

Depositional environments, sediment characteristics, palaeoecological analysis and environmental assessment of an internationally protected shallow Mediterranean lagoon, Gialova Lagoon – Navarino Bay, Greece

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ABSTRACT: This study presents sedimentological, palaeoecological and geochemical data from a shallow Mediterranean coastal lagoon which has been severely influenced by human intervention over the last 70 years. The Gialova Lagoon is protected by international conventions and is listed in the Natura 2000 European Community Network as Special Protection Area (SPA) and Site of Community Importance (SCI). The spatial variability of sediment characteristics such as grain size, total organic carbon (TOC) and moment measures, mean, sorting, kurtosis and skewness were calculated. Moreover, micro- and macrofossil and sediment geochemical analyses were carried out on six gravity core samples. Study of the above parameters indicates that the anthropogenic impact and intervention are reflected in the micro- (ostracods, foraminifera, charophytes) and macrofossil (molluscs) taxa corresponding to different depositional environmental facies, representing a brackish lagoon with the influence of (a) fresh water inflow, (b) shallow marine environment and (c) hypoxic and dystrophic conditions. The geochemical characteristics and the calculation of the degree of sediment contamination using enrichment factors (EF), contamination factors (C_f^i) and the index of geo-accumulation (I_{geo}) indicate a recent relative improvement of the lagoon towards the upper layers of the gravity cores, rendering the lagoon as unpolluted to moderately polluted. This combinatorial study of sediment geochemical characteristics, as well as the downcore micro- and macrofossil assemblages, can be considered as a baseline for future monitoring in accordance with European Union directives, and for any future engineering interventions for the lagoon environmental maintenance and conservation; as this is the first time that geochemical and downcore palaeoecological data have been presented from this lagoon.



KEY WORDS: coastal lagoon, Mediterranean Sea, palaeoecology, pollution index, sedimentology

Coastal lagoons are areas of relatively shallow water, partly isolated from the sea by sandy or shingly barriers (Kjerfve 1994; Bird 2008). Their evolution depends on the tectonic and geomorphological conditions prevailing in the adjacent area and the relative sea level changes. Lagoons are amongst the most threatened aquatic ecosystems, being under constant global change pressures such as sea level rise, storm and river flooding, as well as human activities (Kjerfve 1994; Bird 2008). Many detrimental impacts have been observed as a result of intensive agriculture, aquaculture, industry, overexploitation of the water resources, pollution due to human activities and urbanisation, intense pasturing and overfishing. Human and engineering activities and their impacts on coastal lagoons are reviewed in detail by Duck & Silva (2012), presenting case studies of human intervention in the hydromorphology and physical processes in lagoons.

In Greece there are 24 lagoons, with a total surface of 24,500 ha, of which 10% are natural, 85 % partially natural and 5 % man made (Greek Coastal Zone Management Report 2006). Lagoon sedimentological environments are highly dy-

namic and, due to their location between the land and sea as marginal environments, they present both marine and continental characteristics. Coastal lagoons often receive water run-off from agricultural arable lands that is enriched in nutrients, thus causing eutrophication. Moreover, bottom sediments reload the water column with nutrients via decomposition of the organic matter (Golterman 2004) and thus their grain size and geochemical characteristics are an important geological and environmental factor in the research of aquatic ecosystems. Subsequently, the above-mentioned special features of lagoonal environments and the annual fluctuation of these abiotic factors have a significant effect on the organisms that inhabit these environments, and thus considerably affect the ecosystems and their respective biological assemblages.

Lagoon sediments are considered to receive anthropogenic input of pollutants in urban and suburban areas (Reimann & Caritat 2005; Bellucci *et al.* 2010). In order to assess sediment quality, metal enrichment factors and pollution indices have been introduced (Salomons & Förstner 1984; Müller 1979; Hakanson 1980).

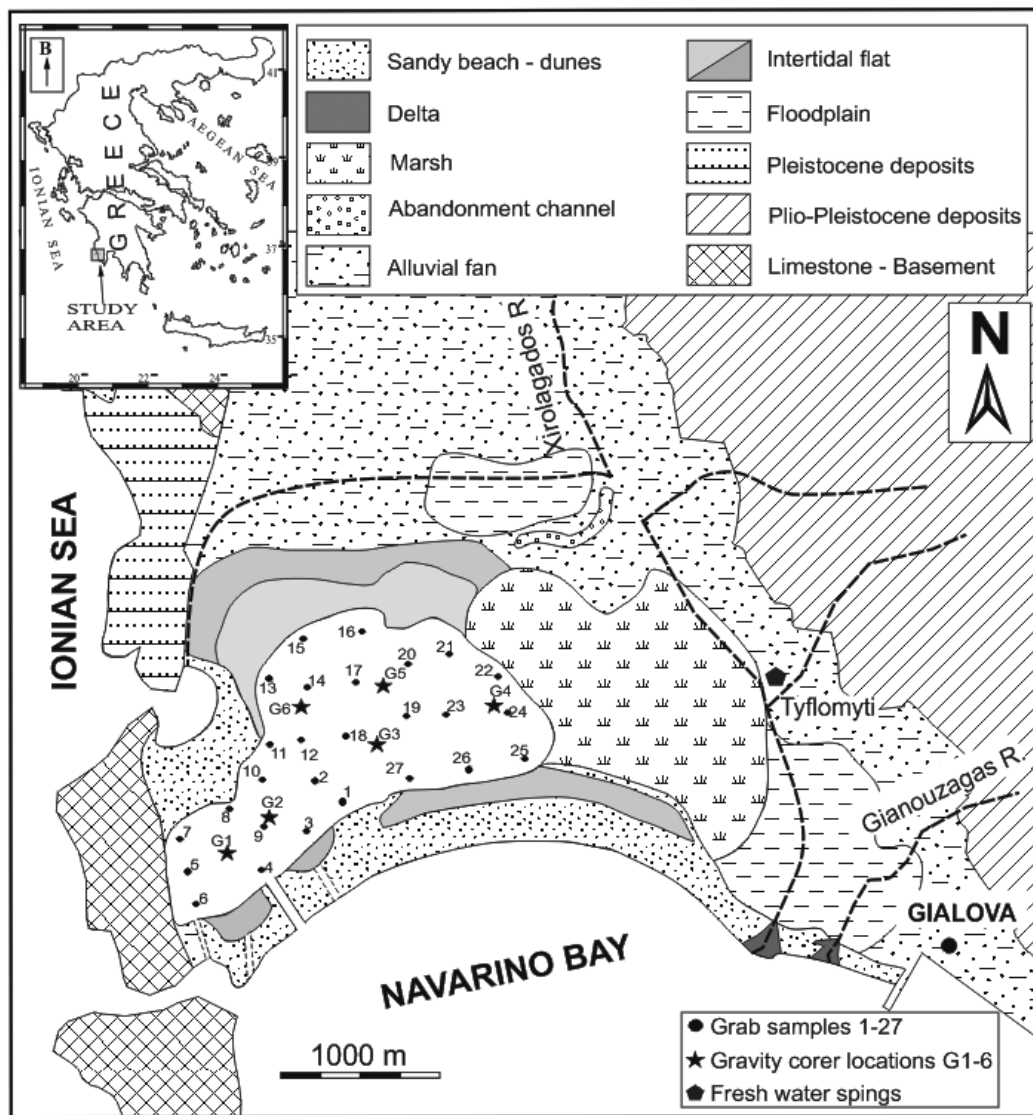


Figure 1 Location map of the study lagoon (inset) and geological map of the area around the Gialova Lagoon, showing locations of the grab and gravity corer samples.

This paper presents the spatial distribution of sedimentological characteristics of bottom lagoon sediments from Gialova Lagoon (southwestern Peloponnese, Greece), such as grain size and moment measures, together with geochemical environmental indices such as the enrichment factor, the contamination factor and the index of geo-accumulation (Abraham & Parker 2008); in order to delineate concentration enrichments or depletions of various metals. Moreover, micro (ostracods, foraminifera, charophytes) and macro (molluscs) subfossil assemblages were used to identify the effect of human activities on the faunal communities and their respective environments.

1. Regional setting

1.1. Geographical and geological setting

The Gialova Lagoon is located in the southwestern end of Peloponnese, and particularly on the northern coast of Navarino Bay (Fig. 1). It is one of the most important ecological areas in Greece and has international significance, as it is listed in the Natura 2000 European Community Network (code GR2550003) as a Special Protected Area (SPA) and Site of Community Importance (SCI) (European Union Council Directive 92/43/EEC on the conservation of natural habitats and of wild fauna and flora). It is a very shallow lagoon with maximum

and mean depths of 1.0 m and 0.5 m respectively, whereas the greatest depths occur in the central parts of the lagoon.

According to the Institute of Geology and Mineral Exploration of Greece (I.G.M.E. 1980) and based on geological mapping, the area belongs to the Gavrovo geotectonic zone (Aubouin 1959). The catchment area and the surrounding land area of the Gialova Lagoon consists of Holocene alluvial deposits and sand dunes, Plio-Pleistocene deposits of conglomerates, marls and fine grained sandstones, and Eocene to Oligocene flysch deposits; whilst the basement of the catchment area consist of Upper Cretaceous to Eocene limestone (Fig. 1). According to the Institute of Geology and Mineral Exploration of Greece (I.G.M.E. 1980) and based on geological mapping, the area belongs to the Gavrovo geotectonic zone.

1.2. Recent evolution of the lagoon

Attempts to drain the lagoon were made in the 1950s. Before 1950, the Xirolagados River flowed into the northern lagoon margin (Fig. 2a). The river fed the lagoon with fine sediments, such as fine sand, silt and clay. The source area of these sediments was the Plio-Pleistocene deposits from the mountainous course of the Xirolagados River and the Holocene sediments from the river lowland area. The Tiflomyti springs, which rose in the eastern lagoon margin, poured fresh water



Figure 2 Changes in the lagoon hydromorphology during the last 70 years: (a) before 1950; (b) after attempts at lagoon drainage; (c) after the 1998 water enrichment works. The Tyflomiti springs are shown by blue dots and artificial sluices by blue arrows. (Personal communication with Hellenic Ornithological Society.)

into the lagoon and the marshes of the Gianouzagas River to the east (Fig. 2a).

During the 1950s, the Xirolagados River was diverted to Voidokilia Bay, located at the northwest side of the lagoon, and the flow of the Tiflomity spring waters was also diverted through an artificial canal in Navarino Bay (Fig. 2b). These works reduced the lagoon area from 7.5 km² to 2.5 km² as water level decreased, reclaiming dry land for farming. In the middle 1980s, an artificial inlet was constructed in the lagoon barrier and, via this inlet, the lagoon was linked with Navarino Bay. Moreover, in the internal part of the lagoon levees were constructed, in order to isolate the fresh and saline waters. The consequences of the drainage works, as well as of the artificial interventions, were the environmental degradation of the lagoon and the initiation of several oxygen depletion and dystrophic events (Koutsoubas *et al.* 2000a). In October 1993, the local ecosystem suffered the impact from an oil spill in the Gulf of Navarino, causing short term, but reversible deterioration (Koutsoubas *et al.* 2000a). A funded ecological study started in the Gialova Lagoon in order to identify the repercussions of the oil spill in the lagoonal environment. The study took place during the years 1995–1996 on a seasonal basis, and provided the first coherent data on the macrobenthic (Koutsoubas *et al.* 2000a) and meiofaunal (McArthur *et al.* 2000) communities of the lagoon, the molluscan (Koutsoubas *et al.* 2000b) and the annelid (Arvanitidis *et al.* 1999) fauna, as well as on the annual cycles of the nutrients and the phytoplankton in the lagoon (Petihakis *et al.* 1999).

Following the suggestions of the management plan of that study, and in order to solve the problem of extreme disturbance due to dystrophic conditions that occurred every autumn in the lagoon, causing significant stress to the ecosystem (Koutsoubas *et al.* 2000a), two canals were opened after 1998, bringing fresh water from the Tiflomity springs and Xirolagados River into the lagoon (Fig. 2c), improving its environmental and ecological status (Chatzigeorgiou *et al.* 2011). However, a second ecological study that followed during 1998–1999 showed changes in the composition of the macrobenthic communities caused by dystrophic events that had continued to occur, possibly also caused by the inflow of fresh water, leading to their degradation (Chatzigeorgiou *et al.* 2011).

2. Material and methods

2.1. Sediment collection and grain size analysis

In order to estimate the spatial distribution of sediment characteristics, 27 surficial samples were collected from the bottom of the lagoon, using a grab; in addition, 23 samples were collected from six locations with the use of a gravity corer (Fig. 1). The grab sampler enables the top 10 cm of sediments

to be sampled, whereas with the gravity corer, sediments from a maximum depth of 45 cm can be sampled. Positioning was achieved with a hand-held GPS (accuracy 3 m). All samples were analysed for their grain size distribution. Material coarser than 4 Φ was dry sieved, whilst fine grained material >4 Φ was analysed using the pipette method, and finally grain size distribution was calculated. Grain size statistical parameters such as mean, sorting, skewness and kurtosis were calculated and sediment classification was made according to Folk (1974) nomenclature. Colours were identified using a Minolta CM-2002 hand-held spectrophotometer based on the Munsell colour chart.

The spatial distribution of sediment characteristics is presented using contour maps which were calculated by Surfer 9 software. The interpretation was based on the kriging technique, which is a statistical method that has been used in studies of geology and marine geology to estimate the distribution of variables in patches of marine sediments.

2.2. Micro- and macrofossil assemblages

Twenty three samples of approximately 15 g of dried sediment were collected from the six cores, namely G1–G6, for macro- and microfossil analysis. Four samples were collected from each core at 0 cm, 10 cm, 25 cm and 38–44 cm, except from core G4 where the fourth sample was not collected as the core was shorter than 30 cm. The specimens were washed through 0.5 mm and 0.063 mm mesh sieves using tap water. Macrofaunal specimens were mainly collected from the coarser sediments, whereas molluscs, ostracods and foraminifera were collected from the finer sediments. The collected specimens were sorted and determined, when possible, to species level, and taxonomic information was checked and updated following the World Register of Marine Species (WoRMS; Appeltans *et al.* 2012). The abundances of the determined taxa per sample were normalised to percentages, and the respective dataset was statistically analysed using correspondence analysis. Cluster analysis based on the Jaccard similarity index (absence–presence) (Jaccard 1912) was also applied to the respective data in order to identify possible sample groups.

2.3. Geochemistry

Approximately 1g \pm 0.1mg of dried sample was digested with 2.35 mL of 65 % HNO₃ and 7 mL of HCl (aqua regia) in a microwave digestion system (Berghof speedwave MWS-3⁺). The temperature digestion protocol was as follows: at 145°C for five minutes; at 170°C for ten minutes; and at 170°C for 15 minutes. The resulting solutions were cooled for 30 minutes and diluted to 10 mL with distilled water. The clear solutions were analysed on a Thermo Scientific iCAP 6000 ICP-OES. The operating conditions were as follows: nebuliser gas flow rates – 0.5 l/min; auxiliary gas flow – 0.5 l/min; plasma gas

Table 1 Surface sediment coordinates, grain size analysis results, total organic carbon (TOC) content and moment measures (Mz, σ , Sk and K_G).

| SAMPLES | COORDINATES | | SAND % | SILT % | CLAY % | TOC % | Mz | σ | Sk | K_G |
|---------|------------------------|------------------------|-----------|-----------|-----------|----------|------|----------|-------|-------|
| | N | E | | | | | | | | |
| 1 | 36.95964 ⁰ | 21.669824 ⁰ | 41.88 | 26.73 | 31.38 | 3.00 | 5.58 | 4.61 | 0.04 | 0.81 |
| 2 | 36.960246 ⁰ | 21.667947 ⁰ | 31.44 | 42.20 | 26.36 | 3.32 | 5.28 | 4.51 | -0.13 | 1.11 |
| 3 | 36.957588 ⁰ | 21.667333 ⁰ | 83.04 | 7.54 | 9.42 | 1.45 | 2.15 | 3.05 | 0.17 | 2.56 |
| 4 | 36.955475 ⁰ | 21.665287 ⁰ | 56.21 | 28.08 | 15.71 | 2.20 | 4.75 | 3.13 | 0.47 | 1.89 |
| 5 | 36.955712 ⁰ | 21.660545 ⁰ | 37.09 | 35.57 | 27.33 | 4.45 | 5.30 | 4.83 | -0.11 | 0.98 |
| 6 | 36.953863 ⁰ | 21.661088 ⁰ | 51.13 | 20.06 | 28.81 | 2.49 | 4.83 | 4.88 | 0.22 | 0.89 |
| 7 | 36.958162 ⁰ | 21.660295 ⁰ | 31.05 | 35.13 | 33.83 | 3.65 | 5.87 | 5.11 | -0.16 | 0.98 |
| 8 | 36.958006 ⁰ | 21.665141 ⁰ | 35.98 | 28.42 | 35.60 | 3.85 | 6.27 | 4.63 | -0.08 | 0.83 |
| 9 | 36.957588 ⁰ | 21.667333 ⁰ | 31.61 | 30.74 | 37.65 | 4.16 | 6.23 | 4.87 | -0.13 | 0.90 |
| 10 | 36.960592 ⁰ | 21.66503 ⁰ | 35.82 | 34.31 | 29.87 | 3.55 | 5.57 | 4.88 | -0.13 | 0.87 |
| 11 | 36.962761 ⁰ | 21.665219 ⁰ | 41.57 | 31.13 | 27.30 | 2.90 | 5.23 | 4.74 | -0.04 | 0.87 |
| 12 | 36.963033 ⁰ | 21.667256 ⁰ | 40.26 | 27.14 | 32.60 | 2.70 | 5.55 | 5.02 | -0.08 | 0.82 |
| 13 | 36.966163 ⁰ | 21.665644 ⁰ | 35.32 | 36.85 | 27.83 | 3.60 | 5.32 | 4.73 | -0.06 | 1.00 |
| 14 | 36.965748 ⁰ | 21.667665 ⁰ | 28.92 | 38.40 | 32.68 | 3.42 | 6.63 | 5.16 | 0.01 | 1.01 |
| 15 | 36.968524 ⁰ | 21.667662 ⁰ | 27.00 | 41.15 | 31.85 | 3.40 | 5.98 | 4.52 | -0.07 | 1.05 |
| 16 | 36.968751 ⁰ | 21.670833 ⁰ | 28.93 | 38.40 | 32.67 | 2.26 | 6.07 | 4.73 | -0.05 | 0.94 |
| 17 | 36.966172 ⁰ | 21.67053 ⁰ | 44.89 | 30.10 | 25.01 | 3.40 | 4.70 | 5.00 | -0.08 | 0.96 |
| 18 | 36.963128 ⁰ | 21.670045 ⁰ | 20.16 | 46.84 | 33.00 | 3.72 | 6.55 | 3.90 | 0.07 | 1.36 |
| 19 | 36.963492 ⁰ | 21.673185 ⁰ | 36.71 | 31.15 | 32.15 | 3.10 | 5.90 | 4.78 | -0.09 | 0.89 |
| 20 | 36.967006 ⁰ | 21.673475 ⁰ | 30.11 | 34.50 | 35.38 | 3.88 | 5.77 | 4.98 | -0.04 | 0.92 |
| 21 | 36.967439 ⁰ | 21.675756 ⁰ | 33.70 | 40.81 | 25.49 | 3.60 | 5.57 | 4.82 | -0.03 | 0.92 |
| 22 | 36.965719 ⁰ | 21.678571 ⁰ | 33.11 | 34.99 | 31.90 | 5.90 | 5.27 | 4.61 | -0.34 | 0.93 |
| 23 | 36.96348 ⁰ | 21.675765 ⁰ | 44.05 | 24.92 | 31.03 | 2.65 | 5.25 | 4.83 | -0.12 | 0.83 |
| 24 | 36.963734 ⁰ | 21.678936 ⁰ | 46.66 | 33.11 | 20.23 | 4.15 | 4.10 | 4.89 | -0.07 | 0.86 |
| 25 | 36.960931 ⁰ | 21.680692 ⁰ | 13.55 | 45.84 | 40.61 | 5.90 | 7.48 | 3.64 | -0.11 | 1.95 |
| 26 | 36.960529 ⁰ | 21.676417 ⁰ | 65.00 | 13.15 | 21.85 | 1.74 | 4.00 | 4.61 | 0.46 | 0.85 |
| 27 | 36.960107 ⁰ | 21.673448 ⁰ | 35.70 | 29.91 | 34.39 | 3.30 | 5.88 | 4.55 | -0.16 | 1.00 |

flow – 15 l/min; pPump rate – 45 rpm; ICP RF power – 1100 W. Aliquots of an ICP multi-element standard solution (100 mg/L Merck) containing the analysed elements, was used in the preparation of calibration solutions. Working standard solutions were prepared by dilution of the stock standard solutions to the desired concentration in 1% HNO₃. The ranges of the calibration curves (six points) were selected to match the expected concentrations for all the elements of the samples analysed by ICP-OES. The correlation coefficient R^2 obtained for all cases was 0.9999. The detection limits (LOD) were calculated as the concentrations of an element given by the standard deviation of a series of ten consecutive measurements of blank solutions. Soil samples from BIPEA's proficiency testing scheme A15 were used to ascertain the accuracy of the measurements. The total organic carbon (TOC) content was determined by the titration method (Gaudette *et al.* 1974).

In order to assess background enrichments or depletions of metals in the Gialova Lagoon sediments, we employed three geochemical indices:

- i. The enrichment factor, EF, (Reimann & de Caritat 2005),

$$EF = (C_{\text{sample}} / C_{\text{Al}}) / (C_{\text{standard}} / C_{\text{Al standard}})$$

where C_{sample} is the concentration of the element and $C_{\text{Al sample}}$ is the concentration of Al in the sample. The Al normalisation is utilised because concentrations of most elements show correlation with Al₂O₃. The selected reference sample is usually an average crust or a local background sample (Blaser *et al.* 2000; Liu *et al.* 2005). Average shale values are used as the standard in our study (Turekian & Wedepohl 1961).

- ii. The contamination factor, C_f^i , (Håkanson 1980; Abraham & Parker 2008),

$$C_f^i = \overline{C_{0-1}^i} / C_n^i$$

Where $\overline{C_{0-1}^i}$ is the mean content of the substance i from at least five sample sites and C_n^i is the pre-industrial reference level for the substance. The pre-industrial reference level determined from various European and American lakes is given by Håkanson (1980).

- iii. The index of geo-accumulation, I_{geo} , (Müller 1979; Abraham & Parker 2008).

$$I_{\text{geo}} = \log_2 [C_n / (1.5 B_n)],$$

where C_n is the measured concentration in the sediment for the metal n , B_n is the background value for the metal n and the factor 1.5 is used because of possible variations of the background data due to lithological variations. However, several researchers (Subramanian & Mohanachandran 1990; Sahu & Bhosale 1991) have used the previous expression using regional backgrounds and on the less than 63–65 μm sediment fraction. In this study, I_{geo} has been calculated using global average shale data from Turekian & Wedepohl (1961) (Rubio *et al.* 2000; Harikumar & Jisha 2010), because they refer to the bulk concentrations and are naturally very rich in fine grain sizes.

In addition, we did establish a local baseline value because metals concentrations are depleted towards the recent sediments. Since pollution effects may extend to a considerable variation in depth, the selection of the low concentration samples for baseline averaging is best done by inspection of the metal trends in the lower core.

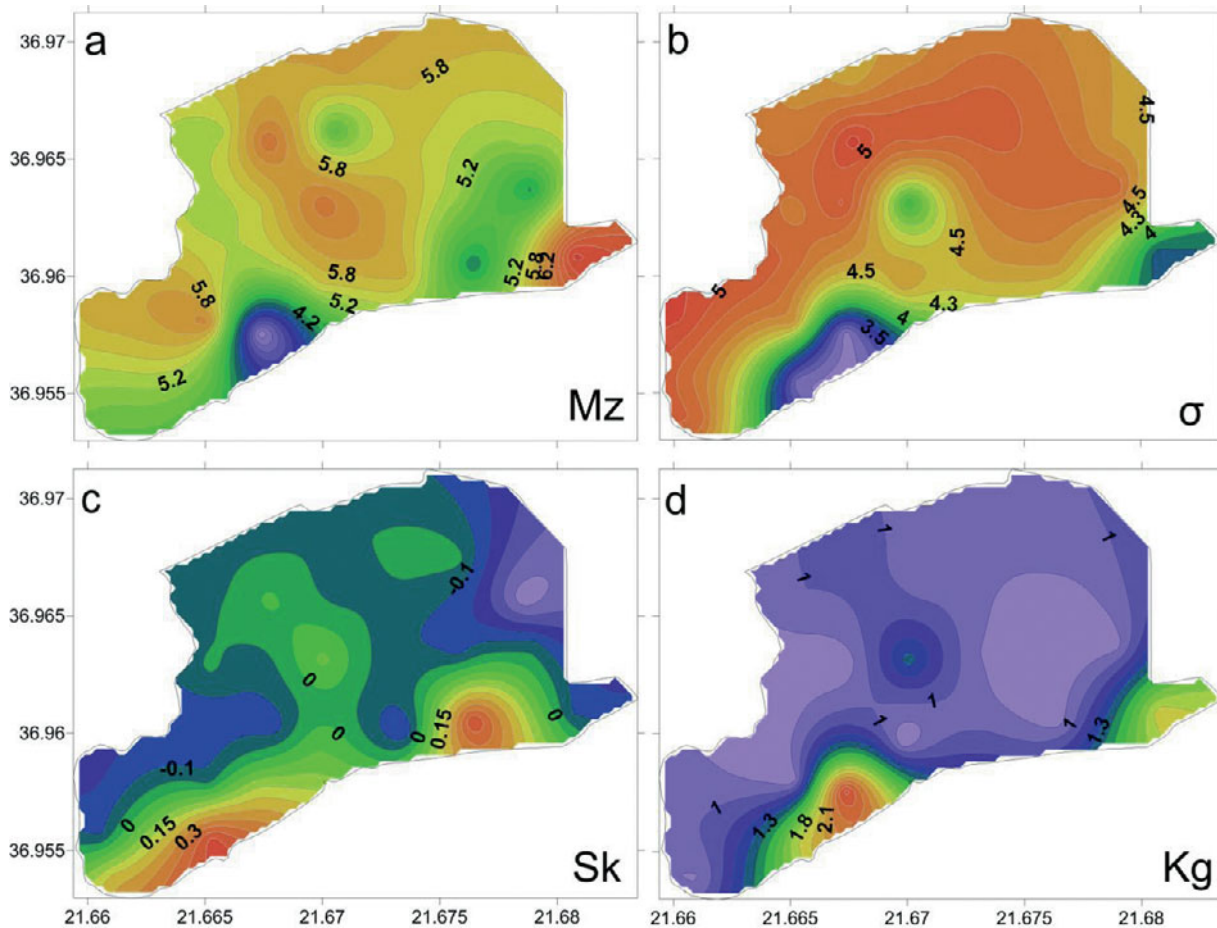


Figure 3 Distributions of the moment measures: (a) mean size (Mz); (b) sorting (σ); (c) skewness (Sk); (d) kurtosis (Kg).

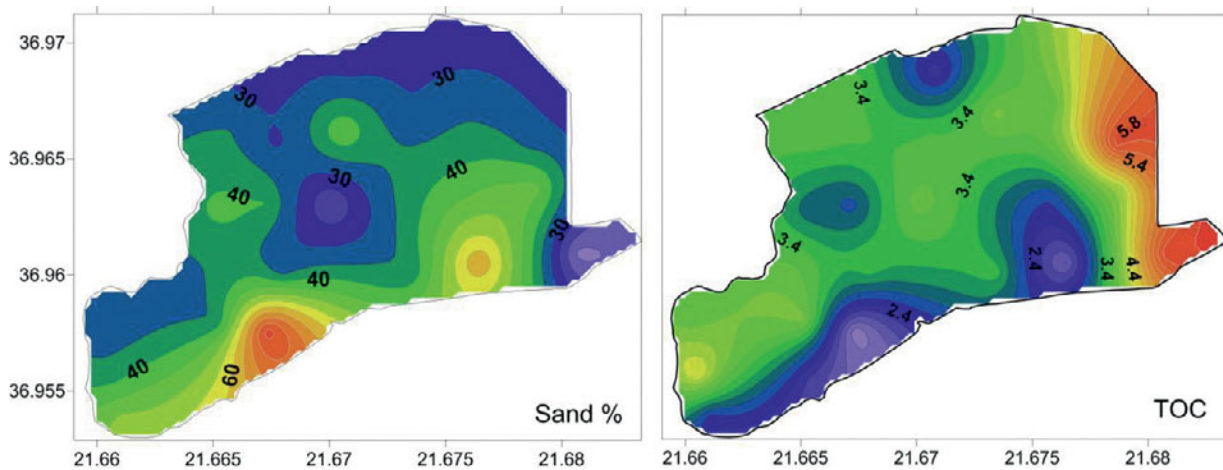


Figure 4 (a) Spatial distribution of sand in the Gialova Lagoon. (b) total organic carbon (TOC).

3. Results

3.1. Lagoon surficial grain size distribution

Based on the grain size analyses of the 27 bottom sediment samples, the mean size (Mz) ranged from 2.15Φ to 7.48Φ , with an average value of 5.45Φ , indicating that the lagoon's sediments are composed mainly of sandy mud and few samples of muddy sand (Fig. 3; Table 1). The spatial distribution of mean size indicates that the finest material is observed at the southeastern lagoon margin ($Mz > 7\Phi$), whilst coarser clastic

sediments are observed near the artificial inlet very close to the barrier island (Figs 3, 4).

The analysis of the moment measures showed that the sediments are characterised by a very poor to an extremely poor sorting (σ), and range from 3.05Φ to 5.16Φ , with an average value of 4.59Φ (Fig. 3b; Table 1). The skewness of the analysed sediments ranged from strongly coarse skewed to strongly fine skewed ($Sk = -0.34 \Phi$ to 0.47Φ), with an average value of -0.02Φ (Fig. 3c; Table 1). In total, 50 % of the samples have near symmetrical skewness, 30 % are fine-skewed, 4 % are

Table 2 Result of major (wt. %) and trace elements (ppm), total organic carbon, grain size analysis of the gravity cores samples G1-6 and their colors according to Munsell color chart.

| DEPTH (m) | Al ₂ O ₃ % | CaO % | Fe ₂ O ₃ % | K ₂ O % | MgO % | MnO % | Na ₂ O % | P ₂ O ₅ % | Cd ppm | Cr ppm | Cu ppm | Ni ppm | Pb ppm | V ppm | Zn ppm | MUD % | SAND % | GRAVEL % | TOC % | COLOR |
|---------------------|----------------------------------|-------|----------------------------------|--------------------|-------|-------------------|---------------------|---------------------------------|--------|--------|--------|--------|--------|--------|--------|-------|--------|----------|-------|-------------------------------|
| GRAVITY CORER G-1 | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 6.73 | 12.13 | 6.98 | 1.62 | 4.47 | 0.21 | 7.92 | 0.19 | 0.76 | 116.72 | 56.59 | 145.75 | 62.31 | 72.63 | 101.13 | 71.74 | 24.51 | 3.75 | 3.53 | 10YR 4/2 - dark grayish brown |
| 0.10 | 7.88 | 17.14 | 6.93 | 1.29 | 3.38 | 0.15 | 3.06 | 0.14 | 1.05 | 119.32 | 64.11 | 181.9 | 36.8 | 69.13 | 101.72 | 47.35 | 51.88 | 0.76 | 4.23 | 10YR 44/1 - dark gray |
| 0.42 | 9.31 | 18.81 | 7.93 | 1.85 | 3.47 | 0.14 | 2.53 | 0.1 | 0.93 | 170.17 | 70.08 | 207.72 | 30.28 | 101.11 | 122.02 | 58.51 | 37.18 | 4.31 | 2.69 | 10B 3/1 - v. dark bluish gray |
| GRAVITY CORER G-2 | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 6.78 | 15.37 | 6.89 | 1.6 | 4.3 | 0.25 | 6.97 | 0.17 | 0.74 | 115.92 | 56.88 | 142.95 | 58.49 | 64.66 | 92.99 | 49.94 | 42.35 | 7.71 | 3.89 | 10Y 4/1 - dark greenish gray |
| 0.10 | 7.78 | 11.72 | 7.22 | 1.27 | 3.39 | 0.17 | 2.98 | 0.14 | 0.91 | 120.14 | 58.1 | 191.36 | 33.4 | 68.87 | 103.2 | 41.97 | 55.04 | 2.99 | 4.60 | 5B 3/1 - v. dark bluish gray |
| 0.25 | 7.33 | 25.28 | 7.15 | 1.52 | 3.59 | 0.21 | 3.65 | 0.11 | 0.9 | 138.33 | 61.52 | 170.6 | 25.7 | 78.28 | 109.75 | 45.69 | 50.91 | 3.40 | 3.31 | 10Y 4/1 - dark greenish gray |
| 0.40 | 8.14 | 15.25 | 7.97 | 1.53 | 3.49 | 0.18 | 2.7 | 0.14 | 0.92 | 139.23 | 63.8 | 210.87 | 24.19 | 82.97 | 120.04 | 53.99 | 37.28 | 8.73 | 2.76 | 10B 4/1 - dark bluish gray |
| GRAVITY CORER G-3 | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 1.91 | 47.88 | 2.01 | 0.46 | 2.08 | 0.1 | 2.42 | 0.11 | 0.24 | 29.07 | 31.62 | 40.54 | 17.12 | 19.09 | 36.18 | 19.07 | 53.93 | 27.00 | 1.56 | 5B 3/1 - v. dark bluish gray |
| 0.10 | 5.19 | 19.28 | 5.35 | 1.03 | 3.85 | 0.25 | 4.7 | 0.21 | 0.6 | 79.07 | 44.22 | 120.62 | 44.8 | 45.69 | 77.77 | 41.38 | 50.44 | 8.18 | 4.37 | 5B 4/1 - dark bluish gray |
| 0.25 | 5.69 | 22.34 | 6.24 | 1.03 | 3.61 | 0.22 | 3.96 | 0.21 | 0.86 | 91.34 | 54.64 | 143.27 | 56.61 | 51.93 | 84.87 | 45.31 | 36.01 | 18.68 | 3.20 | 5B 3/1 - v. dark bluish gray |
| 0.44 | 7.47 | 16.59 | 6.76 | 1.51 | 4.6 | 0.47 | 7.19 | 0.27 | 0.78 | 112.18 | 53.54 | 142.62 | 63.54 | 63.58 | 94.04 | 64.09 | 33.31 | 2.60 | 4.47 | N 2.5 - black |
| GRAVITY CORER G-4 | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 4.3 | 31.05 | 4.29 | 0.75 | 3.26 | 0.21 | 3.43 | 0.17 | 0.48 | 60.41 | 38.03 | 92.44 | 34.31 | 36.83 | 59.26 | 29.74 | 50.52 | 19.74 | 2.65 | 5B 4/1 - dark bluish gray |
| 0.10 | 9.14 | 13.51 | 8.04 | 1.29 | 3.33 | 0.19 | 2.5 | 0.14 | 0.98 | 138.72 | 65.98 | 244.91 | 29.46 | 78.67 | 115.85 | 66.39 | 32.68 | 0.93 | 2.91 | 5B 3/1 - v. dark bluish gray |
| 0.20 | 9.08 | 18.49 | 7.81 | 1.34 | 3.37 | 0.21 | 2.37 | 0.14 | 0.89 | 133.23 | 58.74 | 201.65 | 29.18 | 77.49 | 107.74 | 50.09 | 47.57 | 2.34 | 2.97 | 5B 3/1 - v. dark bluish gray |
| GRAVITY CORER G-5 | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 4.94 | 25.66 | 4.83 | 0.85 | 3.27 | 0.19 | 2.89 | 0.21 | 0.58 | 69.65 | 39.93 | 102.79 | 40.85 | 38.11 | 66.39 | 41.17 | 51.21 | 7.62 | 2.63 | 5B 2.5/1 - bluish black |
| 0.10 | 9.11 | 9.96 | 9.27 | 1.4 | 3.47 | 0.27 | 2.79 | 0.13 | 1.07 | 175.75 | 101.09 | 271.49 | 26.99 | 101.6 | 133.15 | 75.92 | 23.65 | 0.43 | 2.05 | 5B 4/1 - dark bluish gray |
| 0.25 | 8.84 | 18.08 | 7.71 | 1.52 | 3.76 | 0.17 | 2.28 | 0.13 | 0.86 | 143.29 | 67.81 | 208.74 | 33.53 | 81.8 | 122.02 | 62.70 | 30.62 | 6.68 | 2.53 | 5B 3/1 - v. dark bluish gray |
| 0.42 | 8.36 | 17.74 | 8.31 | 1.42 | 3.02 | 0.16 | 1 | 0.13 | 0.97 | 130.51 | 62.43 | 218.99 | 22.61 | 81.29 | 113.88 | 80.36 | 13.41 | 6.22 | 2.17 | 5B 3/1 - v. dark bluish gray |
| GRAVITY CORER G-6 | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 5.86 | 26.02 | 6.11 | 1.36 | 4.2 | 0.22 | 6.49 | 0.13 | 0.72 | 102.12 | 51.23 | 123.58 | 52.07 | 65.67 | 86.22 | 35.41 | 38.94 | 25.65 | 3.52 | 5B 4/1 - dark bluish gray |
| 0.10 | 7.26 | 14.72 | 6.76 | 1.19 | 3.28 | 0.14 | 1.34 | 0.14 | 0.92 | 112.96 | 55.52 | 218.03 | 28.71 | 68.6 | 105.96 | 47.60 | 38.08 | 14.33 | 3.45 | 10BG 4/1 - dark greenish gray |
| 0.25 | 8.56 | 15.64 | 7.65 | 1.29 | 3.05 | 0.16 | 1.19 | 0.14 | 0.86 | 125.63 | 54.06 | 223.71 | 25.25 | 72.57 | 109.26 | 86.18 | 12.31 | 1.51 | 2.14 | 5BG 4/1 - dark greenish gray |
| 0.38 | 6.98 | 25.32 | 7.31 | 1.78 | 2.91 | 0.11 | 4.67 | 0.15 | 0.81 | 176.7 | 55.36 | 181.93 | 22.23 | 102.7 | 102.79 | 76.66 | 19.52 | 3.82 | 2.37 | 5BG 4/1 - dark greenish gray |
| Average | 7.12 | 19.91 | 6.80 | 1.31 | 3.51 | 0.20 | 3.59 | 0.15 | 0.81 | 118.20 | 57.51 | 172.11 | 36.29 | 69.24 | 98.47 | 8.06 | 37.79 | 54.15 | 3.09 | |
| Min | 1.91 | 9.96 | 2.01 | 0.46 | 2.08 | 0.10 | 1.00 | 0.10 | 0.24 | 29.07 | 31.62 | 40.54 | 17.12 | 19.09 | 36.18 | 0.43 | 12.31 | 19.07 | 1.56 | |
| Max | 9.31 | 47.88 | 9.27 | 1.85 | 4.60 | 0.47 | 7.92 | 0.27 | 1.07 | 176.70 | 101.09 | 271.49 | 63.54 | 102.70 | 133.15 | 27.00 | 55.04 | 86.18 | 4.60 | |
| SD | 1.86 | 8.24 | 1.57 | 0.33 | 0.56 | 0.07 | 1.97 | 0.04 | 0.19 | 36.51 | 13.77 | 55.38 | 13.97 | 21.15 | 22.91 | 7.95 | 13.10 | 17.29 | 0.85 | |
| UCC ¹ | 15.2 | 4.2 | 5.00 ^a | 3.4 | 2.2 | 0.08 ^b | 3.9 | – | 98 | 35 | 25 | 20 | 20 | 60 | 71 | – | – | – | – | |
| Shales ² | 15.11 | 3.09 | 6.75 | 3.2 | 2.49 | 0.11 | 1.29 | 0.16 | 0.3 | 90 | 45 | 68 | 20 | 130 | 95 | – | – | – | – | |

1. UCC: Upper Continental Crust values by Taylor and McLennan (1985).

2. Shales: Values reported after Turekian and Wedepohl (1961). The Al₂O₃, CaO, Fe₂O₃, K₂O, MgO, MnO, Na₂O and P₂O₅ concentrations calculated using the appropriate molar ratio.

a: Calculated after the results given by Taylor and McLennan, (1985) for Fe 3.50%.

b: Calculated after the results given by Taylor and McLennan, (1985) for Mn 600 ppm.

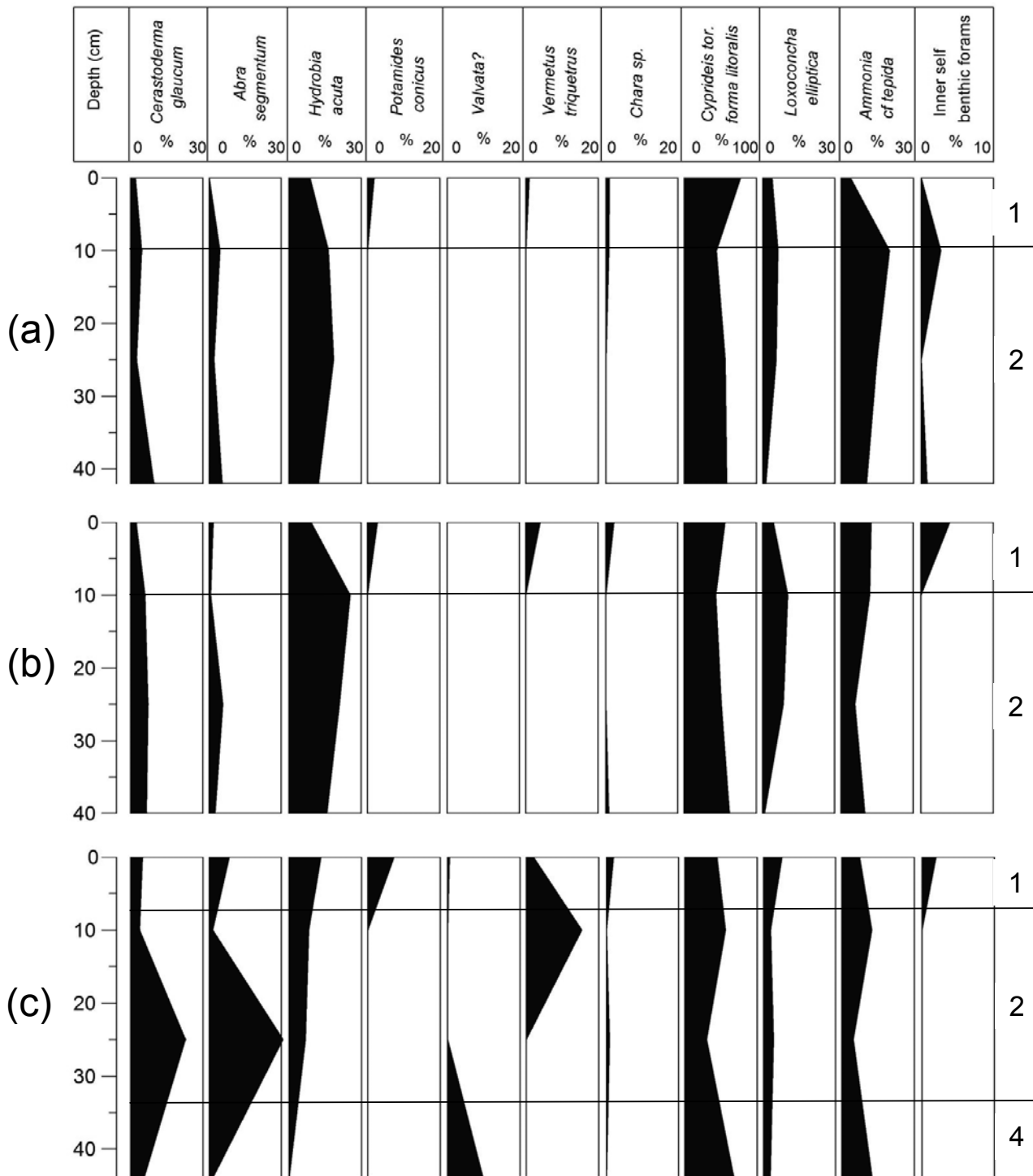


Figure 5 Macro- and micro-faunal abundance plots from gravity cores: (a) G1; (b) G2; (c) G3. Based on Table 3 and the boundaries of the different environmental facies.

strongly fine-skewed and 8 % are strongly coarse-skewed. The calculation of kurtosis displays a fluctuation from very platykurtic to very leptokurtic values ($KG = 0.61-2.56 \Phi$), with an average value of 1.07Φ (Fig. 3d; Table 1). In total, 37 % of the samples show platykurtic values, 45 % mesokurtic values, 8 % leptokurtic values and 11 % very leptokurtic values. The mesokurtic curves were observed only in the northern part of the lagoon, whilst on the basis of its spatial distribution, a variety of kurtosis values were observed.

The vertical grain size distribution of the 22 samples, collected from the six gravity cores, indicate similar characteristics in the sediments from all the cores and thus all the sediments can be classified as sandy mud (Table 2) (Folk 1974).

The presented gravel portion in the gravity corer samples is of biogenic origin, consisting of shells and shell fragments, and for this reason was not included in the sediments' grain size characteristics.

3.2. Micro- and macrofossil assemblages

The study of the macro- and microfossil remains provided taxa that belong to six different phyla (charophytes, molluscs, arthropods, foraminifera, bryozoans and echinoderms). More specifically, two bivalve, five gastropod, four ostracod, ten foraminifera and one charophyte species were determined (Table 3). In all cores, the dominant taxa were *Cyprideis torosa*

Table 3 Abundances of the determined macro- and micro- faunal taxa in percentages extracted from the 23 samples that were collected from the six cores.

| Depth (m) | Bivalvia | | Gastropoda | | | | | | | | | | Ostracoda | | | | Foraminifera | | | | | | | | | | | | |
|--------------------------|-----------------------------|-----------------------|-----------------------|--------------------------|-----------------|----------------------------|-----------|-----------------|-------------------------------|-----------|------------------|---|-----------------------------|----------------------------|-------------------------|----------------|--------------------------|----------------------------|---------------------------------|----------------------------|----------------------------|----------------------|---------------------|-----------------------------|---------------------|----------------------|----------------------------|-----|---|
| | <i>Cerastoderma glaucum</i> | <i>Abra segmentum</i> | <i>Hydrobia acuta</i> | <i>Potamides conicus</i> | <i>Valvata?</i> | <i>Vermetus triquetrus</i> | Rissoidae | Gastropod indet | sea urchin spines (Echinoids) | Bryozoans | <i>Chara sp.</i> | <i>Cyprideis torosa forma litoralis</i> | <i>Loxocochoa elliptica</i> | <i>Heterocypris salina</i> | <i>Leptocythere sp.</i> | Ostracod indet | <i>Ammonia cf tepida</i> | <i>Haynesina germanica</i> | <i>Quinqueloculina seminula</i> | <i>Quinqueloculina sp.</i> | <i>Peneroplis planatus</i> | <i>Cibicides sp.</i> | <i>Bolivina sp.</i> | <i>Elphidium cf crispum</i> | <i>Bulimina sp.</i> | Benthic foram indet. | <i>Globigerinoides sp.</i> | | |
| Gravity Corer G-1 | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 1.9 | 0 | 8.7 | 1.9 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 78 | 3.85 | 0 | 0 | 0 | 3.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.10 | 4.5 | 4.5 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.9 | 45 | 6.36 | 0 | 0 | 0 | 20 | 0 | 0 | 0 | 0 | 0 | 0.9 | 0 | 0 | 1.8 | 0 | 0 | |
| 0.25 | 2.4 | 2 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 56 | 5.65 | 0 | 0 | 0 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.42 | 9.7 | 5.5 | 12 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0.4 | 0 | 59 | 1.26 | 0 | 0 | 0 | 11 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0.4 | 0.4 | 0 | 0 | 0 | |
| Gravity Corer G-2 | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 2.3 | 1.7 | 9.1 | 2.8 | 0 | 4 | 0 | 0 | 0 | 0 | 2.3 | 57 | 4.55 | 0 | 0 | 0 | 13 | 0 | 2.3 | 0 | 0.6 | 1.1 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.10 | 6 | 0.7 | 25 | 0 | 0 | 0 | 0 | 0 | 0.7 | 0 | 0 | 44 | 10.4 | 0 | 0 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.7 | |
| 0.25 | 7.2 | 5.8 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 51 | 8.7 | 0 | 0 | 0 | 5.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.40 | 6.6 | 2.5 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.8 | 63 | 0.83 | 0 | 0 | 0.8 | 9.9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Gravity Corer G-3 | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 4.9 | 8.3 | 13 | 7.3 | 0.5 | 2 | 0 | 0.5 | 0 | 0 | 2 | 44 | 7.8 | 0 | 0 | 0 | 7.3 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.10 | 3.6 | 1.5 | 8 | 0 | 0 | 15 | 0 | 0 | 0 | 0 | 0 | 56 | 2.92 | 0 | 0 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.25 | 23 | 31 | 6.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.8 | 30 | 4.24 | 0 | 0 | 0 | 4.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.44 | 5.6 | 1.4 | 0 | 0 | 9.7 | 0 | 0 | 0 | 0 | 0 | 0 | 68 | 2.78 | 0 | 0 | 0 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Gravity Corer G-4 | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 6.8 | 11 | 13 | 15 | 0.5 | 0.5 | 0 | 0 | 0 | 0 | 3.9 | 39 | 6.34 | 0 | 0 | 0 | 3.9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.10 | 1.7 | 12 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.5 | 53 | 10 | 0 | 0 | 0 | 1.7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.20 | 7.8 | 3.5 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.9 | 61 | 8.7 | 0 | 0 | 0 | 6.1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Gravity Corer G-5 | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 4 | 6 | 9.3 | 1.3 | 0 | 0.7 | 0 | 0 | 0 | 0 | 3.3 | 50 | 13.3 | 0 | 0 | 0 | 6.7 | 0 | 5.3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.10 | 3.1 | 1.6 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 17 | 57 | 4.72 | 0 | 0 | 0.8 | 6.3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.25 | 18 | 11 | 6.8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5.4 | 54 | 1.46 | 0 | 1 | 0 | 2.9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.42 | 13 | 3.6 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 58 | 1.2 | 1.2 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Gravity Corer G-6 | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| 0.00 | 5.1 | 4.3 | 6.8 | 5.1 | 0 | 4.3 | 1.7 | 0 | 0.9 | 0 | 3.4 | 58 | 0 | 0 | 0 | 0 | 8.5 | 0 | 1.7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.10 | 9.2 | 3.8 | 18 | 0 | 0 | 0 | 0 | 0 | 0.8 | 0 | 0 | 65 | 0 | 0 | 0 | 0 | 3.1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.25 | 4.5 | 5.4 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.9 | 67 | 0 | 0 | 0 | 0 | 4.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 0.38 | 11 | 4.4 | 8.8 | 0 | 3.1 | 0 | 0 | 0 | 0 | 0 | 0.6 | 69 | 0 | 0 | 0 | 0 | 3.1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |

forma *litoralis*, *Hydrobia acuta* and *Ammonia cf tepida*. The vertical distribution of all taxa is described in detail per gravity core.

3.2.1. Gravity Corer G1. The dominant taxon in core G1 is *Cyprideis torosa* forma *litoralis*, accompanied in considerable percentages by *Hydrobia acuta* and *Ammonia cf tepida* (Fig. 5; Table 3). At the lowermost part of the core, *C. torosa* is the most abundant taxon, followed by *H. acuta* and *A. cf. tepida*. *Cerastoderma glaucum* and *Abra segmentum* are also present in fair numbers. Two inner shelf taxa, *Elphidium* and *Bulimina*, as well as the marine lagoonal and brackish taxon *Loxocochoa elliptica* (Zaibi et al. 2012), were also identified. The assemblage at 25 cm depth is almost the same, with the exception that *H. acuta*, *L. elliptica* and *A. cf. tepida* numbers increase, although the numbers of *C. glaucum* and *A. segmentum* become reduced and the inner shelf taxa disappear, showing less influence from the sea. At 10 cm depth, the taxa composition changes, characterised by a drop in *C. torosa* and an

increase of *H. acuta*, *L. elliptica* and *A. cf. tepida* and *C. glaucum* and *A. segmentum*; in addition, inner shelf benthic foraminifera (such as *Bolivina*) appear again in very low numbers, accompanied by the presence of a charophyte gyrogonite. At the top of the core, the numbers of the more resistant *C. torosa* rise significantly (78%), whilst *H. acuta*, *L. elliptica* and *A. cf. tepida* and *C. glaucum* and *A. segmentum* numbers drop. A few *Potamides conicus* and *Vermetus triquetrus* shells, and charophyte gyrogonites, are also present.

3.2.2. Gravity Corer G2. In this core, the overall distribution of the main determined taxa is similar to core G1 (Fig. 5; Table 3). The dominant taxon is also *C. torosa* forma *litoralis*, accompanied in significant numbers by *H. acuta* and *A. cf. tepida* (which are more abundant in the middle part of the core). *C. glaucum* and *A. segmentum* are also present, being more abundant towards the lower part of the core. Conversely, *L. elliptica* has a significant presence in the middle of the core, but its numbers decrease at the lower part of the

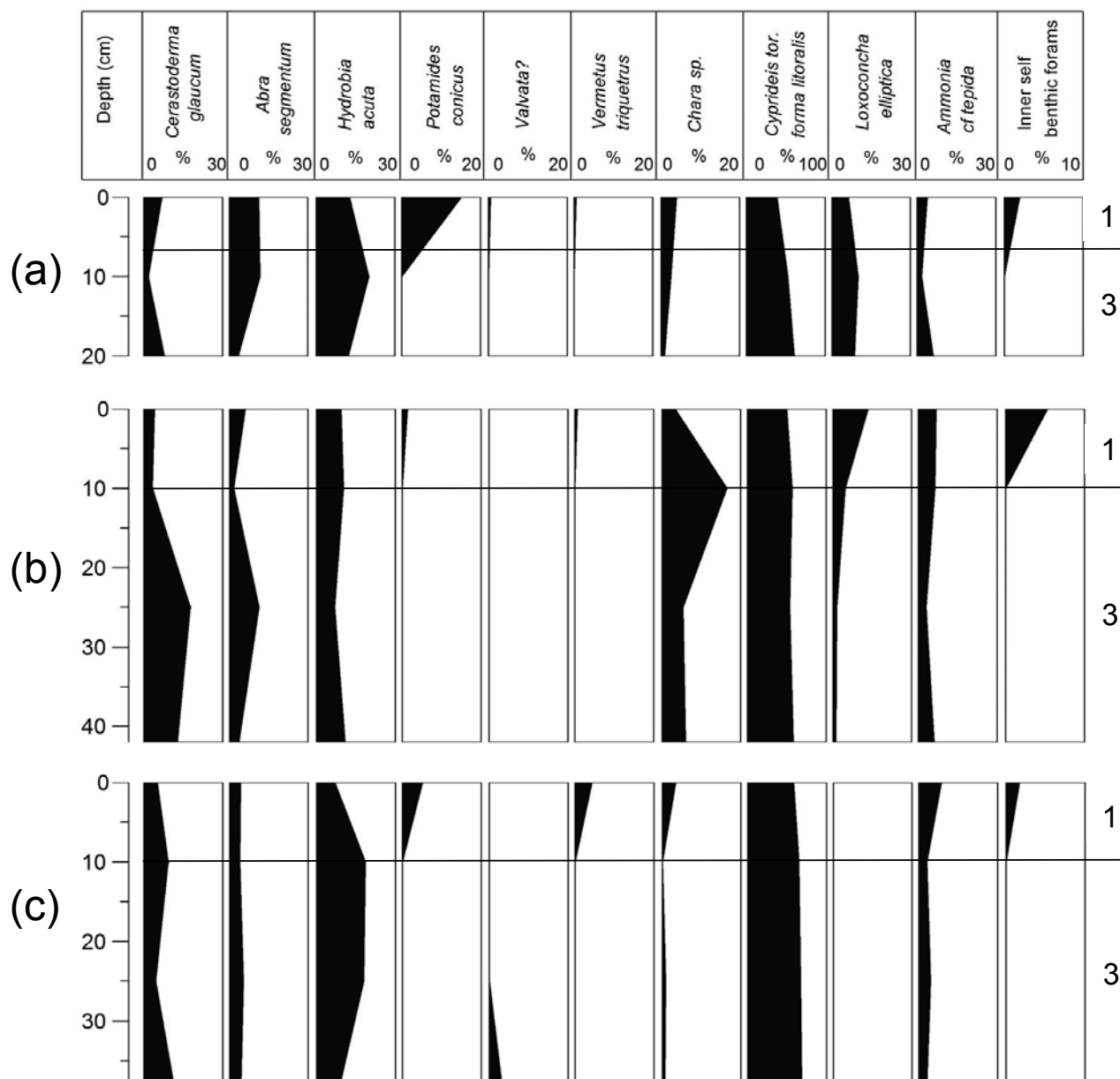


Figure 6 Macro- and micro-faunal abundance plots from gravity cores: (a) G4; (b) G5; (c) G6. Based on Table 3 and the boundaries of the different environmental facies.

core. Inner shelf foraminifera and the gastropod *V. triquetrus* are only found at the uppermost part of the core. Notably, a single *Globigerinoides* sp. test has been found at 10 cm. Finally, charophyte gyrogonites were found at the top and at the lowermost part of the core.

3.2.3. Gravity Corer G3. Despite the dominance of *C. torosa* forma *litoralis* and the contribution of *L. elliptica* and *A. cf tepida*, the composition of the recorded assemblages in core G3 presents a slightly different distribution as compared with cores G1 and G2 (Fig. 5; Table 3). Notably, although the assemblage at the lower part of the core consists of almost all the common taxa except *H. acuta*, the numbers of specimens of the respective taxa are relatively low. The total number of specimens that were collected from the whole processed sample is 72 and the few bivalve valves (*C. glaucum* and *A. segmentum*) were small, whereas the *C. torosa* valves were also small-sized, representing mainly juvenile individuals. *Valvata* shells and charophyte gyrogonites have also been recovered. At 25 cm, *C. glaucum* and *A. segmentum* are significantly abundant, being 23 % and 31 % respectively of the total assem-

blage. The 10 cm depth is characterised by the significant drop in *C. glaucum* and *A. segmentum* numbers and the marked presence of *V. triquetrus*. At the top of the core, inner shelf foraminifera, the gastropods *V. triquetrus* and *P. conicus*, and the presence of charophyte gyrogonites change the composition of the assemblage.

3.2.4. Gravity Corer G4. Once again, *C. torosa* forma *litoralis* dominates this short core. *H. acuta* and *A. segmentum* are major components of the assemblages and *L. elliptica* has a significant contribution (Fig. 6; Table 3). *Ammonia cf tepida* is present in low numbers all over the core, whereas charophyte gyrogonites increase gradually towards the top of the core, suggesting fresh water influence. At the top of the core, inner shelf foraminifera and significant numbers of the gastropod *P. conicus* are present.

3.2.5. Gravity Corer G5. *C. torosa* forma *litoralis* is the dominant taxon, followed by *H. acuta* and *Ammonia cf tepida* (Fig. 6; Table 3). The numbers of the collected specimens of these three taxa are relatively stable along the core. *C. glaucum* and *A. segmentum* numbers are considerably higher at the lower

Table 4 The calculated enrichment factors (EF) and Index of geo-accumulation (I_{geo}) for the analyzed samples.

| Depth (m) | Enrichment Factor (EF) | | | | | | | Index of geo-accumulation (I _{geo}) | | | | | | |
|------------------|------------------------|------|------|------|------|------|------|---|-------|-------|-------|-------|-------|-------|
| | Cd | Cr | Cu | Ni | Pb | V | Zn | Cd | Cr | Cu | Ni | Pb | V | Zn |
| GRAVITY CORE G-1 | | | | | | | | | | | | | | |
| 0.00 | 5.69 | 2.91 | 2.82 | 4.81 | 7.00 | 1.25 | 2.39 | 0.76 | -0.21 | -0.25 | 0.51 | 1.05 | -1.42 | -0.49 |
| 0.10 | 6.71 | 2.54 | 2.73 | 5.13 | 3.53 | 1.02 | 2.05 | 1.22 | -0.18 | -0.07 | 0.83 | 0.29 | -1.50 | -0.49 |
| 0.42 | 5.03 | 3.07 | 2.53 | 4.96 | 2.46 | 1.26 | 2.09 | 1.05 | 0.33 | 0.05 | 1.03 | 0.01 | -0.95 | -0.22 |
| GRAVITY CORE G-2 | | | | | | | | | | | | | | |
| 0.00 | 5.50 | 2.87 | 2.82 | 4.69 | 6.52 | 1.11 | 2.18 | 0.72 | -0.22 | -0.25 | 0.49 | 0.96 | -1.59 | -0.62 |
| 0.10 | 5.89 | 2.59 | 2.51 | 5.47 | 3.24 | 1.03 | 2.11 | 1.02 | -0.17 | -0.22 | 0.91 | 0.15 | -1.50 | -0.47 |
| 0.25 | 6.19 | 3.17 | 2.82 | 5.17 | 2.65 | 1.24 | 2.38 | 1.00 | 0.04 | -0.13 | 0.74 | -0.22 | -1.32 | -0.38 |
| 0.40 | 5.69 | 2.87 | 2.63 | 5.76 | 2.25 | 1.19 | 2.35 | 1.03 | 0.04 | -0.08 | 1.05 | -0.31 | -1.23 | -0.25 |
| GRAVITY CORE G-3 | | | | | | | | | | | | | | |
| 0.00 | 6.33 | 2.56 | 5.56 | 4.72 | 6.77 | 1.16 | 3.01 | -0.91 | -2.22 | -1.09 | -1.33 | -0.81 | -3.35 | -1.98 |
| 0.10 | 5.82 | 2.56 | 2.86 | 5.17 | 6.52 | 1.02 | 2.38 | 0.42 | -0.77 | -0.61 | 0.24 | 0.58 | -2.09 | -0.87 |
| 0.25 | 7.61 | 2.70 | 3.23 | 5.60 | 7.52 | 1.06 | 2.37 | 0.93 | -0.56 | -0.30 | 0.49 | 0.92 | -1.91 | -0.75 |
| 0.44 | 5.26 | 2.52 | 2.41 | 4.24 | 6.43 | 0.99 | 2.00 | 0.79 | -0.27 | -0.33 | 0.48 | 1.08 | -1.62 | -0.60 |
| GRAVITY CORE G-4 | | | | | | | | | | | | | | |
| 0.00 | 5.62 | 2.36 | 2.97 | 4.78 | 6.03 | 1.00 | 2.19 | 0.09 | -1.16 | -0.83 | -0.14 | 0.19 | -2.40 | -1.27 |
| 0.10 | 5.40 | 2.55 | 2.42 | 5.96 | 2.44 | 1.00 | 2.02 | 1.12 | 0.04 | -0.03 | 1.26 | -0.03 | -1.31 | -0.30 |
| 0.20 | 4.94 | 2.46 | 2.17 | 4.94 | 2.43 | 0.99 | 1.89 | 0.98 | -0.02 | -0.20 | 0.98 | -0.04 | -1.33 | -0.40 |
| GRAVITY CORE G-5 | | | | | | | | | | | | | | |
| 0.00 | 5.92 | 2.37 | 2.71 | 4.62 | 6.25 | 0.90 | 2.14 | 0.37 | -0.95 | -0.76 | 0.01 | 0.45 | -2.36 | -1.10 |
| 0.10 | 5.92 | 3.24 | 3.73 | 6.62 | 2.24 | 1.30 | 2.33 | 1.25 | 0.38 | 0.58 | 1.41 | -0.15 | -0.94 | -0.10 |
| 0.25 | 4.90 | 2.72 | 2.58 | 5.25 | 2.87 | 1.08 | 2.20 | 0.93 | 0.09 | 0.01 | 1.03 | 0.16 | -1.25 | -0.22 |
| 0.42 | 5.85 | 2.62 | 2.51 | 5.82 | 2.04 | 1.13 | 2.17 | 1.11 | -0.05 | -0.11 | 1.10 | -0.41 | -1.26 | -0.32 |
| GRAVITY CORE G-6 | | | | | | | | | | | | | | |
| 0.00 | 6.19 | 2.93 | 2.94 | 4.69 | 6.72 | 1.30 | 2.34 | 0.68 | -0.40 | -0.40 | 0.28 | 0.80 | -1.57 | -0.72 |
| 0.10 | 6.38 | 2.61 | 2.57 | 6.68 | 2.99 | 1.10 | 2.32 | 1.03 | -0.26 | -0.28 | 1.10 | -0.06 | -1.51 | -0.43 |
| 0.25 | 5.06 | 2.46 | 2.12 | 5.81 | 2.23 | 0.99 | 2.03 | 0.93 | -0.10 | -0.32 | 1.13 | -0.25 | -1.43 | -0.38 |
| 0.38 | 5.85 | 4.25 | 2.66 | 5.79 | 2.41 | 1.71 | 2.34 | 0.85 | 0.39 | -0.29 | 0.83 | -0.43 | -0.93 | -0.47 |

part of the core, being main components of the assemblage, whilst they are reduced towards the top and their contribution then becomes minor. *L. elliptica* numbers are low at the bottom of the core, but they gradually increase, getting to their highest numbers at the top. Inner shelf foraminifera, and the gastropods *P. conicus* and *V. triquetrus*, are found at the uppermost part of the core sequence in very low numbers. Charophyte gyrogonites have a marked presence along the core, with their highest numbers at the 10 cm depth.

3.2.6. Gravity Corer G6. Core G6 is characterised by the dominance of *C. torosa* forma *litoralis* and the distinct presence of *H. acuta*, particularly in the middle of the core. *C. glaucum*, *A. segmentum* and *A. cf tepida* are present throughout the core, with stable numbers (Fig. 6; Table 3). The total absence of *L. elliptica* from this core can be clearly noted. Once again, inner shelf foraminifera, and the gastropods *P. conicus* and *V. triquetrus*, are found only at the uppermost part of the core. Charophyte gyrogonites in low numbers are found along the core, with their highest values at the top of the core sequence, whereas *Valvata* shells are present at its lowermost part.

3.3. Sediment geochemistry

Table 2 presents the results of major and trace elements, the total organic carbon (TOC) and the grain size characteristics of the gravity core samples, the concentrations of the elements

in the Upper Continental Crust (UCC) (Taylor & McLennan 1985) and the average shale composition (Turekian & Wedepohl 1961).

According to the comparison of the mean composition in our study and the average values reported by Turekian & Wedepohl (1961) for shales, we distinguish Cd, Cr, Cu, Ni, Pb, Zn, MnO, MgO and Na₂O as having mean values been higher than the average shale values. The elements with lower concentrations than average include Al₂O₃, K₂O and V. The Fe₂O₃ and P₂O₅ concentration is similar to that reported for the average shales. Furthermore, CaO, Fe₂O₃, MnO, MgO, Cu, Pb, Ni, and Cr mean values are enriched compared to the concentrations of the Upper Continental Crust (UCC) reported by Taylor & McLennan (1985). By contrast, Al₂O₃, K₂O and Cd are depleted compared to the UCC values. The Na₂O, Zn and V mean values are about the same as the UCC values. TOC profiles of the sediments analysed show variation ranging from 1.56 % to 4.60 % (Table 2).

With respect to the geo-accumulation index, the I_{geo} values are calculated as ≤0 for the majority of the samples analysed, showing unpolluted features for the Gialova Lagoon sediments. Only the I_{geo} values for Cd, Ni and Pb are indicative of the class 1 (Muller 1979), (0 < I_{geo} ≤ 1), showing unpolluted to moderately polluted features (Table 4). The most affected samples are from the G1, G2, G4, G5 and G6 sampling sites.

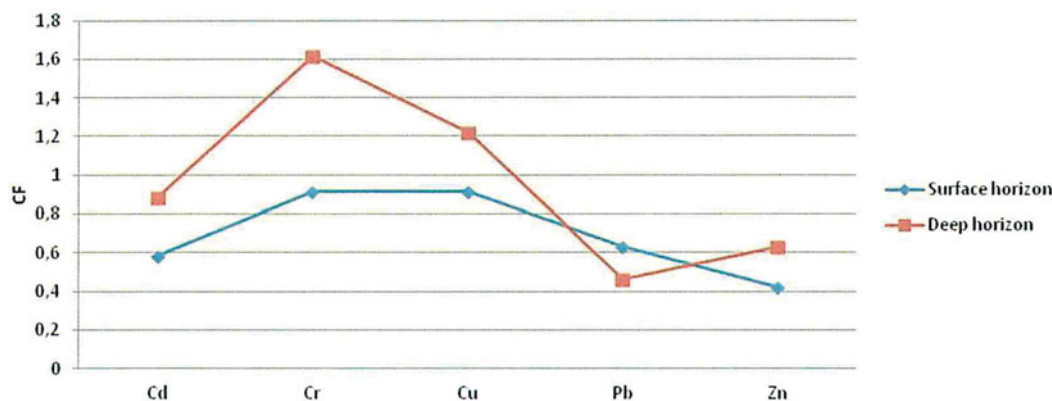


Figure 7 Contamination factor (CF) of various metals for the surface and deep horizon gravity cores samples.

The enrichment factor for elements other than V is well above 1 (Table 4). This is indicative of the source input of fine-grained materials from the surrounding catchment area. In this case, the $EF > 1$ could be interpreted as enrichments caused by anthropogenic activities (Acquavita *et al.* 2010).

The contamination factors of Cd, Cr, Cu, Pb and Zn, calculated for the surface horizon (0.00 cm) and the deep sediment layer (0.40 cm), are below 1 (Fig. 7) indicating a low contamination (Håkanson 1980). In the deep sediment horizon, only Cr and Cu reach values $1 \leq CF < 3$, showing moderate contamination for these samples.

4. Discussion

The hydrology, morphology and sediment characteristics of the Gialova Lagoon are similar to other shallow coastal lagoons of western Greece and are a typical representative example of a Greek lagoon wetland (Avramidis *et al.* 2008, 2013; Papatheodorou *et al.* 2012). The sediment characteristics and the sub-fossil assemblages, as well as the sediment geochemistry and the pollution indices, are influenced by the drainage works that have taken place during the last 70 years in the lagoon (Fig. 2). The attempts at lagoon drainage were started in the 1950s, with the diversion of the Xirolagados River to Voidokilia Bay and the Tiflomity springs, via an artificial channel, to Navarino Bay (Fig. 2b). After 1998, two canals (sluices) were opened, bringing fresh water from both the Xirolagados River and the Tiflomity springs (Fig. 2c).

4.1. Sedimentology – grain size spatial distribution

Based on grain size characteristics, the Gialova Lagoon can be divided into three parts: (i) the northern part, influenced by the Xirolagados River and Tyflomyti springs; (ii) the southern part, adjacent to the sandy barrier and the lagoon inlet; and (iii) the central part, with relatively calm conditions.

The spatial distribution of sand indicates a relatively higher portion (>50 % sand) in a strip area of the lagoon just north of the sandy lagoon barrier and near the inlet (Fig. 4). This high percentage of sand in this area of the lagoon can be related to the littoral sandy material of Navarino Bay, which came into the lagoon via the old inlets, and to the sandy sediments of the barrier, which are transported by the wind into the lagoon. The sorting of the bottom lagoon sediments is relatively constant over the much wider range of mean grain size, as previously mentioned. This evidence suggests a calm condition of a 'quiet water' environment for the lagoon bottom (Buller & MacManus 1972). In the northern part of the lagoon, the grain size cumulative curves are nearly symmetrical and mesokurtic. These evidence a single source area for the sedi-

ments of this part of the lagoon, which is the inflow into the lagoon of the Xirolagados River (Folk 1974). After the diversion of the river mouth, the lagoon has ceased to accept further sediments and the sediments of the northern part can be characterised as 'relict', maintaining their grain size distributions. These 'relict' sediments are extremely poorly sorted, as a result of the flood events of the Xirolagados River. The alternating zones of the Mz values in the northern and northeastern margin of the lagoon suggest the shifting of the Xirolagados River mouth. The spatial distribution of skewness shows the presence of a zone with coarse-skewed values parallel to the long axis of the lagoon. These values are the result of the winnowing process, which causes the removal of the fine-grained tail of distribution (Valia & Cameron 1977; Martins 2003). The winnowing process is caused by re-suspension, with the help of the wind action in the very shallow Gialova Lagoon. The removed fine-grained fraction is deposited again in the area behind the barrier island, which is characterised by a high proportion of sand (Fig. 4). Thus, sediments with fine-skewed values are produced. In this zone, the increased sand content comes from the old inlets and via the winds from the sandy barrier and, to a lesser extent, from the artificial inlet because, currently, a lock-gate in this inlet prevents the entry of sand. The presence of the coarse-skewed values represents a non-depositional or erosional area, whilst the fine-skewed zone is an area dominated by the deposition of fine sediments (Fig. 3c). North of the coarse-skewed zone, the bottom sediments indicate near-symmetrical curves. This suggests the absence of the continuous action of the winnowing process (Duane 1964; Valia & Cameron 1977). In fact, during the summer a large part of the zone with nearly symmetrical values could have been equally dry as today, or it could have been shallower. These conditions prevented the winnowing process and this zone is an area in a state of flux (Martins 2003). It is obvious that the energy in the Gialova Lagoon environment increases gradually from the northern margin to the barrier island.

The increased presence of organic carbon in the eastern part of the lagoon, compared to the western part, is the result of its adjacency with the marshes of the Gianouzagas River. The organic material is transported and deposited in the eastern part, whilst a fast burial prevents the oxidation. The fast burial may be related to flooding events of the Xirolagados River.

4.2. Subfossil assemblages

The temporal and spatial distribution of macro and micro subfossil remains collected from the six cores from the Gialova Lagoon provides information on the evolution of this marginal marine ecosystem. Spatially, the six cores can be separated into three groups: Cores G1 and G2 represent the western part of

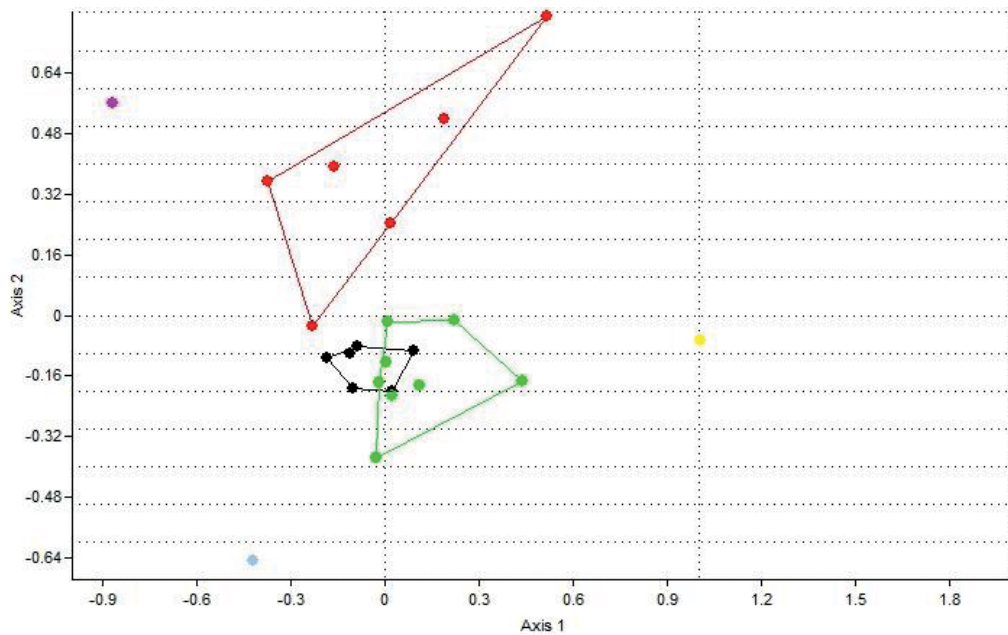


Figure 8 Correspondence analysis plot (axis 1–2) showing the distribution of the 23 samples based on their faunal contents. Red dots = facies 1; black dots = facies 2; green dots = facies 3; grey dot = facies 4; purple dot = sample at 10cm from G3; yellow dot = sample at 25cm from G3.

the lagoon, behind the inlet that connects the lagoon with the Navarino Gulf, hence they are influenced significantly by the marine realm; cores G4, G5 and G6 are located at the north-eastern perimeter of the lagoon, and thus they are influenced by the fresh water input in the lagoon; whereas G3 is found in the middle of the lagoon, depicting a more complicated state (Fig. 1). This segregation is similar to the two zones proposed by McArthur *et al.* (2000) and Koutsoubas *et al.* (2000a), one close to the inlet and one in the inner lagoon.

The 23 samples that were collected from the six cores are characterised mainly by six taxa found almost in every assemblage. These taxa comprise of three mollusc, *C. glaucum*, *A. segmentum* and *H. acuta*, two ostracod, *C. torosa* forma *litoralis* and *L. elliptica*, and one foraminifera species, *A. cf. tepida*, with *C. torosa* forma *litoralis* being clearly the dominant taxon everywhere. All of them are typical shallow brackish lagoonal taxa, tolerant also of a wide range of fluctuations in salinity, temperature and oxygen levels, and thus can be characterised as euryhaline and eurythetmal taxa with a preference to muddy bottoms (Britton 1985; Pascual & Carbonel 1992; Debenay & Guillou 2002; Murray 2006; Kevrekidis *et al.* 2009; Zaibi *et al.* 2012). More specifically, in the Gialova Lagoon such fluctuations occur seasonally, with lower salinities (polysaline) and temperatures in the spring during the end of the wet season (fresh water influxes), and higher salinities (metasaline) and temperatures in the autumn during the end of the dry season (McArthur *et al.* 2000). Consequently, only a small number of tolerant, eurytopic organisms such as the aforementioned ones could survive such harsh conditions in the long term. Furthermore, a number of other taxa were also found in the assemblages in low numbers. Nevertheless, despite their minor contribution in the assemblages, these taxa provide a depiction of the factors that affected temporally the respective lagoon environments. The dominance of *C. torosa* forma *litoralis* in the assemblages indicates that generally low energy environments prevailed in the lagoon (Zaibi *et al.* 2012).

The top ten centimetres in all six cores are characterised by the presence of inner shelf foraminifera (*Peneroplis*, *Cibicides*, *Quinqueloculina*, etc.) and the gastropod *V. triquetrus*, usually located in shallow sublittoral/subtidal marine environments,

as well as by the presence of the gastropod *P. conicus*, indicating a less confined lagoon system with fresh water influx (Kowalke 2005). The presence of charophyte gyrogonites (*Chara* sp.) provides further evidence for the influence of fresh water. In addition, the presence of *Vermetes triquetrus*, especially in G2, G3 and G6, is indicative of shallow protected areas with a marine influence. Consequently, the upper 10cm of all the cores suggest the prevalence of brackish lagoonal conditions in the lagoon; although communication with the Navarino Gulf through the inlet allowed the influx of marine water and, at the same time, inflows of fresh water were discharged into the lagoon through channels from the Xirolagados River and Tyflomyti springs. Consequently, these samples can be attributed to the first facies, as they are clearly separated in the correspondence analysis scatter diagram, and from the dendrogram produced from this analysis (Figs 8, 9). Thus, the depicted lagoonal environment refers to conditions established in the lagoon after the human interventions in the 1980s, when the inlet was constructed and fresh water was allowed again to flow in.

The presence of a *Globigerinoides* sp. test at the top of G2 most likely indicates a random storm event that carried the test in the lagoon.

The middle and lower parts of the six cores present differences which correspond to the three spatial groups. The assemblages in the middle part of cores G1 and G2 (Group 1) consist only of the six main tolerant brackish taxa, indicating brackish, confined lagoonal environments, with significant seasonal fluctuations in salinity and possible dystrophic conditions. The assemblages from the lower part of the two cores also present brackish lagoonal environments; however, the presence of *Chara* sp. gyrogonites suggests the influx of fresh water, at least seasonally. Subsequently, these assemblages characterise the second facies (Figs 8, 9).

Cores from Group 3 (G4, G5 and G6) present almost the same distribution of the six main taxa in the lower and middle levels, with only one exception. Interestingly, charophyte gyrogonites have been found in all the assemblages and particularly in core G5, indicating a constant influx (influence) of fresh water in the lagoon. In addition, a few charophyte gyrogonites have

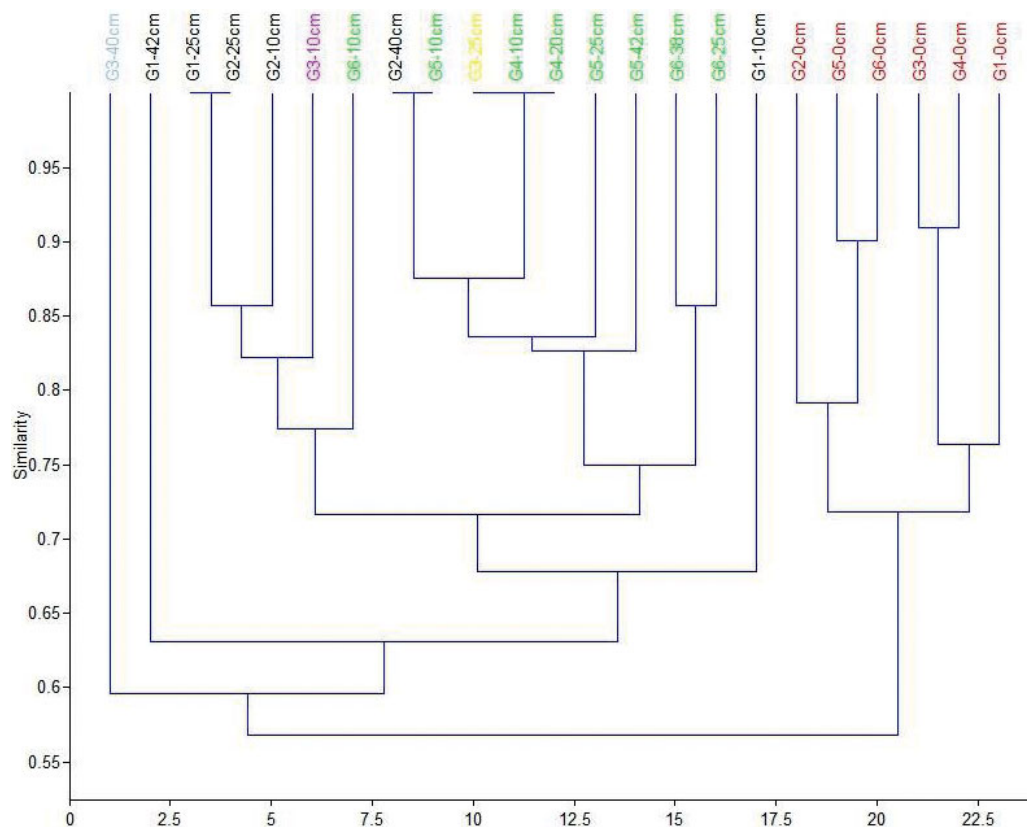


Figure 9 Cluster analysis dendrogram showing the grouping of the 23 samples based on their faunal contents. Red = facies 1; black = facies 2; green = facies 3; grey = facies 4; purple = sample at 10cm from G3; yellow = sample at 25cm from G3.

been recorded from the lower part of Core G6. Therefore, a brackish lagoonal environment with seasonal influx of fresh water can be inferred for the three cores, representing facies three (Figs 8, 9). *L. elliptica* numbers drop in the middle and the lower part of the G5 core whereas, in core G6, *L. elliptica* valves are absent from all samples. Most likely, the impoverishment in *L. elliptica* shows very low energy conditions and no sea water influence, limiting the provision of nutrients (Pascual & Carbonel 1992; Marriner *et al.* 2006; Zaibi *et al.* 2012).

Finally, core G3, which belongs to the second group, is located in approximately the centre of the lagoon. In the middle of the core, a significant increase in *C. glaucum* and *A. segmentum* has been recorded, suggesting favourable conditions for these lagoonal bivalves, which can be correlated with facies two (Figs 8, 9). Nevertheless, at the lower part of the core, the faunal composition changes dramatically; perhaps related to the occasional prevalence of anoxic and/or dystrophic conditions. In particular, the assemblage is clearly impoverished, as only 72 specimens were collected. The majority of the collected material belongs to juvenile individuals, with *C. torosa* forma *litoralis* as the dominant component and with the absence of even common tolerant taxa; we assume that only the easily transported juvenile shells and tests have been recovered from this assemblage. However, a good number of *Valvata* shells (seven individuals), which indicate fresh water influx, have been discovered. Subsequently, brackish anoxic and/or dystrophic conditions characterise the lower part of G3, where fresh water influxes enabled the transportation of lighter shells. This sample characterises the fourth facies, which is clearly separated in the correspondence analysis scatter diagram, and the resultant dendrogram (Figs 8, 9).

In addition to the distribution of the determined taxa from the six cores which represent the different environmental conditions in the lagoon, the correspondence analysis (Fig. 8) and employed cluster analysis (Fig. 9) showed at least four main groups of samples (assemblages) that correspond to four different environmental facies (Figs 5, 6):

1. Brackish lagoonal with shallow marine and fresh water influence;
2. Brackish lagoonal where more confined conditions occur;
3. Brackish lagoonal with fresh water influence;
4. Brackish lagoonal with anoxic and dystrophic conditions.

These observations provide a temporal archive for correlations with previous recent studies on the extant benthic communities of Gialova Lagoon, and particularly on organisms with shells that can be preserved in the sediments (McArthur *et al.* 2000; Koutsoubas *et al.* 2000a, b; Chatzigeorgiou *et al.* 2011). The number of collected valved molluscs taxa from the six cores is clearly reduced, with six gastropod and two bivalve species, compared to the 15 gastropod and five bivalve species from the extant assemblages (Koutsoubas *et al.* 2000a). Nevertheless, three bivalve and seven gastropod species were collected from stations next to the communication channel and thus the marine effect was certainly higher. Despite the smaller number of collected taxa in this study, two of them were not described by Koutsoubas *et al.* (2000b) – the fresh water *Valvata* sp. and the marine *V. triquetrus* – which were traced on the surficial layer (0 cm) and, notably, the latter has been found in all six cores. Another interesting observation is the presence of *P. conicus* only in the surface samples of all cores, and its total absence from all the other deeper ones. This trend matches with the observations of Chatzigeorgiou *et al.* (2011) from their

Table 5 Pearson's correlation coefficients with the significance level of intercorrelation (Sig. = 0.000 indicates $p < 0.001$) for the major and trace elements.

| | | Al ₂ O ₃ | CaO | Fe ₂ O ₃ | K ₂ O | MgO | MnO | Na ₂ O | P ₂ O ₅ | Cd | Cr | Cu | Ni | Pb | V | Zn | TOC | Gravel | Sand | Mud |
|--------------------------------|---------------|--------------------------------|--------|--------------------------------|------------------|--------|--------|-------------------|-------------------------------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| Al ₂ O ₃ | Pearson Corr. | 1,000 | -0.802 | 0.956 | 0.749 | 0.228 | 0.073 | -0.252 | -0.270 | 0.906 | 0.877 | 0.798 | 0.929 | -0.165 | 0.864 | 0.957 | 0.054 | -0.768 | -0.475 | 0.713 |
| | Sig. | - | 0.000 | 0.000 | 0.000 | 0.308 | 0.746 | 0.257 | 0.224 | 0.000 | 0.000 | 0.000 | 0.000 | 0.463 | 0.000 | 0.000 | 0.811 | 0.000 | 0.026 | 0.000 |
| CaO | Pearson Corr. | -0.802 | 1,000 | -0.838 | -0.619 | -0.523 | -0.287 | -0.050 | -0.130 | -0.803 | -0.643 | -0.674 | -0.767 | -0.234 | -0.628 | -0.797 | -0.404 | 0.719 | 0.393 | -0.628 |
| | Sig. | 0.000 | - | 0.000 | 0.002 | 0.013 | 0.195 | 0.825 | 0.566 | 0.000 | 0.001 | 0.001 | 0.000 | 0.295 | 0.002 | 0.000 | 0.062 | 0.000 | 0.071 | 0.002 |
| Fe ₂ O ₃ | Pearson Corr. | 0.956 | -0.838 | 1,000 | 0.797 | 0.291 | 0.110 | -0.140 | -0.219 | 0.928 | 0.916 | 0.857 | 0.929 | -0.098 | 0.910 | 0.975 | 0.045 | -0.731 | -0.576 | 0.773 |
| | Sig. | 0.000 | 0.000 | - | 0.000 | 0.188 | 0.627 | 0.533 | 0.328 | 0.000 | 0.000 | 0.000 | 0.000 | 0.665 | 0.000 | 0.000 | 0.844 | 0.000 | 0.005 | 0.000 |
| K ₂ O | Pearson Corr. | 0.749 | -0.619 | 0.797 | 1,000 | 0.481 | 0.127 | 0.265 | -0.145 | 0.669 | 0.874 | 0.611 | 0.600 | 0.118 | 0.880 | 0.796 | 0.176 | -0.603 | -0.499 | 0.655 |
| | Sig. | 0.000 | 0.002 | 0.000 | - | 0.023 | 0.572 | 0.232 | 0.519 | 0.001 | 0.000 | 0.003 | 0.003 | 0.602 | 0.000 | 0.000 | 0.433 | 0.003 | 0.018 | 0.001 |
| MgO | Pearson Corr. | 0.228 | -0.523 | 0.291 | 0.481 | 1,000 | 0.719 | 0.764 | 0.528 | 0.210 | 0.164 | 0.187 | 0.914 | 0.848 | 0.159 | 0.242 | 0.667 | -0.200 | -0.056 | 0.134 |
| | Sig. | 0.308 | 0.013 | 0.188 | 0.023 | - | 0.000 | 0.000 | 0.011 | 0.348 | 0.465 | 0.405 | 0.050 | 0.000 | 0.479 | 0.278 | 0.001 | 0.372 | 0.805 | 0.551 |
| MnO | Pearson Corr. | 0.073 | -0.287 | 0.110 | 0.127 | 0.719 | 1,000 | 0.575 | 0.714 | 0.032 | -0.030 | 0.109 | -0.079 | 0.658 | -0.063 | 0.024 | 0.479 | -0.176 | -0.008 | 0.087 |
| | Sig. | 0.746 | 0.195 | 0.627 | 0.572 | 0.000 | - | 0.005 | 0.000 | 0.886 | 0.895 | 0.631 | 0.726 | 0.001 | 0.779 | 0.914 | 0.024 | 0.434 | 0.972 | 0.701 |
| Na ₂ O | Pearson Corr. | -0.252 | -0.050 | -0.140 | 0.265 | 0.764 | 0.575 | 1,000 | 0.556 | -0.237 | -0.104 | -0.163 | -0.430 | 0.828 | -0.101 | -0.211 | 0.535 | 0.080 | 0.056 | -0.079 |
| | Sig. | 0.257 | 0.825 | 0.533 | 0.232 | 0.000 | 0.005 | - | 0.007 | 0.289 | 0.644 | 0.468 | 0.046 | 0.000 | 0.656 | 0.345 | 0.010 | 0.722 | 0.804 | 0.725 |
| P ₂ O ₅ | Pearson Corr. | -0.270 | -0.130 | -0.219 | -0.145 | 0.528 | 0.714 | 0.556 | 1,000 | -0.239 | -0.347 | -0.332 | -0.369 | 0.740 | -0.383 | -0.326 | 0.475 | -0.037 | 0.034 | -0.009 |
| | Sig. | 0.224 | 0.566 | 0.328 | 0.519 | 0.011 | 0.000 | 0.007 | - | 0.284 | 0.113 | 0.132 | 0.091 | 0.000 | 0.079 | 0.139 | 0.025 | 0.869 | 0.881 | 0.970 |
| Cd | Pearson Corr. | 0.906 | -0.803 | 0.928 | 0.669 | 0.210 | 0.032 | -0.237 | -0.239 | 1,000 | 0.825 | 0.828 | 0.906 | -0.111 | 0.816 | 0.919 | 0.164 | -0.674 | -0.410 | 0.620 |
| | Sig. | 0.000 | 0.000 | 0.000 | 0.001 | 0.348 | 0.886 | 0.289 | 0.284 | - | 0.000 | 0.000 | 0.000 | 0.622 | 0.000 | 0.000 | 0.466 | 0.001 | 0.058 | 0.002 |
| Cr | Pearson Corr. | 0.877 | -0.643 | 0.916 | 0.874 | 0.164 | -0.030 | -0.104 | -0.347 | 0.825 | 1,000 | 0.830 | 0.850 | -0.229 | 0.993 | 0.926 | -0.073 | -0.693 | -0.547 | 0.733 |
| | Sig. | 0.000 | 0.001 | 0.000 | 0.000 | 0.465 | 0.895 | 0.644 | 0.113 | 0.000 | - | 0.000 | 0.000 | 0.305 | 0.000 | 0.000 | 0.747 | 0.000 | 0.008 | 0.000 |
| Cu | Pearson Corr. | 0.798 | -0.674 | 0.857 | 0.611 | 0.187 | 0.109 | -0.163 | -0.332 | 0.828 | 0.830 | 1,000 | 0.841 | -0.158 | 0.820 | 0.872 | -0.087 | -0.584 | -0.426 | 0.592 |
| | Sig. | 0.000 | 0.001 | 0.000 | 0.003 | 0.405 | 0.631 | 0.468 | 0.132 | 0.000 | 0.000 | - | 0.000 | 0.482 | 0.000 | 0.000 | 0.701 | 0.004 | 0.048 | 0.004 |
| Ni | Pearson Corr. | 0.929 | -0.767 | 0.929 | 0.600 | 0.014 | -0.079 | -0.430 | -0.369 | 0.906 | 0.850 | 0.841 | 1,000 | -0.351 | 0.843 | 0.942 | -0.110 | -0.672 | -0.548 | 0.724 |
| | Sig. | 0.000 | 0.000 | 0.000 | 0.003 | 0.950 | 0.726 | 0.046 | 0.091 | 0.000 | 0.000 | 0.000 | - | 0.109 | 0.000 | 0.000 | 0.627 | 0.001 | 0.008 | 0.000 |
| Pb | Pearson Corr. | -0.165 | -0.234 | -0.098 | 0.118 | 0.848 | 0.658 | 0.828 | 0.740 | -0.111 | -0.229 | -0.158 | -0.351 | 1,000 | -0.237 | -0.186 | 0.640 | 0.080 | 0.069 | -0.089 |
| | Sig. | 0.463 | 0.295 | 0.665 | 0.602 | 0.000 | 0.001 | 0.000 | 0.000 | 0.622 | 0.305 | 0.482 | 0.109 | - | 0.288 | 0.408 | 0.001 | 0.724 | 0.761 | 0.695 |
| V | Pearson Corr. | 0.864 | -0.628 | 0.910 | 0.880 | 0.159 | -0.063 | -0.101 | -0.383 | 0.816 | 0.993 | 0.820 | 0.843 | -0.237 | 1,000 | 0.923 | -0.096 | -0.641 | -0.580 | 0.734 |
| | Sig. | 0.000 | 0.002 | 0.000 | 0.000 | 0.479 | 0.779 | 0.656 | 0.079 | 0.000 | 0.000 | 0.000 | 0.000 | 0.288 | - | 0.000 | 0.671 | 0.001 | 0.005 | 0.000 |
| Zn | Pearson Corr. | 0.957 | -0.797 | 0.975 | 0.796 | 0.242 | 0.024 | -0.211 | -0.326 | 0.919 | 0.926 | 0.872 | 0.942 | -0.186 | 0.923 | 1,000 | 0.008 | -0.703 | -0.534 | 0.728 |
| | Sig. | 0.000 | 0.000 | 0.000 | 0.000 | 0.278 | 0.914 | 0.345 | 0.139 | 0.000 | 0.000 | 0.000 | 0.000 | 0.408 | 0.000 | - | 0.973 | 0.000 | 0.011 | 0.000 |
| TOC | Pearson Corr. | 0.054 | -0.404 | 0.045 | 0.176 | 0.667 | 0.479 | 0.535 | 0.475 | 0.164 | -0.073 | -0.087 | -0.110 | 0.640 | -0.096 | 0.008 | 1,000 | -0.187 | 0.409 | -0.224 |
| | Sig. | 0.811 | 0.062 | 0.844 | 0.433 | 0.001 | 0.024 | 0.010 | 0.025 | 0.466 | 0.747 | 0.701 | 0.627 | 0.001 | 0.671 | 0.973 | - | 0.404 | 0.059 | 0.317 |
| Gravel | Pearson Corr. | -0.768 | 0.719 | -0.731 | -0.603 | -0.200 | -0.176 | 0.080 | -0.037 | -0.674 | -0.693 | -0.584 | -0.672 | 0.080 | -0.641 | -0.703 | -0.187 | 1,000 | 0.307 | -0.693 |
| | Sig. | 0.000 | 0.000 | 0.000 | 0.003 | 0.372 | 0.434 | 0.722 | 0.869 | 0.001 | 0.000 | 0.004 | 0.001 | 0.724 | 0.001 | 0.000 | 0.404 | - | 0.165 | 0.000 |
| Sand | Pearson Corr. | -0.475 | 0.393 | -0.576 | -0.499 | -0.056 | -0.008 | 0.056 | 0.034 | -0.410 | -0.547 | -0.426 | -0.548 | 0.069 | -0.580 | -0.534 | 0.409 | 0.307 | 1,000 | -0.899 |
| | Sig. | 0.026 | 0.071 | 0.005 | 0.018 | 0.805 | 0.972 | 0.804 | 0.881 | 0.058 | 0.008 | 0.048 | 0.008 | 0.761 | 0.005 | 0.011 | 0.059 | 0.165 | - | 0.000 |
| Mud | Pearson Corr. | 0.713 | -0.628 | 0.773 | 0.655 | 0.134 | 0.087 | -0.079 | -0.009 | 0.620 | 0.733 | 0.592 | 0.724 | -0.089 | 0.734 | 0.728 | -0.224 | -0.693 | -0.899 | 1,000 |
| | Sig. | 0.000 | 0.002 | 0.000 | 0.001 | 0.551 | 0.701 | 0.725 | 0.970 | 0.002 | 0.000 | 0.004 | 0.000 | 0.695 | 0.000 | 0.000 | 0.317 | 0.000 | 0.000 | - |

1998 samplings after the opening of two canals that brought fresh water to the lagoon. In addition, they also reported the significant population reduction of *H. acuta*, which is present in considerable populations in almost all core samples and in the previous sampling (Koutsoubas *et al.* 2000b). Conversely, except for *P. conicus*, all the other typical brackish lagoon taxa (*C. glaucum*, *A. segmentum* and *H. acuta*) are found in all core samples. Also, the spatial distribution of foraminifera from the core samples is in agreement with the observations of McArthur *et al.* (2000), where increased numbers of foraminifera are found close to the stations and cores found closer to the communication channel and their numbers decrease in the inner lagoon.

Generally, two trends in the temporal and spatial evolution of the environmental facies can be observed. The first trend is observed in the cores close to the inlet (G1 and G2), where the brackish lagoonal facies with more confined conditions changes at the top of the sequence to brackish lagoonal facies with shallow marine and fresh water influence. The second trend, as has been observed in the inner lagoon, indicates a change from a brackish lagoonal facies with fresh water influence to a brackish lagoonal facies with shallow marine and fresh water influence. Obviously, human interventions caused these changes. In addition, evidence for an intense dystrophic and/or hypoxic event that affected the lagoon is found at the lower part of core G3, and this can be correlated with similar studies on recent assemblages (Koutsoubas *et al.* 2000a; Chat-

zigeorgiou *et al.* 2011), indicating that such events did occur in the past (Fig. 5c, facies 4).

4.3. Geochemistry – environmental pollution

Pearson's correlation was employed in order to access the association between the elements analyzed. Based on this analysis we distinguish two different geochemical groups regarding the inter-element strong positive loadings and significant correlations (Table 5):

- The Al₂O₃, Fe₂O₃, K₂O, Cd, Cr, Cu, Ni, V, Zn group as representative of the clay fraction of the sediments. These elements are closely related to corresponding changes in the concentrations of clays, showing positive loadings with the mud content in the samples analysed, and can thus be considered to represent the aluminosilicate fraction (Koinig *et al.* 2003). In addition, Cd, Cr, Cu, Ni, V and Zn distribution is considered to be affected by anthropogenic inputs in lakes and lagoons (Bellucci *et al.* 2010), such as the notably high Ni content in the analysed samples.
- The TOC, Na₂O, MgO, MnO, P₂O₅ and Pb group. TOC represents the organic fraction in the sediment and can be used as an indicator of the organic matter production (Calvert & Pedersen 1993; Meyers & Ishiwatari 1993), whilst Na is related to salinity and authigenic input (Parker *et al.* 2006). The significant positive correlation between organic matter Mn and Pb indicates that TOC is a good carrier for both elements in the analysed sediments samples.

Table 6 Mean values of heavy metals concentrations in the Gialova Lagoon as compared to other Greek aquatic systems, Upper Continental Crust (UCC), average Shale's compositions and sediment quality guidelines for metals.

| Element Unit | Fe ₂ O ₃ % | MnO % | Fe % | Mn ppm | Zn ppm | Cu ppm | Pb ppm | Ni ppm | Cr ppm | V ppm | Cd ppm | Reference |
|--|-------------------------------------|----------|---------|-----------|-----------|-----------|-----------|-----------|-----------|----------|-----------|------------------------------------|
| Gialova Lagoon, mean | 6.80 | 0.20 | 4.76* | 1549* | 98.47 | 57.51 | 36.29 | 172.11 | 118.20 | 69.24 | 0.81 | This study |
| Messolonghi Lagoon | 3.37 | 0.08* | 2.36* | 630.00 | 60.00 | 20.00 | 16.00 | 84.00 | 101.00 | 75.00 | – | Karageorgis <i>et al.</i> 2012 |
| Koumoundourou Lake | 0.83 | 0.02* | 0.58* | 155.00 | 83.00 | 21.00 | 53.00 | 28.00 | 58.00 | 23.00 | – | Karageorgis <i>et al.</i> 2009 |
| Alikes Lagoon | 5.13 | 0.08 | 3.59* | 630.00 | 68.89 | 29.17 | 9.88 | 134.17 | 251.67 | 109.33 | – | Panagiotaras <i>et al.</i> 2012 |
| Aetoliko Lagoon | 5.96* | 0.11* | 4.17 | 837 | 122.2 | 88 | – | 75.6 | 140 | – | – | Dassenakis <i>et al.</i> 1994 |
| Kleisova Lagoon | 2.34* | 0.07* | 1.64 | 562.00 | 29.00 | 13.00 | – | 62.00 | – | – | – | Papatheodorou <i>et al.</i> 2002 |
| Rhodia Lagoon | 4.04* | 0.11* | 2.83 | 867 | 72 | 37 | 36 | 124 | 231 | 112 | – | Karageorgis 2007 |
| Tsoukalio Lagoon | 4.52* | 0.15* | 3.16 | 1191 | 76 | 31 | 26 | 131 | 274 | 108 | – | Karageorgis 2007 |
| Logarou Lagoon | 6.78* | 0.12* | 4.74 | 922 | 105 | 44 | 25 | 221 | 302 | 153 | – | Karageorgis 2007 |
| Tsopeli Lagoon | 5.69* | 0.09* | 3.98 | 665 | 100 | 48 | 26 | 168 | 295 | 129 | – | Karageorgis 2007 |
| Navarino Bay – upper sediments | – | – | – | – | 352.00 | 66.00 | – | 151.00 | – | – | – | Varnavas <i>et al.</i> 1987 |
| Agiasma Lagoon – station A | – | – | – | – | 49.54 | 25.86 | 75.69 | 42.31 | – | – | – | Christophoridis <i>et al.</i> 2007 |
| UCC | 5.00* | 0.08* | 3.5 | 600 | 71 | 25 | 20 | 20 | 35 | 60 | – | Taylor McClennan 1985 |
| Shales | 6.75* | 0.11* | 4.72 | 850 | 95 | 45 | 20 | 68 | 90 | 130 | – | Turekian Wedepohl 1961 |
| threshold effect level (TEL) in sediment quality guidelines for metals | – | – | – | – | 123 | 35.7 | 35 | 18 | 37.3 | – | 0.6 | Burton 2002 |
| Pre-industrial reference level | – | – | – | – | 175 | 50 | 70 | – | 90 | – | 1 | Håkanson 1980 |
| Toxic-response factor | – | – | – | – | 1 | 5 | 5 | – | 2 | – | 30 | Håkanson 1980 |

* Calculated concentrations using the appropriate molar ratio

In order to better evaluate the pollution level in the Gialova Lagoon sediments, an attempt was made to compare our data with the data reported for Greek lagoons by other researchers. We also include for comparison data presented for the Navarino Bay surface sediments (Varnavas *et al.* 1987) because they are influenced by the same surrounding environmental conditions. In respect to other data obtained for Greek lagoon sediments, the Gialova mean Cu value (57.51 ppm) is in the same order of magnitude as the value reported by Varnavas *et al.* (1987) for the upper sediments in Navarino Bay (66.00 ppm). By contrast, Christophoridis *et al.* (2007) reported an average value of 25.86 ppm for the Cu in the Agiasma Lagoon, station A sediments. The Ni values in our study are about the same as the Ni concentrations reported by Varnavas *et al.* (1987) and Karageorgis (2007). The Christophoridis *et al.* (2007) Ni values are about four times lower than the Ni values in our study (Table 6).

However, the Zn mean concentration in the Gialova Lagoon sediments (98.47 ppm) is well below that of the 352.00 ppm reported for the upper sediments in Navarino Bay (Varnavas *et al.* 1987). The relationship of Cu, Ni and Zn with the clay content of the samples in our study is well documented by their positive loadings with Al₂O₃. This is in agreement with the findings of Varnavas *et al.* (1987), who reported the same geochemical association for Cu, Ni, Zn and Al₂O₃ in the Navarino sediments. In addition, the well-above-1 enrichment factors calculated for the Gialova Lagoon samples (Table 6) indicate an additional contribution to the total distribution of Cd, Cr, Cu, Ni, Pb and Zn in the sediments.

Furthermore, the enrichment for Ni can be related to the weathering of the bauxite deposits from the surrounding environment. This finding is consistent with the high Ni/Al ratios reported by Varnavas *et al.* (1987) for the Navarino sediments. Moreover, the Zn value (49.54 ppm) in the Agiasma Lagoon sediments are about 50 % lower than the zinc concentration for the Gialova Lagoon sediments (Table 6). In addition, the average values reported in our study for Fe₂O₃, MnO, Zn, Cu, Pb and Ni in the Gialova Lagoon sediments are above those reported by Panagiotaras *et al.* (2012) for the Alikes Lagoon sediments (Table 6). This trend is opposite for the elements V and Cr, where the Alikes Lagoon sediments are enriched compared to those of the Gialova Lagoon (Table 6).

Based on concentrations of total organic carbon (TOC) and total extractable heavy metals of the six gravity cores, and using the pollution indices, the Gialova Lagoon seems to be unpolluted with moderate contamination. In all the gravity cores, a down-core increase of the total metals was observed, indicating a recent environmental improvement of the lagoon status. An exception to these general trends is Pb, which indicates a slight enrichment of the surface sediments (Fig. 7), whilst it shows high positive loading with TOC and thus this can be attributed to the urban anthropogenic activities (Christophoridis *et al.* 2007). Comparing the mean values of heavy metals concentrations of the Gialova Lagoon with sediment quality guidelines for metals (Burton 2002), it seems that metal concentrations of Cr, Cu and Ni in the lagoon sediments exceed the threshold effect level (Table 6). According to spatial distribution of heavy metals, it seems that the most unaffected station is G3, which is located in the central part of the lagoon, whilst the peripheral sampling sites are more affected. The concentrations of trace metals in the sediments of the Gialova Lagoon, although higher than other Greek coastal lagoons (Table 6), still suggest that the lagoon appears to be unaffected by human activity. The contamination factors for Cd, Cr, Cu, Pb and Zn presented here are comparable with those reported for the Messolonghi Lagoon (Karageorgis *et al.* 2012), the Rodia, Tsoukalio, Logarou and Tsopeli lagoons (Karageorgis 2007), the Aetoliko Lagoon (Dasenakis *et al.*

1994), the Klisova Lagoon (Papatheodorou *et al.* 2002) and the Koumoundourou Lake (Karageorgis *et al.* 2009).

Using the sediment quality guidelines for metals reported by Burton (2002) and MacDonald *et al.* (2000), we consider that the most affected surface sediments for Cd are from the G1, G2 and G6 sampling sites. Using the same criterion for Cr, all the sampling sites exceed the threshold effect level (37.3 ppm, Table 6). The same features are obvious for Cu and Ni. By contrast, Zn shows values below the threshold effect level (123 ppm) for all the analysed sediment samples. The G1, G2, G5 and G6 sampling sites show Pb values for the surface sediments that are above the threshold effect level (35 ppm). The G4 sampling site is considered unaffected for Pb, using the sediment quality guidelines for metals reported by Burton (2002) and MacDonald *et al.* (2000).

5. Conclusions

During the last 70 years, the Gialova Lagoon has suffered from different human interventions which have influenced the sediment dispersal mechanism, the subfossil assemblages and the environmental status of the lagoon. During the last 20 years, efforts to restore the hydrological regime, and the removal of constructions for drainage of the lagoon, have improved the ecosystem status. The lagoon has the characteristics of a typical shallow coastal Mediterranean lagoon and it is protected by international conventions; therefore it can be considered an important case study concerning how inter-temporal human interventions and restoration efforts are reflected in the depositional environments, palaeoecology and sediment geochemistry. The present study indicates the influence of the hydrologic changes on spatial sediment characteristics. The vertical examination of subfossil taxa in gravity cores shows different depositional environments and environmental facies of a shallow brackish lagoon, dominated by periods of marine influence and fresh water inflow and by hypoxic and dystrophic conditions. Generally, it has been clear that in the past the Gialova Lagoon was partitioned into two zones, one around the inlet (cores G1 and G2), where rather confined brackish lagoonal conditions prevailed with an influence from the shallow marine shelf, and another one at the internal perimeter of the lagoon (cores G4, G5 and G6), where the brackish lagoonal environment was affected by fresh water influx. Nevertheless, the bottom surface sediments present evidence indicating that, recently, conditions in the lagoon have become more homogeneous, characterised by shallow shelf and fresh water influence.

Moreover, the vertical distribution of the lagoon's metal concentrations and contamination indices indicates the gradual improvement in the lagoon's environmental conditions. The above spatiotemporal sedimentological, geochemical and palaeoecological data constitute a useful tool and an environmental baseline for the decision makers to develop procedures, according to European Union Directives, to control and limit the environmental risk of the protected and threatened Gialova Lagoon ecosystem.

6. References

- Abraham, G. M. S. & Parker, R. J. 2008. Assessment of heavy metal enrichment factors and the degree of contamination in marine sediments from Tamaki Estuary, Auckland, New Zealand. *Environmental Monitoring Assessment* **136**, 227–38.
- Acquavita, A., Predonzani, S., Mattassi, G., Rossin, P., Tamberlich, F., Falomo, J. & Valic, I. 2010. Heavy metal contents and distribution in coastal sediments of the Gulf of Trieste (Northern Adriatic Sea, Italy). *Water Air Soil Pollution* **211**, 95–111.
- Appeltans, W., Bouchet, P., Boxshall, G. A., De Broyer, C., de Voogd, N. J., Gordon, D. P., Hoeksema, B. W., Horton, T.,

- Kennedy, M., Mees, J., Poore, G. C. B., Read, G., Stöhr, S., Walter, T.C. & Costello, M. J. (eds) 2012. *World Register of Marine Species*. Accessed at <http://www.marinespecies.org> on 2013-08-05.
- Arvanitidis, C., Koutsoubas, D., Dounas, K. & Eleftheriou, A. 1999. Annelid fauna of a Mediterranean lagoon (Gialova Lagoon, southwest Greece): community structure in a severely fluctuating environment. *Journal of the Marine Biological Association of the UK* **79**, 849–86.
- Aubouin, J. 1959. Contribution à l'étude géologique de la Grèce septentrionale: le confins de l'Épire et de la Thessalie. *Annales Géologiques Pays Hellenique* **10**, 1–483.
- Avramidis, P., Bouzos, D., Antoniou, V. & Kontopoulos, N. 2008. Application of grain size trend analysis and spatiotemporal changes, as a tool for lagoon management. Case study of the Kotychi Lagoon, western Greece. *Geologica Carpathica* **59**, 261–68.
- Avramidis, P., Bekiari V., Kontopoulos, N. & Kokidis, N. 2013. Shallow coastal lagoon sediment characteristics and water physicochemical parameters – Myrtari Lagoon, Mediterranean Sea, western Greece. *Fresenius Environmental Bulletin* **22**, 1625–38.
- Bellucci, L. G., Giuliani, S., Mugnai, C., Frignani, M., Paolucci, D., Albertazzi, S. & Ruiz Fernandez, A. C. 2010. Anthropogenic metal delivery in sediments of Porto Marghera and Venice Lagoon (Italy). *Soil Sediment Contamination* **19**, 42–57.
- Bird, E. 2008. *Coastal geomorphology. An introduction*. Chichester: John Wiley and Sons Ltd.
- Blaser, P., Zimmermann, S., Luster, J. & Shoty, W. 2000. Critical Examination of Trace Element Enrichments and Depletions in Soils: As, Cr, Cu, Ni, Pb and Zn in Swiss Forest Soils. *The Science of the Total Environment* **249**, 257–80.
- Britton, R. H. 1985. Life cycle and production of *Hydrobia acuta* Drap. (Gastropoda: Prosobranchia) in a hypersaline coastal lagoon. *Hydrobiologia* **122**, 219–30.
- Buller, A. T. & MacManus, J. 1972. Simple metric sedimentary statistics used to recognize different environments. *Sedimentology* **18**, 1–21.
- Burton, J. G. A. 2002. Sediment quality criteria in use around the world. *Limnology* **3**, 65–75.
- Calvert, S. E. & Pedersen, T. F. 1993. Geochemistry of Recent oxic and anoxic marine sediments: Implications for the geological record. *Marine Geology* **113**, 67–88.
- Chatzigeorgiou, G., Reizopoulou, S., Maidanou, M., Naletaki, M., Orneraki, E., Apostolaki, E. & Arvanitidis, C. 2011. Macrobenthic community changes due to dystrophic events and freshwater inflow: Changes in space and time in a Mediterranean lagoon (Gialova Lagoon, SW Greece). *Estuarine Coastal and Shelf Science* **94**, 111–21.
- Christophoridis, A., Stamatis, N. & Orfanidis, S. 2007. Sediment heavy metals of a Mediterranean coastal lagoon: Agiasma, Nestos Delta, Eastern Macedonia (Greece). *Transitional Waters Bulletin* **1**, 33–43.
- Dassenakis, M., Krasakopoulou, E. & Matzara, B. 1994. Chemical Characteristics of Aetoliko Lagoon, Greece, after an Ecological Shock. *Marine Pollution Bulletin* **28**, 427–33.
- Debenay, J. P. & Guillou, J. J. 2002. Ecological transitions indicated by foraminiferal assemblages in paralic environments. *Estuaries* **25**(6A), 1107–20.
- Duane, D. B. 1964. Significance of skewness in recent sediments, western Pamlico Sound, North Carolina. *Journal of Sedimentary Petrology* **34**, 864–74.
- Duck, R. & Silva, J. F. 2012. Coastal lagoons and their evolution: A hydromorphological perspective. *Estuarine, Coastal and Shelf Science* **110**, 2–14.
- Folk, R. I. 1974. *Petrology of sedimentary rocks*. Austin, Texas: Hemphill. 170 pp.
- Gaudette, H., Flight, W., Toner, L. & Folger, D. 1974. An inexpensive titration method for the determination of organic carbon in recent sediments. *Journal of Sedimentary Petrology* **44**, 249–53.
- Golterman, H. 2004. *The Chemistry of Phosphate and Nitrogen Compounds in Sediments*. Dordrecht, The Netherlands: Kluwer Academic Publishers.
- Greek Coastal Zone Management Report. 2006. *National Report of Greece on Coastal Zone Management*. Athens: Ministry of Environment, Physical Planning and Public Works. 92 pp.
- Håkanson, L. 1980. An Ecological Risk Index for Aquatic Pollution Control: A Sedimentological Approach. *Water Research* **14**, 975–1001.
- Harikumar, P. S. & Jisha, T. S. 2010. Distribution pattern of trace metal pollutants in the sediments of an urban wetland in the southwest coast of India. *International Journal of Engineering Science and Technology* **2**, 840–50.
- Institute of Geology and Mineral Exploration of Greece (I.G.M.E.). 1980. *Geological Map Sheet Koroni–Pylos–Skhiza*. Scale 1:50000.
- Jaccard, P. 1912. The distribution of the flora in the alpine zone. *New Phytologist* **11**(2), 37–50.
- Karageorgis, A. P. 2007. Geochemical study of sediments from the Amvrakikos Gulf lagoon complex, Greece. *Transitional Waters Bulletin* **3**, 3–8.
- Karageorgis, A. P., Katsanevakis, S. & Kaberi, H. 2009. Use of enrichment factors for the assessment of heavy metal contamination in the sediments of Koumoundourou Lake, Greece. *Water Air Soil Pollution* **204**, 243–58.
- Karageorgis, A. P., Sioulas, A., Krasakopoulou, E., Anagnostou, C. L., Hatiris, G. A., Kyriakidou, H. & Vasilopoulos, K. 2012. Geochemistry of surface sediments and heavy metal contamination assessment: Messolonghi lagoon complex, Greece. *Environmental Earth Science* **65**, 1619–29.
- Kevrekidis, T., Kasapis, K. & Kalpia, V. 2009. Life cycle, population dynamics, growth and production of *Abra segmentum* (Mollusca, Bivalvia) at low salinities in a Mediterranean lagoon. *Helgoland Marine Research* **63**, 277–85.
- Kjerfve, B. 1994. *Coastal lagoon processes*. Amsterdam: Elsevier. 576 pp.
- Koinig, K. A., Shoty, W., Lotter, A. F., Ohlendorf, C. & Sturm, S. 2003. 9000 years of geochemical evolution of lithogenic major and trace elements in the sediment of an alpine lake: the role of climate, vegetation, and land-use history. *Journal of Paleolimnology* **30**, 307–20.
- Koutsoubas, D., Dounas, C., Arvanitidis, C., Kornilos, S., Petihakis, G., Triantafyllou, G. & Eleftheriou, A. 2000a. Macrobenthic community structure and disturbance assessment in Gialova Lagoon, Ionian Sea. *ICES Journal of Marine Science* **57**, 1472–80.
- Koutsoubas, D., Arvanitidis, C., Dounas, C. & Drummond, L. 2000b. Community structure and dynamics of the molluscan fauna in a mediterranean lagoon (Gialova Lagoon, SW Greece). *Belgian Journal of Zoology* **130**, 135–42.
- Kowalke, T. 2005. Mollusca in marginal marine and inland saline aquatic ecosystems – examples of Cretaceous to extant evolutionary dynamics. *Zitteliana* **A45**, 35–64.
- Liu, W. H., Zhao, J. Z., Ouyang, Z. Y., Söderlund, L. & Liu, G. H. 2005. Impacts of Sewage Irrigation on Heavy Metal Distribution and Contamination in Beijing, China. *Environment International* **31**, 805–12.
- McArthur, V., Koutsoubas, D., Lambadariou, N. & Dounas, C. 2000. The meiofaunal community structure of a Mediterranean lagoon (Gialova Lagoon, Ionian Sea). *Helgoland Marine Research* **54**, 7–17.
- MacDonald, D. D., Ingersoll, C. G. & Berger, T. A. 2000. Development and evaluation of consensus-based sediment quality guidelines for freshwater ecosystems. *Archives of Environmental Contamination Toxicology* **39**, 20–31.
- Marriner, N., Morhange, C. & Doumet-Serhal, C. 2006. Geoarchaeology of Sidon's ancient harbours, Phoenicia. *Journal of Archaeological Science* **33**, 1514–35.
- Martins, L. R. 2003. Recent sediments and grain size analysis. *Gravel* **1**, 90–105.
- Meyers, P. A. & Ishiwatari, R. 1993. Lacustrine organic geochemistry—an overview of indicators of organic matter sources and diagenesis in lake sediments. *Organic Geochemistry* **20**, 867–900.
- Müller, G. 1979. Schwermetalle in den sedimenten des Rheins-Veränderungenseit 1971. *Umschau* **79**, 778–83.
- Murray, J. W. 2006. *Ecology and Applications of Benthic Foraminifera*. Cambridge: Cambridge University Press. 440 pp.
- Panagiotaras, D., Papoulis, D., Kontopoulos N. & Avramidis, P. 2012. Geochemical processes and sedimentological characteristics of Holocene lagoon deposits, Alikes Lagoon Zakynthos island, Western Greece. *Geological Journal* **47**, 372–87.
- Papatheodorou, G., Hotos, G., Geraga, M., Avramidou, D. & Vori-nakis, T. 2002. Heavy metal concentrations in sediments of Klisova Lagoon (Southeast Messolonghi – Aetolikon Lagoon complex), W. Greece. *Fresenius Environmental Bulletin* **11**, 951–56.
- Papatheodorou, G., Avramidis, P., Fakiris, E., Christodoulou, D. & Kontopoulos, N. 2012. Bed diversity in the shallow water environment of Pappas lagoon in Greece. *International Journal of Sediment Research* **27**, 1–17.
- Parker, A. G., Goudie, A. S., Stokes, S., White, K., Hodson, M. J., Manning, M. & Kennet, D. 2006. A record of Holocene climate change from lake geochemical analyses in southeastern Arabia. *Quaternary Research* **66**, 465–76.
- Pascual, A. & Carbonel, P. 1992. Distribution and annual variations of *Loxococoncha elliptica* in the Gernika estuary (Bay of Biscay). *Geobios* **25**, 495–503.
- Petihakis, G., Triantafyllou, G., Koutsoubas, D., Allen, I. & Dounas, C. 1999. Modelling the annual cycles of nutrients and phyto-

- plankton in a Mediterranean lagoon (Gialova, Greece). *Marine Environmental Research* **48**, 37–58.
- Reimann, C. & de Caritat, P. 2005. Distinguishing between Natural and Anthropogenic Sources for Elements in the Environment: Regional Geochemical Surveys versus Enrichment Factors. *The Science of the Total Environment* **337**, 91–107.
- Rubio, B., Nombela, M. A. & Vilas, F. 2000. Geochemistry of Major and Trace Elements in Sediments of the Ria de Vigo (NW Spain): an Assessment of Metal Pollution. *Marine Pollution Bulletin* **40**, 968–80.
- Sahu, K. C. & Bhosale, U. 1991. Heavy metal pollution around the island city of Bombay, India. Part I: quantification of heavy metal pollution of aquatic sediments and recognition of environmental discriminants. *Chemical Geology* **91**, 263–83.
- Salomons, W. & Förstner, U. 1984. *Metals in the hydrocycle*. Berlin, Heidelberg, New York, Tokyo: Springer-Verlag. 368 pp.
- Subramanian, V. & Mohanachandran, G. 1990. Heavy metals distribution and enrichment in the sediments of Southern East Coast of India. *Marine Pollution Bulletin* **2**, 324–30.
- Taylor, S. R. & McLennan, S. M. 1985. *The Continental Crust: its Composition and Evolution*. Oxford, London, Edinburgh, Boston, Palo Alto, Melbourne: Blackwell Scientific. 312 pp.
- Turekian, K. K. & Wedepohl, K. H. 1961. Distribution of the Elements in Some Major Units of the Earth's Crust. *Geological Society of America Bulletin* **72**, 175–92.
- Valia, H. S. & Cameron, B. 1977. Skewness as a paleoenvironmental indicator. *Journal of Sedimentary Petrology* **47**, 784–93.
- Varnavas, S. P., Panagos, A. G. & Laios, G. 1987. Trace elements in surface sediments of Navarino Bay, Greece. *Environmental Geology and Water Sciences* **10**, 159–68.
- Zaïbi, C., Carbonel, P., Kamoun, F., Fontugne, M., Azri, C., Jedoui, Y. & Montacer, M. 2012. Evolution of the sebkha Dreïaa (South-Eastern Tunisia, Gulf of Gabes) during the Late Holocene: Response of ostracod assemblages. *Revue de Micropaléontologie* **55**, 83–97.

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