16.4 m at 1800 m ($R^2 = 0.45, P < 0.001$). Canopy height was not correlated significantly with altitude within the narrower sampled altitudinal range on Topap Oso, where average height was 33.3 m. On Pueh points at 600 m had an average canopy height of 32 m. Canopy height at points from 800 m to 1200 m averaged 28.1 m, significantly lower than points from 600 m ($P < 0.01$) but not different from each other.

Canopy cover, assessed using a canopy cover index, was not correlated with altitude but did vary significantly between sites (paired t -tests with Bonferroni correction, maximum $P < 0.001$). Cover on Mulu averaged 20% higher than Pueh, and 40% higher than Topap Oso. Total woody plant basal area was correlated with altitude within the sampled range only on Mulu ($R^2 = 0.17, P < 0.001$).

DISCUSSION

This study of Bornean birds provides Sundaland's first example of an altitudinal study of montane bird species richness and turnover derived from a single set of replicated, quantitative surveys (but see Harris *et al.* 2012 for a study of occurrence). Of the three mountains surveyed, Mt Mulu has the most complete primary forest altitudinal gradient. Richness on Mulu is not correlated with patterns of community nestedness and turnover, or with values predicted by the MDE. Richness peaks at 600 m, and this peak appears simply to be the result of overlap of distinct lowland and montane bird communities, a result consistent with the faunal overlap hypothesis (Beck & Chey 2008). While low-to-mid-altitude richness peaks have previously been attributed to lowland-montane community overlap (Herzog *et al.* 2005, Romdal & Rahbek 2009), more recent studies have downplayed the importance of this phenomenon (McCain & Beck 2015) or found it to lack explanatory power (Beck *et al.* 2017). On the other two Bornean mountains, richness again is not correlated with community nestedness or turnover. MDE could not be determined for those mountains because of the short altitudinal range surveyed. Mt Pueh shows an apparent mid-altitude peak in richness,

whereas richness on Mt Topap Oso appears to peak at 600 m (or its richness-peak is outside the survey range). On all three mountains, lowland and montane communities are predictable assemblages that are distinct from one another, and they overlap to varying degrees.

Richness, turnover, nestedness

The species richness gradient on Mt Mulu displays a low-altitude plateau, then it rises to a hump-shaped peak at 600 m, followed by a monotonic decline (Figure 3). This pattern agrees with about 25% of avian altitudinal gradient patterns worldwide (McCain 2009). The same pattern has been shown to be common in moths (Beck *et al.* 2017), plants (Grytnes *et al.* 2008) and mammals (McCain 2005), including at several sites in South-East Asia. This finding reinforces the general observation that richness does not always decline monotonically with altitude, a phenomenon inconsistent with the global latitudinal diversity gradient (Rohde 1992). The difference between altitudinal and latitudinal patterns suggests that the forces influencing richness on tropical mountains differ at least partially from those acting globally. Such differences are not surprising given the massive difference in scale. The limited spatial extent of altitudinal gradients, for example, likely leads to overlap between distinct montane communities based on spill-over in marginal habitat. Such differences are worth keeping in mind as biogeographers try to relate species richness patterns on mountains to those across the globe (Rahbek 2005).

When combined with richness estimates, turnover and nestedness values can help identify the community characteristics producing an altitudinal richness pattern. As in most vertebrate datasets worldwide (McCain & Beck 2015), the richness and turnover peaks on Mulu do not coincide. McCain & Beck (2015) acknowledge that this does not exclude the possibility that mid-altitude richness peaks are the result of a broad zone of overlap between distinct highland and lowland communities, but report that in at least half of the datasets they examine there is no evidence for such distinct communities. In Borneo, however, distinct lowland and montane bird communities are quite clear, and Mulu provides an example illustrating that the richness peak and turnover peak are products of different (but related) phenomena. The low-mid-altitude richness peak at 600 m occurs because montane species richness increases more rapidly than lowland species richness declines with altitude (Figure 6). The pronounced, single turnover peak on Mulu occurs between 900–1200m because most lowland species are exhausted above 900m, whereas virtually all species above that altitude are montane (Figure 6). Richness at 900 m resembles richness at 1200 m, but

this numerical similarity obscures a significant change in species composition. This change is apparent not only in the ordination, but also in the contrast between high turnover and low nestedness values across this interval. For these reasons, a lack of coincidence between turnover and richness values for gradients with significantly non-random species range distributions may be consistent with the faunal overlap hypothesis.

In testing the faunal community overlap hypothesis in Bornean geometrid moths, Beck & Chey (2008) predicted that (1) species richness will decline with altitude when montane species are excluded, and (2) the distribution of species' altitudinal range midpoints will be bimodal, one mode for lowland species and one for montane species. Although this hypothesis was not found to have much explanatory power in Bornean moths (Beck & Chey 2008), both predictions are true for Mulu birds (Figures 5 and 6). Species' altitudinal range midpoints are distributed bimodally (Figure 5), and richness of lowland species declines with altitude above 300 m. Lower richness below 300 m on Mulu is probably attributable to disturbance of low-altitude forest rather than intrinsically lower richness, since undisturbed lowland forest in Borneo is known to be extremely rich in species (Smythies 1999). Overlap of two distinct communities has also been found to explain the richness peak in small-mammal communities on Mt Kinabalu, in Sabah, north Borneo, although this peak occurs at a much higher altitude (Nor 2001). Similar patterns have also been found in African birds (Romdal & Rahbek 2009).

Unfortunately, extensive forest disturbance due to widespread shifting cultivation at low altitude prevented us from surveying entire gradients on the other two Bornean mountains. The scarcity of full altitudinal gradients still covered by primary forest is unsurprising given that over 38% of Borneo's lowland forest has been converted to plantations since the mid-1970s, and 56% of the remaining lowland forest has been heavily disturbed (Gaveau *et al.* 2014). Even minor forest disturbance has been shown to have a significant effect on species assemblages of moths (Beck *et al.* 2006) and birds in Borneo (Cleary *et al.* 2007, Edwards *et al.* 2011, Johns 1996, Lambert 1992), necessitating the truncation of survey ranges on Pueh and Topap Oso. Even so, surveys on these mountains appear to have captured altitudes of peak richness, if not the total shape of richness–altitude curves. The survey analyses also illuminate the main zones of community turnover on both mountains, as evidenced by the distinct difference in ordination space between sites at 1200 m and those at lower altitudes. Interestingly, we find that peak richness occurs at higher altitude (800–1000 m) on Pueh than on Mulu and Topap Oso. This shift on Pueh may have two causes: the upward expansion of lowland species in the absence of a rich montane community of potential

competitors, and the downward expansion of the few montane species that are present on this relatively small, isolated, coastal mountain resulting from telescoping of vegetation zones. The higher altitude of peak richness on Pueh is consistent with the idea that lowland species are able to live higher on a mountain when released from competition (Terborgh & Weske 1975), in this case because of the limited montane avifauna due to its isolation and small size. Other researchers, however, have cautioned against inferring competition without considering alternative explanations (Cadena & Loiselle 2007), and more research is required to make strong claims about the mechanisms responsible for Pueh's community patterns.

MDE and other null models

Null models of McCain & Beck (2015), including the MDE, predict different richness and turnover patterns than occur on Mt Mulu. The models' poor fit is likely related to our finding that peak richness is caused deterministically by overlap of distinct lowland and montane faunal communities. Nevertheless, MDE models have been shown to have at least some explanatory power in respect to plant community distributions on Borneo's Mt Kinabalu (Grytnes *et al.* 2008), which at 4095 m is nearly twice as high as Mt Mulu. In that case, a statistical framework combining both ecological factors and the null models was most successful in predicting species richness, an approach promoted by the models' original proponents (Colwell *et al.* 2005). We do not rule out the importance of null processes in influencing bird distribution on Mulu, but faunal overlap appears to provide a better explanation of the low-mid-altitude richness hump.

Community composition

Combined ordination of all survey points from the three mountains revealed differences in community composition between mountains and between altitudes on individual mountains. Pueh and Topap Oso further differ from each other in occurrence of key species. For example, on Pueh species that are most common across the gradient were chestnut-winged babbler and yellow-bellied warbler, whereas on Topap Oso they were spectacled bulbul, Bornean barbet, red-throated barbet and blue-eared barbet. On the other hand, as dictated by common sense, sites at adjacent altitudes on the same mountain are on average more similar in respect to species composition than to sites at the same altitude on different mountains. This was especially true on Pueh and Topap Oso for the lowest three altitudes sampled (600,

800 and 1000 m). Overall, for Pueh and Topap Oso, as mountains of similar size, differences in community composition suggest that factors other than altitude (e.g. geography, climate, degree of isolation and possibly habitat disturbance on the lower slopes; cf. Lomolino 2001) are playing important roles in structuring avian communities.

On all three mountains, sites at 1200 m exhibit a significant shift in species composition from those at 1000 m and below. This community change is reflected quantitatively in richness and turnover values as well as visually on the ordination plot. This break even occurs on Pueh, with its relatively impoverished higher montane community, which suggests a primarily abiotic rather than biotic cause (Jankowski *et al.* 2013). If the 800–1000 m ceiling apparent in many lowland species' distributions on typical Bornean mountains (e.g. Mulu and Topap Oso) is caused by competition with montane avifauna, we would expect a significant uphill shift in species' upper range limits on Pueh, where many montane competitors are absent (Pueh has half as many montane species as Mulu) (Terborgh & Weske 1975). In fact, some species do appear to expand their ranges upward on Pueh relative to the other mountains (e.g. grey-headed canary-flycatcher and square-tailed drongocuckoo, *Surniculus lugubris*), contributing to its richness peak at higher altitudes. But these range-shifts are not without limits, and usually consist of only a few hundred metres, suggesting they are constrained ultimately by climate or habitat. Abiotic factors are therefore likely to play a role in limiting distributions of bird species, whether directly through physiological limits or indirectly through habitat structure (Lomolino 2001).

Conclusions

The decline of species richness with latitude is a pervasive pattern globally, but the decline of species richness with altitude is a much less uniform pattern. This study, conducted on three distinctly situated Bornean mountains, presents the first published quantitative surveys of altitudinal gradients of birds in Sundaland. Species richness peaks at 600 m on Mt Mulu (and probably Mt Topap Oso), but several hundred metres higher on Pueh. Only limited conclusions can be drawn from the surveys of partial altitudinal gradients, which highlights the importance of studying and conserving the few remaining intact forest gradients that remain in the Sunda region. Continuous gradients of altitude not only provide habitat for a wide diversity of species, but also represent crucial but diminishing opportunities to understand the processes that have produced and structured biodiversity in the past.

Patterns of lowland versus montane community richness, plus overall patterns of turnover and nestedness, on Mulu explain the formation of the mountain's low-mid altitude richness peak. This peak is caused by overlap of lowland and montane communities. However, the peak in richness does not coincide with the peak in turnover, but occurs in a lower altitudinal band where the lowland bird community is still mostly intact and some montane species begin to appear. The distribution of species ranges producing this peak is not consistent with null predictions of the mid-domain effect on Mulu, but the narrower surveyed range on the other two mountains does not allow us to rule out this effect on those mountains. The gradual and overlapping transition from lowland to highland species supports the idea that bird species ranges in Borneo are not distributed randomly with respect to each other, but rather form relatively distinct communities by altitude. This supports the idea that a faunal overlap can produce a mid-low altitude peak in richness that does not coincide with peak turnover.

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Appendix 1. Bird species recorded in this study by Bornean mountain. Classification and order follow IOC World Bird List (v 4.4). (<http://www.worldbirdnames.org/>). Species with five or more detections on Mulu are designated as 'lowland' (n = 82) or 'montane' (n = 40) based on whether >75% of detections occur above or below the altitude of peak turnover (between 900 and 1200 m). Those species with five or more detections split more evenly between these two altitudinal zones are designated mid-altitude ('mid-alt'; n = 10).

Species	English name	Topap Oso	Pueh	Mulu
<i>Arborophila hyperythra</i>	Red-breasted partridge	X		X
<i>Haematortyx sanguiniceps</i>	Crimson-headed partridge	X		montane
<i>Rollulus rouloul</i>	Crested partridge	X		X
<i>Lophura ignita</i>	Crested fireback	X		X
<i>Argusianus argus</i>	Great argus	X	X	lowland
<i>Spilornis cheela</i>	Crested serpent eagle	X	X	lowland
<i>Spilornis kinabaluensis</i>	Mountain serpent eagle	X		montane
<i>Ictinaetus malaiensis</i>	Black eagle			X
<i>Macropygia ruficeps</i>	Little cuckoo-dove		X	montane
<i>Chalcophaps indica</i>	Common emerald dove	X	X	lowland
<i>Treron curvirostra</i>	Thick-billed green pigeon		X	
<i>Ducula badia</i>	Mountain imperial pigeon	X	X	montane
<i>Centropus sinensis</i>	Greater coucal			lowland
<i>Rhinortha chlorophaea</i>	Raffles's malkoha	X	X	lowland
<i>Zanlostomus javanicus</i>	Red-billed malkoha		X	lowland
<i>Phaenicophaeus curvirostris</i>	Chestnut-breasted malkoha	X	X	X
<i>Chrysococcyx xanthorhynchus</i>	Violet cuckoo			X
<i>Chrysococcyx minutillus</i>	Little bronze cuckoo		X	X
<i>Cacomantis sonneratii</i>	Banded bay cuckoo	X	X	lowland
<i>Cacomantis merulinus</i>	Plaintive cuckoo	X	X	lowland
<i>Cacomantis sepulcralis</i>	Rusty-breasted cuckoo	X	X	
<i>Surmiculus lugubris</i>	Square-tailed drongo-cuckoo	X	X	X
<i>Hierococcyx vagans</i>	Moustached hawk-cuckoo			lowland
<i>Hierococcyx bocki</i>	Dark hawk-cuckoo			montane
<i>Hierococcyx fugax</i>	Malaysian hawk-cuckoo	X	X	X
<i>Cuculus micropterus</i>	Indian cuckoo	X	X	X
<i>Cuculus lepidus</i>	Sunda cuckoo	X		montane
<i>Glaucidium brodiei</i>	Collared owl			montane
<i>Harpactes kasumba</i>	Red-naped trogon	X	X	lowland
<i>Harpactes diardii</i>	Diard's trogon	X	X	lowland
<i>Harpactes orrhophaeus</i>	Cinnamon-rumped trogon		X	
<i>Harpactes duvaucelii</i>	Scarlet-rumped trogon	X	X	lowland
<i>Harpactes oreskios</i>	Orange-breasted trogon	X	X	X
<i>Actenoides concretus</i>	Rufous-collared kingfisher	X	X	
<i>Lacedo pulchella</i>	Banded kingfisher	X	X	lowland
<i>Ceyx erithaca</i>	Oriental dwarf kingfisher	X		X
<i>Nyctyornis amictus</i>	Red-bearded bee-eater	X	X	X
<i>Bericornis comatus</i>	White-crowned hornbill	X	X	
<i>Buceros rhinoceros</i>	Rhinoceros hornbill	X	X	mid-alt
<i>Rhinoplax vigil</i>	Helmeted hornbill	X	X	lowland
<i>Anorrhinus galeritus</i>	Bushy-crested hornbill	X	X	lowland
<i>Rhyticeros undulatus</i>	Wreathed hornbill	X	X	X
<i>Megalaima chrysopogon</i>	Golden-whiskered barbet	X	X	lowland
<i>Megalaima mystacophanos</i>	Red-throated barbet	X	X	lowland
<i>Megalaima monticola</i>	Mountain barbet	X	X	montane
<i>Megalaima henricii</i>	Yellow-crowned barbet	X		lowland
<i>Megalaima pulcherrima</i>	Golden-naped barbet			montane
<i>Megalaima duvaucelii</i>	Blue-eared barbet	X	X	lowland
<i>Megalaima eximia</i>	Bornean barbet	X	X	mid-alt
<i>Caloramphus fuliginosus</i>	Brown barbet	X		lowland
<i>Picumnus innominatus</i>	Speckled piculet	X	X	
<i>Sasia abnormis</i>	Rufous piculet	X	X	lowland
<i>Chrysophlegma miniaceum</i>	Banded woodpecker	X	X	X
<i>Chrysophlegma mentale</i>	Checker-throated woodpecker	X	X	mid-alt
<i>Picus puniceus</i>	Crimson-winged woodpecker	X	X	X
<i>Dinopium rafflesii</i>	Olive-backed woodpecker		X	X
<i>Blythipicus rubiginosus</i>	Maroon woodpecker	X	X	mid-alt
<i>Reinwardtipicus validus</i>	Orange-backed woodpecker	X		X
<i>Micropternus brachyurus</i>	Rufous woodpecker	X	X	lowland

Appendix 1. Continued

Species	English name	Topap Oso	Pueh	Mulu
<i>Meiglyptes tristis</i>	Buff-rumped woodpecker		X	X
<i>Meiglyptes tukki</i>	Buff-necked woodpecker		X	lowland
<i>Loriculus galgulus</i>	Blue-crowned hanging parrot		X	lowland
<i>Calyptomena viridis</i>	Green broadbill	X	X	lowland
<i>Calyptomena hosii</i>	Hose's broadbill	X		X
<i>Calyptomena whiteheadi</i>	Whitehead's broadbill			X
<i>Psarisomus dalhousiae</i>	Long-tailed broadbill			X
<i>Eurylaimus javanicus</i>	Banded broadbill	X	X	lowland
<i>Eurylaimus ochromalus</i>	Black-and-yellow broadbill	X	X	lowland
<i>Corydon sumatranus</i>	Dusky broadbill		X	
<i>Hydroornis schwaneri</i>	Bornean banded pitta	X		
<i>Erythropitta arquata</i>	Blue-banded pitta	X		
<i>Erythropitta granatina</i>	Garnet pitta	X		
<i>Gerygone sulphurea</i>	Golden-bellied gerygone	X	X	montane
<i>Hemipus picatus</i>	Bar-winged flycatcher-shrike	X	X	X
<i>Hemipus hirundinaceus</i>	Black-winged flycatcher-shrike	X	X	
<i>Tephrodornis virgatus</i>	Large woodshrike	X	X	lowland
<i>Philentoma pyrhoptera</i>	Rufous-winged philentoma	X	X	lowland
<i>Philentoma velata</i>	Maroon-breasted philentoma	X	X	lowland
<i>Aegithina viridissima</i>	Green iora	X	X	lowland
<i>Coracina larvata</i>	Sunda cuckooshrike	X		X
<i>Coracina fimbriata</i>	Lesser cuckooshrike	X	X	X
<i>Pericrocotus igneus</i>	Fiery minivet		X	
<i>Pericrocotus solaris</i>	Grey-chinned minivet	X		montane
<i>Pericrocotus speciosus</i>	Scarlet minivet	X	X	X
<i>Pachycephala hypoxantha</i>	Bornean whistler		X	montane
<i>Erpornis zantholeuca</i>	White-bellied erpornis	X	X	lowland
<i>Pteruthius aeralatus</i>	Blyth's shrike-babbler	X	X	montane
<i>Oriolus xanthonotus</i>	Dark-throated oriole	X	X	X
<i>Oriolus cruentus</i>	Black-and-crimson oriole	X		mid-alt
<i>Dicrurus leucophaeus</i>	Ashy drongo	X	X	montane
<i>Dicrurus aeneus</i>	Bronzed drongo	X	X	lowland
<i>Dicrurus hottentottus</i>	Hair-crested drongo	X	X	mid-alt
<i>Dicrurus paradiseus</i>	Greater racket-tailed drongo	X	X	lowland
<i>Rhipidura albicollis</i>	White-throated fantail	X	X	montane
<i>Rhipidura perlata</i>	Spotted fantail	X	X	lowland
<i>Hypothymis azurea</i>	Black-naped monarch	X	X	lowland
<i>Terpsiphone paradisi</i>	Asian paradise flycatcher	X	X	lowland
<i>Platylophus galericulatus</i>	Crested jay	X	X	X
<i>Platysmurus leucopterus</i>	Black magpie	X		X
<i>Cissa jefferyi</i>	Bornean green magpie			montane
<i>Dendrocitta cinerascens</i>	Bornean treepie	X		montane
<i>Corvus enca</i>	Slender-billed crow			X
<i>Eupetes macrocerus</i>	Rail-babbler	X		
<i>Culicicapa ceylonensis</i>	Grey-headed canary-flycatcher	X	X	lowland
<i>Pycnonotus melanoleucos</i>	Black-and-white bulbul	X		X
<i>Pycnonotus atriceps</i>	Black-headed bulbul			X
<i>Pycnonotus montis</i>	Bornean bulbul	X		montane
<i>Pycnonotus squamatus</i>	Scaly-breasted bulbul	X	X	lowland
<i>Pycnonotus cyaniventris</i>	Grey-bellied bulbul	X	X	lowland
<i>Pycnonotus eutilotus</i>	Puff-backed bulbul			lowland
<i>Pycnonotus flavescens</i>	Flavescens bulbul			montane
<i>Pycnonotus plumosus</i>	Olive-winged bulbul	X	X	X
<i>Pycnonotus brunneus</i>	Asian red-eyed bulbul	X	X	lowland
<i>Pycnonotus erythroptalmos</i>	Spectacled bulbul	X	X	lowland
<i>Alophoixus finschii</i>	Finsch's bulbul	X		X
<i>Alophoixus ochraceus</i>	Ochraceous bulbul	X	X	montane
<i>Alophoixus bres</i>	Grey-cheeked bulbul	X	X	lowland
<i>Alophoixus phaeocephalus</i>	Yellow-bellied bulbul	X	X	lowland
<i>Tricholestes criniger</i>	Hairy-backed bulbul	X	X	lowland
<i>Iole olivacea</i>	Buff-vented bulbul	X	X	lowland

Appendix 1. Continued

Species	English name	Topap Oso	Pueh	Mulu
<i>Ixos malaccensis</i>	Streaked bulbul	X	X	lowland
<i>Hemixos flavala</i>	Ashy bulbul	X	X	mid-alt
<i>Abroscopus superciliaris</i>	Yellow-bellied warbler	X	X	montane
<i>Phyllergates cuculatus</i>	Mountain tailorbird		X	montane
<i>Horornis vulcanius</i>	Sunda bush warbler			montane
<i>Urosphena whiteheadi</i>	Bornean stubtail			montane
<i>Phylloscopus trivirgatus</i>	Mountain leaf warbler		X	montane
<i>Seicercus montis</i>	Yellow-breasted warbler	X	X	montane
<i>Orthotomus atrogularis</i>	Dark-necked tailorbird	X	X	X
<i>Orthotomus sericeus</i>	Rufous-tailed tailorbird	X	X	lowland
<i>Orthotomus ruficeps</i>	Ashy tailorbird	X	X	lowland
<i>Pomatorhinus montanus</i>	Chestnut-backed scimitar babbler	X	X	mid-alt
<i>Stachyris nigriceps</i>	Grey-throated babbler	X	X	mid-alt
<i>Stachyris poliocephala</i>	Grey-headed babbler	X	X	X
<i>Stachyris maculata</i>	Chestnut-rumped babbler	X	X	lowland
<i>Stachyris leucotis</i>	White-necked babbler	X	X	X
<i>Stachyris nigricollis</i>	Black-throated babbler	X	X	lowland
<i>Stachyris erythroptera</i>	Chestnut-winged babbler	X	X	lowland
<i>Stachyridopsis rufifrons</i>	Rufous-fronted babbler	X	X	lowland
<i>Macronus ptilosus</i>	Fluffy-backed tit-babbler	X	X	lowland
<i>Alcippe brunneicauda</i>	Brown fulvetta	X	X	lowland
<i>Napothera crassa</i>	Mountain wren-babbler			montane
<i>Napothera epilepidota</i>	Eyebrowed wren-babbler	X	X	lowland
<i>Malacocincla sepiaria</i>	Horsfield's babbler	X	X	lowland
<i>Malacocincla malaccensis</i>	Short-tailed babbler	X	X	lowland
<i>Malacopteron magnirostre</i>	Moustached babbler	X	X	lowland
<i>Malacopteron affine</i>	Sooty-capped babbler	X	X	lowland
<i>Malacopteron cinereum</i>	Scaly-crowned babbler	X	X	lowland
<i>Malacopteron magnum</i>	Rufous-crowned babbler	X	X	lowland
<i>Trichastoma rostratum</i>	White-chested babbler			lowland
<i>Trichastoma bicolor</i>	Ferruginous babbler	X	X	X
<i>Pellorneum pyrrogenys</i>	Temminck's babbler	X	X	mid-alt
<i>Pellorneum capiistratum</i>	Black-capped babbler	X		lowland
<i>Garrulax palliatus</i>	Sunda laughingthrush			montane
<i>Garrulax mitratus</i>	Chestnut-capped laughingthrush	X		montane
<i>Garrulax calvus</i>	Bare-headed laughingthrush	X		montane
<i>Yuhina everetti</i>	Chestnut-crested yuhina	X	X	montane
<i>Oculocincta squamifrons</i>	Pygmy white-eye	X	X	X
<i>Chlorocharis emiliae</i>	Mountain blackeye			montane
<i>Zosterops palpebrosus</i>	Oriental white-eye		X	
<i>Zosterops atricapilla</i>	Black-capped white-eye	X		montane
<i>Zosterops everetti</i>	Everett's white-eye	X	X	X
<i>Irena puella</i>	Asian fairy-bluebird	X	X	lowland
<i>Sitta frontalis</i>	Velvet-fronted nuthatch	X	X	lowland
<i>Gracula religiosa</i>	Common hill myna			X
<i>Chlamydochaera jefferyi</i>	Fruithunter			X
<i>Copsychus saularis</i>	Oriental magpie-robin			X
<i>Copsychus pyrropygus</i>	Rufous-tailed shama	X		X
<i>Copsychus malabaricus</i>	White-rumped shama	X	X	lowland
<i>Cyornis unicolor</i>	Pale blue flycatcher	X	X	lowland
<i>Cyornis caerulatus</i>	Sunda blue flycatcher	X		
<i>Cyornis superbus</i>	Bornean blue flycatcher	X	X	lowland
<i>Cyornis turcosus</i>	Malaysian blue flycatcher		X	lowland
<i>Cyornis concretus</i>	White-tailed flycatcher	X	X	X
<i>Cyornis umbratilis</i>	Grey-chested jungle flycatcher	X	X	lowland
<i>Eumyias thalassinus</i>	Verditer flycatcher	X	X	
<i>Brachypteryx montana</i>	White-browed shortwing			montane
<i>Vauriella gularis</i>	Eyebrowed jungle flycatcher	X		montane
<i>Enicurus ruficapillus</i>	Chestnut-naped forktail			X
<i>Enicurus leschenaulti</i>	White-crowned forktail		X	
<i>Enicurus borneensis</i>	Bornean forktail			X
<i>Ficedula dumetoria</i>	Rufous-chested flycatcher	X	X	X

Appendix 1. Continued

Species	English name	Topap Oso	Pueh	Mulu
<i>Ficedula hyperythra</i>	Snowy-browed flycatcher		X	montane
<i>Ficedula westermanni</i>	Little pied flycatcher		X	X
<i>Muscicapella hodgsoni</i>	Pygmy flycatcher			montane
<i>Chloropsis sonnerati</i>	Greater green leafbird	X	X	lowland
<i>Chloropsis cyanopogon</i>	Lesser green leafbird	X	X	lowland
<i>Chloropsis cochinchinensis</i>	Blue-winged leafbird		X	
<i>Chloropsis kinabaluensis</i>	Bornean leafbird	X		montane
<i>Prionochilus maculatus</i>	Yellow-breasted flowerpecker	X	X	lowland
<i>Prionochilus xanthopygius</i>	Yellow-rumped flowerpecker	X	X	lowland
<i>Prionochilus thoracicus</i>	Scarlet-breasted flowerpecker	X		
<i>Dicaeum chrysorrheum</i>	Yellow-vented flowerpecker		X	
<i>Dicaeum trigonostigma</i>	Orange-bellied flowerpecker	X	X	lowland
<i>Dicaeum monticolum</i>	Black-sided flowerpecker	X	X	montane
<i>Dicaeum cruentatum</i>	Scarlet-backed flowerpecker		X	
<i>Chalcoparia singalensis</i>	Ruby-cheeked sunbird		X	X
<i>Anthreptes simplex</i>	Plain sunbird	X	X	lowland
<i>Anthreptes malacensis</i>	Brown-throated sunbird		X	X
<i>Hypogramma hypogrammicum</i>	Purple-naped sunbird	X	X	lowland
<i>Leptocoma brasiliana</i>	Van Hasselt's sunbird		X	
<i>Leptocoma calcostetha</i>	Copper-throated sunbird		X	
<i>Cinnyris jugularis</i>	Olive-backed sunbird			X
<i>Aethopyga siparaja</i>	Crimson sunbird	X	X	lowland
<i>Aethopyga temminckii</i>	Temminck's sunbird	X	X	montane
<i>Arachnothera longirostra</i>	Little spiderhunter	X	X	lowland
<i>Arachnothera crassirostris</i>	Thick-billed spiderhunter		X	
<i>Arachnothera robusta</i>	Long-billed spiderhunter	X	X	
<i>Arachnothera flavigaster</i>	Spectacled spiderhunter		X	
<i>Arachnothera chrysogenys</i>	Yellow-eared spiderhunter		X	X
<i>Arachnothera modesta</i>	Grey-breasted spiderhunter	X	X	lowland
<i>Arachnothera everetti</i>	Bornean spiderhunter	X		X
<i>Arachnothera juliae</i>	Whitehead's spiderhunter			X
	Totals	155	151	187

Appendix 2. Simper scores reflecting species' contributions to pairwise differences between groups of sites from specified mountains and altitudes in Borneo. The 'csum' represents the cumulative sum of the contribution of the top n species. The 15 most influential species are displayed for each comparison. Species found to be influential in 10 or more of these 14 pairwise comparisons are marked with a single asterisk (*), while species found to be influential in only one of these pairwise comparisons are marked with a double asterisk (**).

Species	mean ± SD	csum
Comparison: Mt Mulu 600 m to Mt Pueh 600 m		
<i>Megalaima mystacophanos</i>	0.033 ± 0.02	0.033
<i>Megalaima chrysopogon</i> *	0.031 ± 0.023	0.064
<i>Pomatorhinus montanus</i> *	0.027 ± 0.021	0.091
<i>Megalaima duvaucelii</i>	0.025 ± 0.02	0.115
<i>Stachyris erythroptera</i> *	0.023 ± 0.019	0.138
<i>Megalaima henricii</i>	0.021 ± 0.021	0.16
<i>Pycnonotus erythroptalmos</i>	0.02 ± 0.018	0.18
<i>Eurylaimus ochromalus</i>	0.019 ± 0.017	0.199
<i>Alcippe brunneicauda</i> *	0.018 ± 0.017	0.217
<i>Rhyticeros undulatus</i>	0.018 ± 0.017	0.235
<i>Rhipidura perlata</i>	0.018 ± 0.018	0.252
<i>Megalaima eximia</i> *	0.017 ± 0.022	0.269
<i>Orthotomus sericeus</i>	0.016 ± 0.019	0.285
<i>Pycnonotus cyaniventris</i>	0.016 ± 0.016	0.3
<i>Culicicapa ceylonensis</i>	0.015 ± 0.016	0.316
Comparison: Mt Mulu 600 m to Mt Topap Oso 600 m		
<i>Megalaima chrysopogon</i> *	0.047 ± 0.021	0.047
<i>Pomatorhinus montanus</i> *	0.044 ± 0.021	0.091
<i>Megalaima mystacophanos</i>	0.035 ± 0.023	0.126
<i>Megalaima duvaucelii</i>	0.03 ± 0.022	0.155
<i>Stachyris erythroptera</i> *	0.029 ± 0.021	0.185
<i>Alcippe brunneicauda</i> *	0.025 ± 0.022	0.21
<i>Megalaima henricii</i>	0.024 ± 0.022	0.234
<i>Eurylaimus ochromalus</i>	0.022 ± 0.022	0.256
<i>Megalaima eximia</i> *	0.021 ± 0.026	0.277
<i>Pycnonotus cyaniventris</i>	0.02 ± 0.019	0.297
<i>Culicicapa ceylonensis</i>	0.018 ± 0.02	0.315
<i>Pycnonotus erythroptalmos</i>	0.018 ± 0.017	0.333
<i>Argusianus argus</i>	0.017 ± 0.014	0.35
<i>Rhipidura perlata</i>	0.017 ± 0.019	0.367
<i>Malacopteron magnum</i> **	0.013 ± 0.016	0.379
Comparison: Mt Pueh 600 m to Mt Topap Oso 600 m		
<i>Stachyris erythroptera</i> *	0.041 ± 0.026	0.041
<i>Alcippe brunneicauda</i> *	0.032 ± 0.024	0.073
<i>Pycnonotus erythroptalmos</i>	0.026 ± 0.023	0.099
<i>Rhyticeros undulatus</i>	0.025 ± 0.023	0.124
<i>Megalaima chrysopogon</i> *	0.024 ± 0.025	0.148
<i>Eurylaimus ochromalus</i>	0.023 ± 0.021	0.171
<i>Pomatorhinus montanus</i> *	0.022 ± 0.021	0.194
<i>Orthotomus sericeus</i>	0.022 ± 0.026	0.216
<i>Rhipidura perlata</i>	0.02 ± 0.023	0.235
<i>Megalaima duvaucelii</i>	0.018 ± 0.02	0.254
<i>Buceros rhinoceros</i>	0.018 ± 0.022	0.272
<i>Malacopteron magnirostre</i>	0.017 ± 0.02	0.289
<i>Argusianus argus</i>	0.017 ± 0.018	0.306
<i>Malacocincla malaccensis</i> **	0.017 ± 0.02	0.323
<i>Cyornis unicolor</i> **	0.017 ± 0.023	0.34
Comparison: Mt Pueh 800 m to Mt Topap Oso 800 m		
<i>Stachyris erythroptera</i> *	0.049 ± 0.028	0.049
<i>Megalaima chrysopogon</i> *	0.035 ± 0.023	0.083
<i>Alcippe brunneicauda</i> *	0.03 ± 0.024	0.114
<i>Pycnonotus erythroptalmos</i>	0.023 ± 0.021	0.137
<i>Culicicapa ceylonensis</i>	0.021 ± 0.022	0.157

Appendix 2. Continued

Species	mean ± SD	csum
<i>Megalaima duvaucelii</i>	0.02 ± 0.02	0.178
<i>Rhyticeros undulatus</i>	0.019 ± 0.023	0.197
<i>Oriolus xanthonotus</i>	0.018 ± 0.022	0.215
<i>Malacopteron magnirostre</i>	0.018 ± 0.017	0.233
<i>Stachyridopsis rufifrons</i>	0.017 ± 0.019	0.25
<i>Argusianus argus</i>	0.017 ± 0.017	0.267
<i>Pomatorhinus montanus</i> *	0.017 ± 0.019	0.284
<i>Arachnothera longirostra</i>	0.016 ± 0.021	0.3
<i>Megalaima mystacophanos</i>	0.016 ± 0.017	0.317
<i>Eurylaimus ochromalus</i>	0.016 ± 0.019	0.333
Comparison: Mt Mulu 900 m to Mt Pueh 800/1000 m		
<i>Alcippe brunneicauda</i> *	0.038 ± 0.026	0.038
<i>Megalaima chrysopogon</i> *	0.038 ± 0.029	0.076
<i>Abroscopus supercilii</i>	0.035 ± 0.032	0.111
<i>Stachyris erythroptera</i> *	0.034 ± 0.029	0.146
<i>Pomatorhinus montanus</i> *	0.028 ± 0.027	0.174
<i>Culicicapa ceylonensis</i>	0.022 ± 0.022	0.196
<i>Rhipidura perlata</i>	0.022 ± 0.023	0.218
<i>Pellorneum pyrrogenys</i>	0.021 ± 0.026	0.239
<i>Hemixos flava</i>	0.021 ± 0.023	0.26
<i>Pycnonotus erythroptalmos</i>	0.019 ± 0.022	0.279
<i>Megalaima eximia</i> *	0.017 ± 0.022	0.297
<i>Macronus ptilosus</i>	0.017 ± 0.02	0.314
<i>Arachnothera longirostra</i>	0.017 ± 0.02	0.331
<i>Megalaima duvaucelii</i>	0.016 ± 0.021	0.347
<i>Oriolus xanthonotus</i>	0.015 ± 0.02	0.362
Comparison: Mt Mulu 900 m to Mt Topap Oso 800/1000 m		
<i>Pomatorhinus montanus</i> *	0.042 ± 0.044	0.042
<i>Stachyris erythroptera</i> *	0.039 ± 0.039	0.082
<i>Alcippe brunneicauda</i> *	0.033 ± 0.032	0.115
<i>Macronus ptilosus</i>	0.03 ± 0.035	0.145
<i>Megalaima chrysopogon</i> *	0.03 ± 0.033	0.174
<i>Hemixos flava</i>	0.029 ± 0.036	0.203
<i>Pellorneum pyrrogenys</i>	0.027 ± 0.042	0.231
<i>Pycnonotus erythroptalmos</i>	0.026 ± 0.028	0.257
<i>Culicicapa ceylonensis</i>	0.025 ± 0.03	0.282
<i>Argusianus argus</i>	0.022 ± 0.027	0.304
<i>Megalaima eximia</i> *	0.021 ± 0.027	0.326
<i>Megalaima mystacophanos</i>	0.02 ± 0.026	0.346
<i>Rhipidura perlata</i>	0.02 ± 0.031	0.366
<i>Stachyridopsis rufifrons</i>	0.018 ± 0.026	0.384
<i>Buceros rhinoceros</i>	0.017 ± 0.024	0.401
Comparison: Mt Pueh 1000 m to Mt Topap Oso 1000 m		
<i>Abroscopus supercilii</i>	0.061 ± 0.016	0.061
<i>Megalaima chrysopogon</i> *	0.031 ± 0.024	0.092
<i>Stachyris erythroptera</i> *	0.028 ± 0.021	0.12
<i>Alcippe brunneicauda</i> *	0.028 ± 0.021	0.148
<i>Hemixos flava</i>	0.026 ± 0.023	0.174
<i>Megalaima eximia</i> *	0.024 ± 0.024	0.198
<i>Rhipidura perlata</i>	0.024 ± 0.025	0.222
<i>Culicicapa ceylonensis</i>	0.022 ± 0.021	0.244
<i>Pomatorhinus montanus</i> *	0.019 ± 0.019	0.263
<i>Arachnothera longirostra</i>	0.019 ± 0.019	0.282
<i>Megalaima mystacophanos</i>	0.018 ± 0.018	0.3
<i>Pycnonotus erythroptalmos</i>	0.017 ± 0.017	0.316
<i>Pellorneum pyrrogenys</i>	0.015 ± 0.016	0.332
<i>Stachyridopsis rufifrons</i>	0.015 ± 0.017	0.346
<i>Eurylaimus ochromalus</i>	0.013 ± 0.016	0.36
Comparison: Mt Pueh 1200 m to Mt Topap Oso 1200 m		
<i>Abroscopus supercilii</i>	0.055 ± 0.019	0.055
<i>Megalaima chrysopogon</i> *	0.045 ± 0.02	0.099

Appendix 2. Continued

Species	mean ± SD	csum
<i>Pomatorhinus montanus</i> *	0.04 ± 0.021	0.139
<i>Pteruthius aeralatus</i>	0.033 ± 0.019	0.172
<i>Seiurus montis</i>	0.029 ± 0.023	0.201
<i>Pellorneum pyrrogenys</i>	0.026 ± 0.023	0.227
<i>Megalaima monticola</i>	0.024 ± 0.011	0.251
<i>Alcippe brunneicauda</i> *	0.023 ± 0.02	0.275
<i>Hemixos flavala</i>	0.023 ± 0.021	0.298
<i>Culicicapa ceylonensis</i>	0.02 ± 0.019	0.317
<i>Megalaima eximia</i> *	0.02 ± 0.018	0.337
<i>Aethopyga temminckii</i>	0.02 ± 0.019	0.357
<i>Arachnothera longirostra</i>	0.02 ± 0.018	0.376
<i>Megalaima mystacophanos</i>	0.019 ± 0.019	0.395
<i>Stachyris nigriceps</i>	0.018 ± 0.02	0.413
Comparison: Mt Mulu 1200 m to Mt Pueh 1200 m		
<i>Megalaima monticola</i>	0.04 ± 0.026	0.04
<i>Megalaima pulcherrima</i>	0.039 ± 0.03	0.079
<i>Abroscopus superciliosus</i>	0.039 ± 0.02	0.118
<i>Megalaima chrysopogon</i> *	0.039 ± 0.02	0.157
<i>Pomatorhinus montanus</i> *	0.029 ± 0.02	0.186
<i>Megalaima eximia</i> *	0.029 ± 0.02	0.215
<i>Garrulax mitratus</i>	0.026 ± 0.021	0.241
<i>Alcippe brunneicauda</i> *	0.024 ± 0.017	0.265
<i>Aethopyga temminckii</i>	0.023 ± 0.02	0.288
<i>Pteruthius aeralatus</i>	0.023 ± 0.018	0.311
<i>Pellorneum pyrrogenys</i>	0.021 ± 0.017	0.333
<i>Seiurus montis</i>	0.021 ± 0.018	0.353
<i>Hemixos flavala</i>	0.018 ± 0.016	0.371
<i>Arachnothera longirostra</i>	0.017 ± 0.014	0.388
<i>Culicicapa ceylonensis</i>	0.016 ± 0.015	0.404
Comparison: Mt Mulu 1200 m to Mt Topap Oso 1200 m		
<i>Megalaima pulcherrima</i>	0.059 ± 0.043	0.059
<i>Megalaima monticola</i>	0.047 ± 0.027	0.106
<i>Garrulax mitratus</i>	0.038 ± 0.031	0.144
<i>Megalaima eximia</i> *	0.038 ± 0.026	0.182
<i>Aethopyga temminckii</i>	0.038 ± 0.035	0.219
<i>Pteruthius aeralatus</i>	0.029 ± 0.029	0.249
<i>Pellorneum pyrrogenys</i>	0.024 ± 0.023	0.273
<i>Alophoixus ochraceus</i>	0.023 ± 0.024	0.296
<i>Hemixos flavala</i>	0.023 ± 0.02	0.319
<i>Dendrocitta cinerascens</i> **	0.022 ± 0.027	0.34
<i>Gerygone sulphurea</i> **	0.02 ± 0.031	0.36
<i>Pomatorhinus montanus</i> *	0.019 ± 0.02	0.379
<i>Megalaima chrysopogon</i> *	0.018 ± 0.019	0.397
<i>Seiurus montis</i>	0.018 ± 0.024	0.416
<i>Pycnonotus montis</i>	0.018 ± 0.019	0.433
Comparison: Mt Pueh 600 m to Mt Pueh 1200 m		
<i>Abroscopus superciliosus</i>	0.044 ± 0.017	0.044
<i>Megalaima chrysopogon</i> *	0.027 ± 0.021	0.071
<i>Pteruthius aeralatus</i>	0.026 ± 0.016	0.096
<i>Megalaima eximia</i> *	0.025 ± 0.017	0.121
<i>Stachyris erythroptera</i> *	0.024 ± 0.018	0.145
<i>Pomatorhinus montanus</i> *	0.023 ± 0.019	0.168
<i>Seiurus montis</i>	0.022 ± 0.019	0.191
<i>Pycnonotus erythropteros</i>	0.021 ± 0.017	0.212
<i>Pellorneum pyrrogenys</i>	0.021 ± 0.021	0.233

Appendix 2. Continued

Species	mean ± SD	csum
<i>Hemixos flavala</i>	0.021 ± 0.018	0.254
<i>Alcippe brunneicauda</i> *	0.02 ± 0.017	0.274
<i>Rhyticeros undulatus</i>	0.017 ± 0.016	0.291
<i>Eurylaimus ochromalus</i>	0.016 ± 0.014	0.307
<i>Orthotomus sericeus</i>	0.016 ± 0.017	0.322
<i>Stachyris nigriceps</i>	0.016 ± 0.015	0.338
Comparison: Mt Topap Oso 600 m to Mt Topap Oso 1200 m		
<i>Megalaima monticola</i>	0.043 ± 0.017	0.043
<i>Megalaima eximia</i> *	0.032 ± 0.024	0.075
<i>Pycnonotus erythropteros</i>	0.026 ± 0.023	0.101
<i>Argusianus argus</i>	0.025 ± 0.024	0.126
<i>Alcippe brunneicauda</i> *	0.024 ± 0.025	0.15
<i>Megalaima chrysopogon</i> *	0.023 ± 0.025	0.173
<i>Megalaima mystacophanos</i>	0.023 ± 0.025	0.196
<i>Culicicapa ceylonensis</i>	0.022 ± 0.024	0.218
<i>Hemixos flavala</i>	0.022 ± 0.024	0.239
<i>Pycnonotus montis</i>	0.021 ± 0.024	0.261
<i>Megalaima duvaucelii</i>	0.02 ± 0.023	0.281
<i>Pomatorhinus montanus</i> *	0.019 ± 0.023	0.3
<i>Stachyridopsis rufifrons</i>	0.019 ± 0.024	0.319
<i>Alophoixus ochraceus</i>	0.019 ± 0.023	0.337
<i>Megalaima henricii</i>	0.018 ± 0.023	0.355
Comparison: Mt Mulu 600 m to Mt Mulu 1200 m		
<i>Megalaima chrysopogon</i> *	0.046 ± 0.021	0.046
<i>Megalaima monticola</i>	0.042 ± 0.027	0.088
<i>Megalaima pulcherrima</i>	0.041 ± 0.031	0.129
<i>Megalaima mystacophanos</i>	0.034 ± 0.021	0.163
<i>Pomatorhinus montanus</i> *	0.033 ± 0.021	0.197
<i>Alcippe brunneicauda</i> *	0.033 ± 0.019	0.229
<i>Megalaima duvaucelii</i>	0.029 ± 0.02	0.258
<i>Megalaima eximia</i> *	0.028 ± 0.026	0.286
<i>Aethopyga temminckii</i>	0.026 ± 0.024	0.312
<i>Garrulax mitratus</i>	0.026 ± 0.022	0.339
<i>Stachyris erythroptera</i> *	0.026 ± 0.018	0.365
<i>Megalaima henricii</i>	0.022 ± 0.021	0.387
<i>Pteruthius aeralatus</i>	0.02 ± 0.02	0.407
<i>Eurylaimus ochromalus</i>	0.018 ± 0.019	0.426
<i>Hemixos flavala</i>	0.018 ± 0.014	0.444
Comparison: Mt Mulu 50 m to Mt Mulu 1800 m		
<i>Megalaima duvaucelii</i>	0.051 ± 0.011	0.051
<i>Pycnonotus erythropteros</i>	0.049 ± 0.013	0.1
<i>Malacopteron affine</i> **	0.045 ± 0.017	0.145
<i>Megalaima mystacophanos</i>	0.04 ± 0.021	0.185
<i>Megalaima pulcherrima</i>	0.04 ± 0.018	0.224
<i>Stachyris erythroptera</i> *	0.033 ± 0.016	0.257
<i>Horornis vulcanius</i> **	0.03 ± 0.021	0.287
<i>Stachyris maculata</i> **	0.029 ± 0.014	0.316
<i>Arachnothera longirostra</i>	0.026 ± 0.019	0.343
<i>Garrulax mitratus</i>	0.024 ± 0.018	0.367
<i>Phylloscopus trivirgatus</i> **	0.024 ± 0.014	0.391
<i>Stachyris nigricollis</i> **	0.024 ± 0.017	0.414
<i>Eurylaimus ochromalus</i>	0.022 ± 0.018	0.437
<i>Megalaima chrysopogon</i> *	0.022 ± 0.017	0.459
<i>Orthotomus sericeus</i>	0.021 ± 0.017	0.48