

## Research Article

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### Author for correspondence:

\*Niti Sukumal,

Email: [niti\\_230@hotmail.com](mailto:niti_230@hotmail.com)

The authors dedicate this paper to the memory of Atthapol Sunsernboon, ranger at HuaiOo Substation, DoiWiangLa Wildlife Sanctuary, who assisted during data collection.

# Assessing the status of the Green Peafowl's "expected stronghold" in dry forests along the Salawin River, north-west Thailand

Niti Sukumal<sup>1\*</sup> , Somying Thunhikorn<sup>2</sup> and Tommaso Savini<sup>1</sup> 

<sup>1</sup>Conservation Ecology Program, King Mongkut's University of Technology Thonburi, 49 Soi Thian Thale 25, Bang Khun Thian Chai Thale Road, Tha Kham, Bang Khun Thian, Bangkok 10150, Thailand and <sup>2</sup>Department of National Parks, Wildlife and Plant Conservation, 61 Phaholyothin Road, Chatuchak, Bangkok 10900, Thailand

## Summary

The suitable habitats of the Endangered Green Peafowl *Pavo muticus* have declined by 80% over the past century due to extensive anthropogenic degradation. Currently, only six strongholds remain in mainland South-east Asia. While there are estimates of the species' presence and status in five of these, the sixth one, defined as an "expected stronghold" located in the suitable dry forest along the Salawin River between Thailand and Myanmar, is not well documented. This study assessed the status of the area as a stronghold for the Green Peafowl's long-term survival by estimating 1) the population density, 2) the current extension of suitable habitats, and 3) the threats to its survival. The area had an estimated density of 0.27 calling males/km<sup>2</sup> (CI = 0.07–1.01) inhabiting 9,154 km<sup>2</sup> of a mosaic of forest types, including mixed dry pine, mixed deciduous, and dry dipterocarp forests. Higher estimates were reported in other strongholds, including 0.8 birds/km<sup>2</sup> in Bago Yoma (Myanmar), 15.8 birds/km<sup>2</sup> in north-east Thailand, 1.13–11.34 birds/km<sup>2</sup> in HuaiKhaKhaeng Wildlife Sanctuary (Thailand), 0.15–1.7 birds/km<sup>2</sup> in northern Cambodia, and 0.15–4.69 birds/km<sup>2</sup> in eastern Cambodia / south-central Vietnam. Hunting, habitat disturbance, and the presence of humans posed the greatest danger to the species across the surveyed area. Our results confirm the potential of the area for the species' long-term survival. However, this "expected stronghold," which could sustain a population of over 10 calling males/km<sup>2</sup> like other high-quality strongholds, is in dire need of a comprehensive management plan to help reduce anthropogenic pressure. Ultimately, transboundary management between Thailand and Myanmar is crucial for the long-term repopulation of this stronghold.

## Introduction

The 'Endangered' Green Peafowl *Pavo muticus*, a South-east Asian dry forest specialist (Brickle 2002) and a good indicator of the habitat status (Savini *et al.* 2021), has seen its historical range decline by more than 80% over the past century (BirdLife International 2018) due primarily to the extraction of commercially valuable timber species such as teak (Sodhi *et al.* 2004) and extensive agricultural expansion (Ratnam *et al.* 2016). Most of the remaining suitable habitats over the species' range, recently grouped into six strongholds, large areas of suitable habitats with a high probability of Green Peafowl occurrence (Sukumal *et al.* 2020a), remain under threat of habitat degradation (logging, excessive free-range grazing, and extended bushfires) and persistently high hunting pressure (Sukumal *et al.* 2015).

The species' population status has now been estimated in five of those strongholds: eastern Cambodia/south-central Vietnam (Sukumal *et al.* 2015, Nuttall *et al.* 2016, Tak *et al.* 2022), north-eastern Cambodia (Loveridge *et al.* 2017), western Thailand (Sukumal *et al.* 2017), northern Thailand (Saridnirun *et al.* 2021), and the Bago Yoma range in Myanmar (Lay Win *et al.* in review). Due to a lack of data on the species' presence and status, the sixth stronghold, located along the Salawin River and consisting of dry forest patches between the north-west of Thailand and the states of Kayah and Kayin in Myanmar, was classified as an "expected stronghold" based on the predicted high probability of its occurrence (Sukumal *et al.* 2020a). On the Thai side, the dry forest patches are covered by five protected areas considered highly suitable for the species, despite the history of human encroachment and logging concessions, especially before their official designation (Delang 2005). On the Myanmar side, a large amount of dry forest along the border falls partly under eight reserve forests, three wildlife sanctuaries and 34 community forests protected at the state level under the Salawin Peace Park initiative (Paul 2018). The rest of the area is covered by small-scale agriculture and villages.

Therefore, the aim of this study was to determine the current status of the sixth stronghold. We started by estimating the Green Peafowl's status in all five protected areas on the Thai side of

the Salawin River using line transect surveys and distance sampling, proven to be suitable for the species (Sukumal *et al.* 2015). Second, using remote-sensing data, we measured the extent of the remaining suitable habitat in light of the encroachment activities going on within protected areas. Third, using available data from Spatial Monitoring and Report Tool (SMART, see Methods), collected systematically by park rangers during patrols, for each of the protected areas investigated, we assessed the spatial distribution of threats to the species.

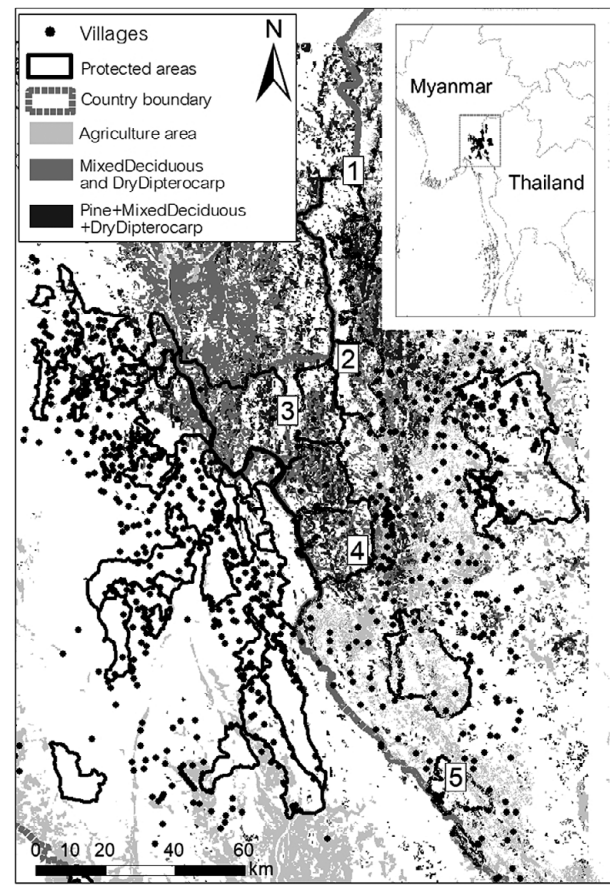
## Methods

### Study area

Overall, the study focused on dry forest habitats found along the Salawin River (also spelled as Salween River) in north-west Thailand and neighbouring Myanmar. The survey of Green Peafowl was conducted in five protected areas in north-western Thailand: 1) DoiWiangLa Wildlife Sanctuary (WS) ( $18^{\circ}54' \text{ N } 97^{\circ}54' \text{ E}$ ) covering a total area of  $467 \text{ km}^2$ , 2) MaeYuamFangKwa WS ( $18^{\circ}23' \text{ N } 97^{\circ}54' \text{ E}$ ),  $292 \text{ km}^2$ , 3) Salawin WS ( $18^{\circ}18' \text{ N } 97^{\circ}51' \text{ E}$ ),  $875 \text{ km}^2$ , 4) Salawin National Park (NP) ( $18^{\circ}10' \text{ N } 97^{\circ}44' \text{ E}$ ),  $721 \text{ km}^2$ , and 5) MaeMoei NP ( $17^{\circ}28' \text{ N } 98^{\circ}04' \text{ E}$ ),  $1,142 \text{ km}^2$ , with an altitudinal range of 200 to 1,000 m. The area is covered by dry dipterocarp forest and mixed deciduous forest, which includes teak and isolated patches of mostly evergreen forest at a higher elevation (Figure 1). Villages are found in relatively high numbers both around and within the protected areas (Figure 1). The area has a dry season from November to April and a wet season from May to October.

### Line transects surveys and density estimation

Green Peafowl density was estimated using 19 line transects set along accessible trails and roads (forest interior) in all five protected areas. Because of the difficulty in accessing the large study area, we focused on areas where Green Peafowl had been reported both by rangers and villagers in the last five years. At each selected site, we set 2–3 transects 500 m apart (from the end of one and the beginning of the other). We mainly focused on the density estimate for the whole study area but also provided stratified density estimates of each transect, bearing in mind the possibility of bias as some of the detections may originate from overlapping areas between the transects. Transects were monitored during the breeding season (January–March 2021) when the birds frequently call. Only vocal detections were recorded (i.e. not visual detections) within 1 km on both sides of the transects. Of the 19 transects, seven were set in DoiWiangLa WS, three in MaeYuamFangKwa WS, two in Salawin WS, four in Salawin NP, and three in MaeMoei NP. Each transect was 2 km in length with the exception of one transect of 1.5 km in MaeMoei NP and another 3 km transect in DoiWiangLa WS. Each transect was walked by different observers during the daily peak calling periods, 06h30–08h30 in the morning and 16h45–18h45 in the evening, for a total of six times per transect (twice a day for three consecutive days) (Sukumal *et al.* 2015, 2017). Observers walked the transects together at the beginning of the survey in order to standardize their data collection, in particular to estimate the distance to calling birds and minimize detection errors between observers. The distance to each calling bird was assigned to 100-m distance intervals, or 50-m distance intervals for closer records. Double counting of individual calling birds within the same location was corrected by removing calls from the same location during one survey to estimate the species' occurrence.



**Figure 1.** The study site, focusing on dry forests both inside and outside protected areas. The transect survey was conducted inside and around five protected areas in the north-west of Thailand, consisting of 1) DoiWiangLa Wildlife Sanctuary (WS), 2) MaeYuamFangKwa WS, 3) Salawin WS, 4) Salawin National Park (NP) and 5) MaeMoei NP.

Density was estimated using DISTANCE version 6.0 by pooling the detections along all transects in each protected area. The data were controlled for outliers, which were not found, before conducting the analysis. All key functions, namely uniform, half-normal, hazard rate, and negative exponential functions, with series adjustments, cosine, simple polynomial, and hermite polynomial, were examined to select the best detection function and model with the lowest AIC following Buckland *et al.* (2001).

### Defining suitable habitat for Green Peafowl and areas with high predicted occupancy

We predicted the probability of Green Peafowl occurrence over our study site by using species distribution modelling techniques, which associate environmental variables with known species' occurrence records to identify environmental conditions that are suitable for the species, enabling the identification of suitable environments in space and estimation of the species' probability of occurrence across a study region. Species distribution modelling requires two types of data: 1) biological data that describe the known distribution or occurrence localities of a species, mostly obtained from occurrence records or field surveys, and 2) environmental data consisting of either continuous or categorical data within a certain range (Pearson 2007). We gathered location records of Green Peafowl from both villagers and patrolling rangers, combined them with

**Table 1.** Density estimates for Green Peafowl in the north-eastern part of DoiWiangLa Wildlife Sanctuary during the breeding season of February–March 2021.

Site	Transect ID	Total length (km)	Survey effort (transect walk frequency)	Effective strip width (ESW)	Density estimates (calling birds/km <sup>2</sup> )	95% CI	Number detected
Global estimate	all 19 Transects	38.5	118	386 m	0.27	0.07–1.01	49
DoiWiangLa WS	Total	5	8	347 m	1.76	0.17–17.76	49
	Transect 1	3	8	527 m	0.71	0.51–0.98	18
	Transect 2	2	8	426 m	2.27	1.94–2.64	31

data from the line transect survey using the ArcGIS 10.3.1 program (Esri, Redlands, USA), and plotted the distance and bearing of calling birds along the transects. In total, 19 locations were recorded, six from villagers, four from patrolling rangers and nine from our line transects. For each location point, we created a 500-m radius circular plot based on an average of the effective strip width (ESW) of detection along the transects (see Table 1), obtained from density estimation using the DISTANCE 6.0 program (Thomas *et al.* 2010). When two circular plots overlapped by >5%, we selected only the location of the latest record to represent the area. We finally derived 13 locations for habitat assessment and probability of occurrence modelling over our study site. We constructed the probability of occurrence models using a generalized linear model (GLM) with logit link and binomial error distribution to study the relationship between a given variable and the probability of occurrence, using the presence/absence of Green Peafowl as the response variable. Presence data were derived from 13 locations of Green Peafowl recording, while absence data were derived from random locations along the transects, at least 1 km apart, where we did not detect any Green Peafowl during the survey. We created 500-m radius circular plots around the presence and absence locations and determined the habitat types, altitude, distance from the center to the village, and distance from the center to the country's border, which were used as the predictor environmental variables. Habitat types were obtained from Climate Change Land Cover (CCI-LC) year 2000, downloaded from <http://maps.elie.ucl.ac.be/CCI/viewer/download.php>. We started with 38 original habitat types; firstly we eliminated unsuitable habitats such as bare areas, urban areas, water bodies, and secondly we reclassified and renamed the remaining 19 habitat types into 14 major habitat types (Table S1 in the online supplementary material). We combined the habitat type map with the land use type map 2018, obtained from Land Development Department, Thailand (available at <http://dinonline.ldd.go.th/Default.aspx>), to update the remaining 14 habitat types. All predictor variables were tested for multicollinearity using a pairwise-correlation matrix (Spearman rho,  $\rho$ ). All the variables had a correlation coefficient of <0.6 (Zuur *et al.* 2010) and were included in the model. Each predictor variable was standardized by dividing the value by the standard deviation in order to transform the data to the same scale. The best GLM model was selected based on the lowest Akaike's Information Criterion (AIC) value (Burnham and Anderson 1998). A confidence interval of 85% was used to consider variables influencing the probability of peafowl occurrence. The analysis was conducted using program R (R Development Core Team 2014).

### Spatial distribution of threats

Threats to Green Peafowl were defined based on data entered in the Spatial Monitoring and Reporting Tool (SMART) available for the

five protected areas. SMART is an integrated conservation management system using data collected by rangers while patrolling to inform management plans in each protected area (Cronin *et al.* 2021). Data were collected at least twice a month on defined routes covering over 90% of the protected area by rangers on patrol during 2019–2021 and other records of threats along patrolling routes within the protected areas between October 2013 and January 2017. We categorized threats to Green Peafowl into three categories: 1) Activities linked with hunting, which included direct hunting records and presence of camping sites; 2) habitat disturbance, which included free-ranging cattle, logging, and forest fire; and 3) presence of humans, which included data on non-timber forest product (NTFP) collection, presence of human tracks, garbage dumps, and fishing.

## Results

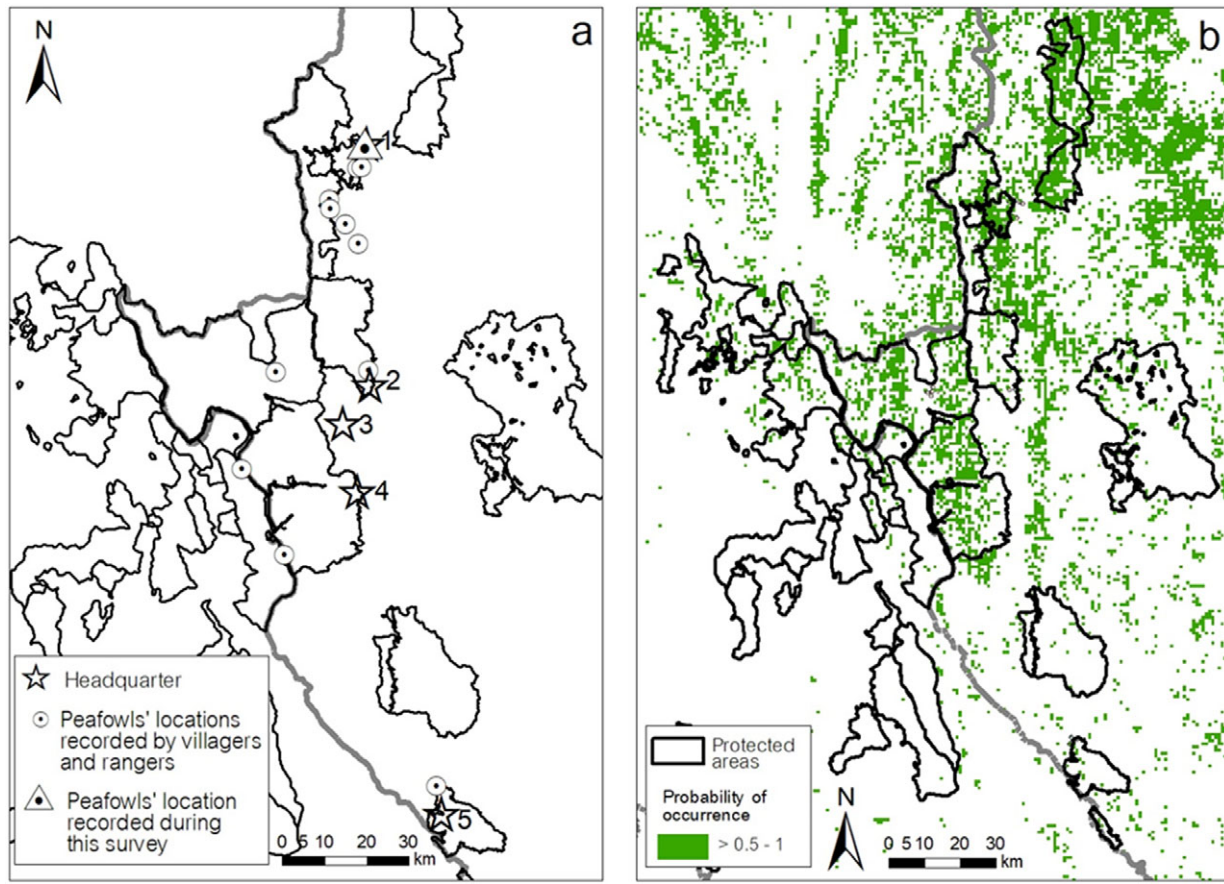
### Density estimation

A total of 49 detections were recorded after walking 19 transects 118 times, totalling 241 km (See triangle in Figure 2a). The density estimate for the whole area was of 0.27 calling males/km<sup>2</sup> (CI = 0.07–1.01). However, we detected Green Peafowl only along the two transects in DoiWiangLa WS, namely one 3-km transect denoted by transect 1 and one 2-km transect hereafter denoted by transect 2. To meet the recommended minimum number of 40 detections for reliable density estimation (Buckland *et al.* 2001), both transects were walked eight times rather than six times as we did for other transects. From the 49 independent calling birds detected after eliminating double counts, 18 detections were found on transect 1 and 31 on transect 2. The overall density estimated for this protected area was 1.76 calling males/km<sup>2</sup> (CI = 0.17–17.76), while the stratified density estimates for transects 1 and 2 were 0.71 calling males/km<sup>2</sup> (CI = 0.51–0.98) and 2.27 calling males/km<sup>2</sup> (CI = 1.94–2.64), respectively (Table 1).

Since January 2019, a total of 10 locations were recorded outside the survey period based on reliable reports from villagers (photos were provided) and rangers (See circles in Figure 2a). Eight detections were within protected areas, while two were in the agricultural landscape surrounding them.

### Remaining dry forest habitat over study site

Over the range of our study site, the remaining dry forests covered an area of 9,154 km<sup>2</sup> and comprised mixed deciduous and dry dipterocarp forests, 5,005 km<sup>2</sup>, and pine with mixed deciduous and dry dipterocarp forests, 4,149 km<sup>2</sup>. Only 1,951 km<sup>2</sup> or 39% of the dry forests over our study site were in protected areas both in Thailand and Myanmar (Figure 1).



**Figure 2.** a) Recorded locations of Green Peafowl during the survey (triangle with dot as the center), and outside the survey period by villagers and rangers; b) predicted probability of occurrence of Green Peafowl from modelling, based on line transect surveys and detection records of Green Peafowl in the study site.

### Suitable habitat and probability of occurrence

The best model (lowest AIC) of suitable habitat had only one variable of dry forest containing pine, mixed deciduous (including teak) and dry dipterocarp forests. The probability of occurrence was higher when the area of pine, mixed deciduous and dry dipterocarp forests increased (Table 2).

The predicted suitable habitat (probability of occurrence >0.5) for Green Peafowl over our study site both in Thailand and Myanmar (17°09'–19°23' N, 96°54'–98°28' E) was 5,693 km<sup>2</sup> in size, and just 22% of this area (1,275 km<sup>2</sup>) fell inside protected areas (Figure 2b).

### Threats to Green Peafowl

The threat data revealed an extensive distribution of threats to the species over the five protected areas (Figure 3a, b, c); habitat disturbance presented the highest number of threats with 795 records, followed by 649 records of hunting and 191 records of human presence.

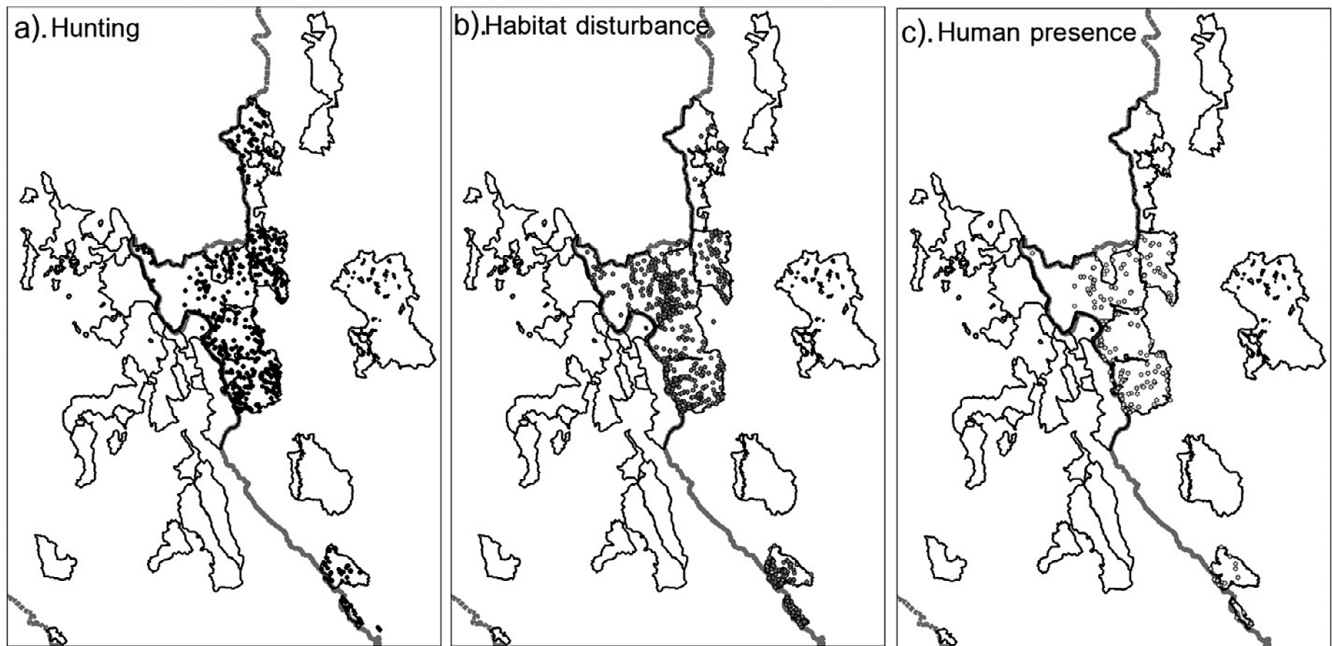
### Discussion

Overall, the detections of Green Peafowl within the five protected areas were low resulting in a low estimated density of 0.27 calling males/km<sup>2</sup>. This density estimate mirrors that of YokDon National Park in south-central Vietnam (0.25 calling males/km<sup>2</sup>), considered

**Table 2.** Parameters that influence the probability of occurrence model based on line transect surveys and detection records of Green Peafowl in the study site.

Parameters	$\beta$	SE	Lower 85 % CI	Upper 85 % CI
Pine with mixed deciduous and dry dipterocarp forest	0.78	0.38	0.26	1.39

heavily disturbed primarily by extensive fire and high cattle grazing (Sukumal *et al.* 2015). Although some villagers and rangers reported encountering the birds nearby during the survey, we were unable to detect any calling males along most of the transects, except for a cluster found in DoiWiangLa WS, most likely due to a combination of a low population number and high disturbance from human settlements (Swaddle *et al.* 2015). The Bago Yoma range in central Myanmar had a density estimate of 0.8 calling males/km<sup>2</sup> in a dry forest mixed with teak (Lay Win *et al.* in review). DoiWiangLa WS alone had an estimated density of 1.76 calling birds/km<sup>2</sup> (ranging from 0.71 to 2.27), within the range (1.13–11.34 calling males/km<sup>2</sup>) reported for the dry dipterocarp and mixed deciduous forests in the species' stronghold of HuaiKhaKhaeng Wildlife Sanctuary in western Thailand (Sukumal *et al.* 2017). However, the value is much lower than the density of 13.55–19.87 calling males/km<sup>2</sup> estimated in a similar mixed habitat



**Figure 3.** Occurrence of different disturbance types across all five protected areas: a) direct hunting records and presence of camping sites; b) free-ranging cattle, logging, and forest fire; and c) presence of non-timber forest product (NTFP) collection, presence of human tracks, garbage dumps, and fishing.

structure in the nearby northern Thailand stronghold (Saridnirun *et al.* 2021). The relatively high density recorded in the northern stronghold could be a natural consequence of the relatively few predators found in the area (G. Saridnirun unpubl. data). As predators are also rare in our study site, the low estimated density could be a consequence of ongoing anthropogenic disturbance in the area.

Suitable dry forest habitats along the Salawin River cover an area of 9,154 km<sup>2</sup> and comprise all the main habitat types selected by the species over its range, in particular mixed deciduous, including the suitable mixed teak forest, and dry dipterocarp forests (5,005 km<sup>2</sup>) and pine with mixed deciduous and dry dipterocarp forests (4,149 km<sup>2</sup>). The size and habitat composition of the area can potentially hold a Green Peafowl population large enough to guarantee its long-term survival (Sukumal *et al.* 2020b), notwithstanding the surrounding agricultural landscape (Figure 1). In the northern stronghold, Saridnirun *et al.* (2021) estimated 15 calling birds/km<sup>2</sup> in the dry dipterocarp forest, 19 calling birds/km<sup>2</sup> in the mixed deciduous forest, and 24 calling birds/km<sup>2</sup> in the mixed pine forest similarly surrounded by agricultural landscapes encroaching on the protected areas (Figure 1). With effective management, farmland in proximity to the natural habitat can hold viable populations, with densities ranging from 1.83 birds/km<sup>2</sup> in areas surrounding small forest fragments (Shwe *et al.* 2021) to 14.29 calling males/km<sup>2</sup> in areas surrounding large continuous forest patches (Saridnirun *et al.* 2021).

Our results show clearly the extensive distribution of threats such as habitat disturbance, especially by free-ranging cattle, logging, forest fire, and high hunting pressure over the species' suitable dry forest habitats. Habitat disturbance and hunting pressure can drastically reduce Green Peafowl populations even in highly suitable habitats. In the 1,155 km<sup>2</sup> of dry dipterocarp forest in YokDon National Park, southcentral Vietnam, the estimated density was as low as 0.25 calling males/km<sup>2</sup> (ranging from 0.03 to 0.7) due to heavy grazing, forest fire, and hunting (Sukumal *et al.* 2015).

#### Importance of the area as a stronghold for Green Peafowl

The area was defined as an “expected stronghold” for Green Peafowl by Sukumal *et al.* (2020a) due to the large extent of suitable habitats and the high predicted probability of occurrence of the species therein, despite a lack of prior confirmation of its presence in the area. The extent of suitable habitat is considered a major limiting factor for the species' long-term survival. Green Peafowl is currently listed as ‘Endangered’ mainly due to habitat loss as a consequence of agricultural expansion over its range (BirdLife International 2018). Despite the relatively low estimated density, most likely a consequence of ongoing anthropogenic disturbance, the amount of suitable habitat, the main factor limiting the species over its whole range, arguably justifies considering the area along the Salawin River as an area with a good opportunity for the Green Peafowl population recovery. However, as management implementation is urgently needed, we can currently suggest the area be considered as a “potential stronghold pending management implementation.” Green Peafowl have shown high resilience to extreme population decline by recovering to large numbers within a relatively small amount of time once disturbances are limited with management (Sukumal *et al.* 2015, 2017). Reducing anthropogenic pressure in the area could therefore guarantee the recovery of the species to the level of >10 calling males/km<sup>2</sup>, as estimated in well-established populations in similar habitats (Sukumal *et al.* 2017; Saridnirun *et al.* 2021), which could guarantee its long-term survival (Sukumal *et al.* 2020b). For instance, focusing on the Green Peafowl populations around the headquarters of DoiWiangLa WS (number 1 in Figure 2a) and MaeYuamFangKwa WS (number 2 in Figure 2a) might facilitate its recovery and expansion to other protected areas nearby. The recovery recorded in HuaiKhaKhaeng Wildlife Sanctuary shows the species' ability to disperse from a source population and repopulate areas up to 15 km away following an increase in the protection level (Simcharoen *et al.* 1995; Sukumal *et al.* 2017). Data on sporadic detections during the actual survey

showed dispersing juveniles up to 22 km away from the only recorded large population (number 1 in Figure 2a). Moreover, the species can adapt well to agricultural areas around the stronghold if well managed. A density estimate of 1.13–2.63 males/km<sup>2</sup> was reported in agricultural fields surrounding a 30-ha forest fragment protected by resident monks (Shwe et al. 2021), whereas 14.2 calling males/km<sup>2</sup> were estimated in an agricultural landscape surrounding protected areas managed by local communities for ecotourism (Saridnirun et al. 2021).

Myanmar and Thai authorities should take concerted actions to protect and manage the transboundary dispersal of the species, which could help its recovery, especially on the Thai side. Sporadic detections, probably, of birds dispersing from the Myanmar side occurred near the border (i.e. the Salawin River) in areas where no resident populations were detected in this survey. Reducing anthropogenic impact in the area by limiting agricultural expansion, cattle grazing, and fire and preventing poaching could increase the population in the mid- to long-term. Unfortunately, data on the species' presence on the Myanmar side are sparse and only available as fragmented secondary data from general wildlife surveys. Due to recent political instability, specific surveys cannot hold in the near future, and implementation of any management activity may prove challenging.

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**Supplementary Material.** To view supplementary material for this article, please visit <https://doi.org/10.1017/S0959270922000338>.

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