

# Indirect evidence of pathogen-associated altered oocyte production in queens of the invasive yellow crazy ant, *Anoplolepis gracilipes*, in Arnhem Land, Australia

M.D. Cooling<sup>1\*</sup>, B.D. Hoffmann<sup>2</sup>, M.A.M. Gruber<sup>1</sup> and P.J. Lester<sup>1</sup>

<sup>1</sup>School of Biological Sciences, Victoria University of Wellington, PO Box 600, Wellington 6140, New Zealand; <sup>2</sup>CSIRO, Tropical Ecosystems Research Centre, PMB 44, Winnellie, NT 0822, Australia

## Abstract

*Anoplolepis gracilipes* is one of the six most widespread and pestiferous invasive ant species. Populations of this invader in Arnhem Land, Australia have been observed to decline, but the reasons behind these declines are not known. We investigated if there is evidence of a pathogen that could be responsible for killing ant queens or affecting their reproductive output. We measured queen number per nest, fecundity and fat content of queens from *A. gracilipes* populations in various stages of decline or expansion. We found no significant difference in any of these variables among populations. However, 23% of queens were found to have melanized nodules, a cellular immune response, in their ovaries and fat bodies. The melanized nodules found in dissected queens are highly likely to indicate the presence of pathogens or parasites capable of infecting *A. gracilipes*. Queens with nodules had significantly fewer oocytes in their ovaries, but nodule presence was not associated with low ant population abundances. Although the microorganism responsible for the nodules is as yet unidentified, this is the first evidence of the presence of a pathogenic microorganism in the invasive ant *A. gracilipes* that may be affecting reproduction.

**Keywords:** invasive species, melanized nodules, biological control

(Accepted 23 August 2017; First published online 18 September 2017)

## Introduction

Insects are host to diverse microbial communities (Dillon & Dillon, 2004; Anderson *et al.*, 2011) that include a wide range of pathogenic taxa (Schmid-Hempel, 1998). Pathogens can have a range of effects on their insect hosts, including reduced reproduction (Calleri *et al.*, 2006; Dunn *et al.*, 2012). Symptoms include low-energy reserves and reduced fertility. Low-fat content alone can impact survival. Ants with a lower fat content, for example, have been found to be less able to survive physiologically stressful conditions (Elmes *et al.*, 1999). Fat

content also plays a role in immunity and is thus one indicator of general health (Fellous & Lazzaro, 2010).

In many social insects, the queen or queens are the sole reproducing females in the colony. They alone produce workers that are responsible for all the other tasks essential for colony functioning such as brood care, colony defense and foraging (Alaux *et al.*, 2011). In polygynous ant species, population size and spread is usually positively correlated with queen number (Ingram, 2002a, b; McGlynn, 2010). Therefore, low queen numbers in polygyne colonies may constrain colony growth and expansion (Nonacs, 1991). Low queen numbers may be due to low queen production or relatively high queen mortality.

In social insect societies, pathogens can affect colony health via the queen, either by killing the queen directly, or negatively affecting her reproduction. For example, the microsporidian, *Kneallhazia solenopsae*, has been shown to cause a

\*Author for correspondence:

Phone: +1 (778) 872-8328

Fax: +1 (604) 926-1891

E-mail: [meghan.cooling@gmail.com](mailto:meghan.cooling@gmail.com)

reduction in colony size in fire ants, potentially by destroying the fat bodies of queens, and thereby reducing queen fertility (Briano *et al.*, 1995a). In *Apis mellifera* colonies in Australia, the most common cause of death of larval queens is the *Black queen cell virus* (Nguyen *et al.*, 2011). Similarly, Valles *et al.* (2013) found *Solenopsis invicta* queens infected with *S. invicta virus 3* (SINV-3) had fewer eggs in their ovaries than uninfected queens. Clearly there are a wide variety of pathogens that can influence the reproductive rates of social insects.

Social insects, like other social animals, are thought to be particularly vulnerable to disease because their extremely close living conditions may increase pathogen transmission rates (Schmid-Hempel, 1998). Pathogens affecting social insects are especially well studied in the honey bee, *A. mellifera*, because of the myriad of pathogens associated with colony collapse disorder (Cox-Foster *et al.*, 2007; Oldroyd, 2007; vanEngelsdorp *et al.*, 2009). Other hymenopteran species may also harbor many pathogens and parasites that affect their population dynamics (Briano *et al.*, 1995b; Briano, 2005; Cameron *et al.*, 2011). Pathogens of invasive species are of particular interest because of their potential role in biocontrol (Oi & Valles, 2009; Valles *et al.*, 2012; Lester *et al.*, 2015; Sébastien *et al.*, 2015; Gruber *et al.*, 2017). To date, the most well-studied invasive ant species, *S. invicta*, has been found to host over 40 species of pathogens and parasites (Oi & Valles, 2009).

The yellow crazy ant, *Anoplolepis gracilipes* (F. Smith), is one of the six most widespread, abundant and damaging invasive ants globally (Holway *et al.*, 2002). It is prevalent in the Pacific region and is thought to most likely be native to Southeast Asia (Wetterer, 2005; Drescher *et al.*, 2011). Multiple studies have identified numerous microorganisms associated with this ant (Sébastien *et al.*, 2011; Gruber, 2012), and several pathogens have recently been detected (Cooling *et al.*, 2016), but the effect these pathogens may have on the host population has yet to be documented. Population crashes or declines have been observed in this species, without obvious causes (Haines & Haines, 1978; Gruber *et al.*, 2012; Cooling & Hoffmann, 2015), though pathogens have been suggested (Gruber, 2012; Cooling & Hoffmann, 2015). The nature of these observed population declines, the lack of worker symptoms and the slow rate of decline led to the hypothesis that if a pathogen was involved, it was likely to be affecting the queens either through direct mortality and/or reducing their reproductive output (Cooling & Hoffmann, 2015). Here we investigated if there is evidence of a pathogen that could be responsible for killing ant queens or affecting their reproductive output in *A. gracilipes* populations in Arnhem Land, Australia. Specifically we compared: (1) queen number, (2) queen fecundity and (3) the queen fat content among populations of varying worker abundance levels. We hypothesized that low- and medium-abundance populations would have significantly fewer queens per nest, and associated queens would have a lower reproductive output and reduced energy reserves than queens from high-abundance populations.

## Materials and methods

### Population selection

We used six populations classed into three population types (low, medium and high abundance), previously defined from a combination of spatial surveys, population abundances and historical observations (Cooling & Hoffmann, 2015; Cooling, 2016; Cooling *et al.*, 2016). See Supplementary

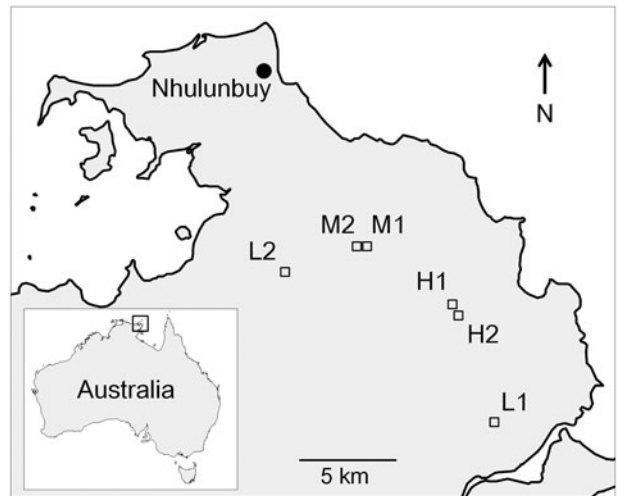


Fig. 1. Study site locations (red squares) in NE Arnhem Land. L, M and H denote low-, medium- and high-abundance populations, respectively. Sites M1 and M2 are 450 m apart and H1 and H2 are 400 m apart and are discrete populations.

Material S1 for details of the surveying process and surveyed population maps. Population selection was determined by both *A. gracilipes* worker abundance in each population at the time of sampling and each population's history of decline or expansion. The methods used to determine abundance are described in detail in Cooling *et al.* (2016). Briefly, we selected four populations where *A. gracilipes* abundance was in various stages of decline in the absence of any human intervention (fig. 1) (low- and medium-abundance populations), and two that were expanding (high-abundance populations). The two low-abundance populations (L1, L2) were remnants of much larger populations that had declined substantially in the previous 5 years (Cooling & Hoffmann, 2015; Cooling *et al.*, 2016). Worker abundances at L1 and L2 were  $31.75 \pm 4.78$  workers (mean of eight plots, each consisting of five pooled pitfall trap samples  $\pm$  standard error) and  $21 \pm 10$  workers, respectively. The two medium-abundance populations (M1, M2) had higher abundances (mean  $\pm$  SE:  $80 \pm 22$  workers and  $57 \pm 9$  workers, respectively) than the low-abundance populations and covered a wider area. These populations were stable, but belonged to populations that were previously larger and had been observed to decline (Cooling *et al.*, 2016). The two high-abundance populations (H1, H2) were expanding, had abundances two orders of magnitude higher (mean  $\pm$  SE:  $2749 \pm 327$  and  $1401 \pm 264$ , respectively) than the low- and medium-abundance populations, and covered large areas (Cooling *et al.*, 2016).

### Nest and queen number

Queens were collected in February 2013 and 2015. Each year, three 30 m transects spaced a minimum of 30 m apart were haphazardly placed in each population. The length of each transect extending out 1 m on one side ( $30 \text{ m}^2$ ) was searched for queens by scraping away litter with a trowel and moving rocks and logs. Every *A. gracilipes* nest found was excavated and the queens were collected, counted and the nest they were from recorded. Because this ant does not

utilize persistent nest sites and individuals regularly move among nests (Hoffmann, 2014), a nest was defined as where workers and eggs, larvae, or pupae were present. A congregation of workers without the presence of eggs, larvae, or pupae was not considered a nest. A nest was considered completely excavated when the flow of workers exiting the nest had stopped and no queens had been found for at least 10 min. Nests within the same transect were considered separate from each other when there were no underground connections between chambers housing workers, eggs, larvae, pupae, or queens. To account for the effect of environmental differences among populations, at each nest and every 5 m mark, starting at 0 m along a transect, the following parameters were measured: canopy cover (%) (estimated visually), litter cover (%) (estimated visually), number of *Acacia* plants over 0.5 m tall, number of *Eucalyptus* plants over 0.5 m tall, rock cover (%), deadwood cover (%) and litter depth (average of three measurements) in a 1 m<sup>2</sup> area with the nest at the center. In populations where no, or very few, nests were found along transects, the surrounding area was searched intensively, using the same methods as above of scraping away litter with a trowel to reveal nest entrances and looking under rocks and logs, to locate additional nests. The environmental measures associated with these nests were measured in the same manner as described above. Queens were chilled in a refrigerator at 4°C for 1.5 h, then placed into RNAlater<sup>®</sup> Stabilization Solution (Life Technologies, Auckland, New Zealand) alive for optimal preservation (RNAlater Handbook, 2006). The vials of RNAlater<sup>®</sup> were left at room temperature for 24 h before being placed in a freezer to allow the liquid to permeate the tissue. In order to increase replication, an additional 16 queens were randomly collected in February 2014 (eight from L1, eight from M2). These additional queens were included in the potential fecundity analysis only.

Generalized mixed effect models (GLMMs) with Poisson distributions were used to test if any of the environmental variables affected queen number per nest and nest number. Population, transect and year were random effects, with transect nested within population. Population type (low-, medium- or high-abundance population), canopy cover, litter depth, litter cover, number of *Acacia* and number of *Eucalyptus* were fixed effects. We did not use the sequential Bonferroni technique to correct for multiple comparisons, as it has been shown to substantially increase the chance of a making a type II error (failing to reject a false null hypothesis) (Moran, 2003). GLMMs would be inappropriate if the data were spatially correlated; however, Mantel tests (9999 permutations) indicated that there was no spatial autocorrelation in the data (queen number:  $R^2 = 0.062$ ,  $P = 0.086$ ; nest number:  $R^2 = -0.011$ ,  $P = 0.893$ ).

All data analyses were performed using R v 3.0.2 (R Core Team, 2013). The package 'ade4' was used for the Mantel tests (Dray & Dufour, 2007). The package 'lme4' was used for the GLMMs (Bates *et al.*, 2014). *Post-hoc* comparisons used the package 'lsmeans' (Russell *et al.*, 2015).

#### Potential fecundity

To investigate reproductive output among population type, 278 queens (74 from low-abundance; 103 from medium-abundance; 101 from high-abundance populations; 143 from 2013; 16 from 2014; 119 from 2015) were dissected and ovariole number, the number of mature oocytes and

length of longest basal oocyte (to the nearest 10 µm) counted. These variables are indicators of egg-laying rates in ants (Dalecky *et al.*, 2005). The fat bodies were removed, placed in water and frozen for fat content analysis. Before dissection, head width and gaster width were measured to the nearest 0.05 mm with an ocular micrometer mounted on a dissecting scope. Head width is a standard and accurate measure of overall body size (Hölldobler & Wilson, 1990; Kaspari, 1993) and gaster width is a measure of egg-laying capacity in ants (Dalecky *et al.*, 2005). While dissecting, it was discovered that some queens had melanized nodules, a cellular immune response in insects (Stanley *et al.*, 2012), in their ovaries and fat bodies (fig. 2). A Poisson GLMM was used to test which variables affected oocyte number. Population, nest, year, queen number (how many queens were present in the original nest), head width, gaster width, ovariole number and length of longest basal oocyte were random effects, with the nest variable nested within population. Nodulation status (whether or not a queen had nodules), population type and their interaction term were fixed effects. *Post-hoc* comparisons were performed on population type and the interaction term. When interactions are present, comparisons can be difficult to extract from GLMMs as significance is measured relative to the lowest level of each factor (the intercept). To overcome this limitation, we undertook pairwise tests of significance using least-square means. However, as least-square means compares the means of terms to each other instead of the intercept, the GLMM interaction term results may differ slightly from those of the *post-hoc* comparisons.

#### Egg laying

We hypothesized queens from high-abundance populations would have higher egg-laying rates than queens from medium- and low-abundance populations. To test this hypothesis, we experimentally compared egg-laying rates between queens from the different populations. Egg laying was quantified in February 2013 using the maximum number of queens that were found, being ten and four queens from the two low-abundance populations, and 20 from each of the high- and medium-abundance populations. The egg-laying rate was measured in queens by placing one queen and three workers in a collection jar (5.5 cm high and 4.5 cm across) containing a 1 cm thick layer of damp plaster. The jar sides were coated with Insect-A-Slip (BioQuip Products, USA) to prevent ants escaping, and the outside of the jar wrapped in white paper to provide a dark environment. Ants were maintained without food, at approximately 25 ± 2°C, and in constant darkness to simulate conditions inside the nest. Ants were placed in the jars within 3 h of collection. After 24 h, the ants were removed and eggs counted. Preliminary tests showed that worker presence was critical for egg laying to occur and that 24 h was sufficient to observe egg laying. Queens were later dissected and examined for nodules, but because few egg-laying queens had nodules (18/96), analyses were restricted to egg laying by population type (i.e., not compared between queens with and without nodules within each population type) and egg laying between queens with and without nodules pooled from each population type. The number of eggs laid by queens among the three different population types was compared using a Kruskal–Wallis *H* test because the data were not normally distributed. Likewise, the number of eggs laid by queens with and without nodules was compared using a Mann–Whitney *U* test.

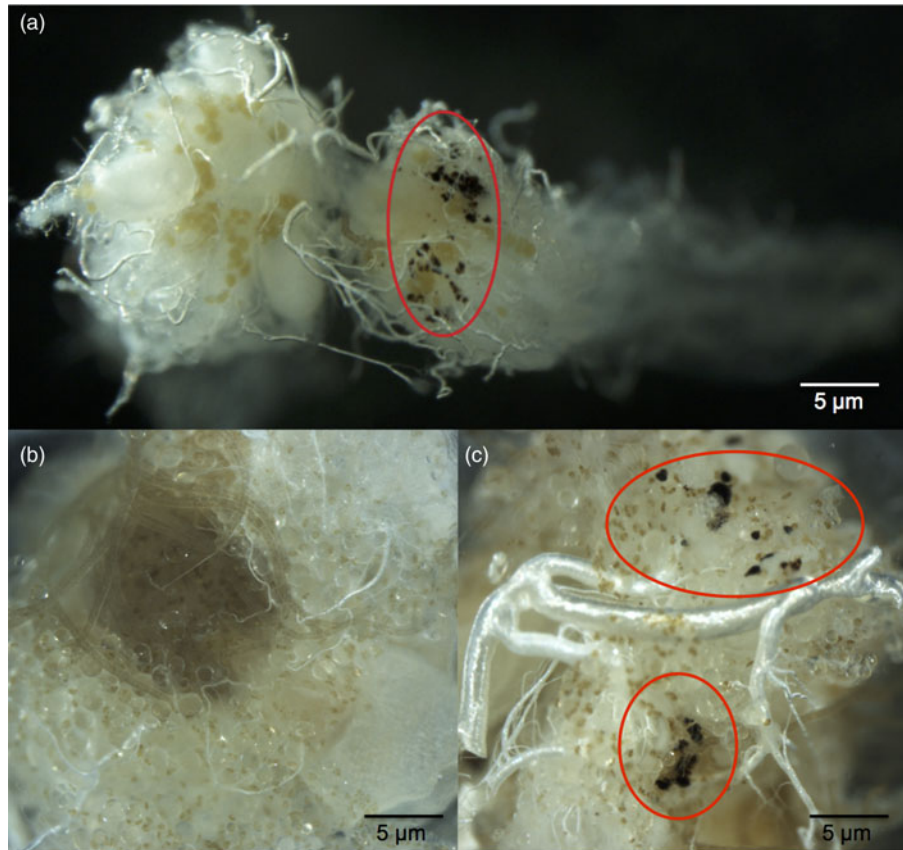


Fig. 2. (a) Ovaries from a queen with nodules. The ovary on the right contains melanized nodules (circled in red), the ovary on the left does not. (b) Fat bodies of a queen without nodules. (c) Nodules (circled in red) induced in the fat bodies of a queen by the authors, using spores of the entomopathogenic fungus, *Metarhizium anisopliae*. See Supplementary Material S1 for the injection process methods.

### Fat content

We measured queen fat content collected in February 2015 using a petroleum ether extraction following the methods of Richards & Packer (1994). During queen dissections (discussed above), the fat bodies were isolated from the queen, collected in water and frozen at  $-20^{\circ}\text{C}$  until processing. The fat samples were then placed in a drying oven at approximately  $65^{\circ}\text{C}$  for 10 days until their weight stabilized. The samples were weighed on a Kern ABT 220-4 M analytical balance to the nearest ten-thousandth of a gram four times over 2 days. After that period, 2 ml of petroleum ether was added to each sample. Samples were left for 5 days, at which point the fat had dissolved into the petroleum ether. The remaining petroleum ether was decanted from each sample into another container to remove the dissolved fat. Samples were then placed in a drying oven at  $65^{\circ}\text{C}$  for a further 10 days. Samples were then reweighed, with the difference between the initial and final weight taken to be the amount of fat each queen possessed. A GLMM with a  $\gamma$  distribution and log-link function was used to test which variables affected fat content. Population, nest, queen number per nest and head width were random effects, with nest nested within population. Nodulation status, population type and their interaction term were fixed effects.

### Results

#### Nest and queen number

A total of 284 queens were collected from 61% of the 171 nests excavated (table 1). Fifty percent of nests had fewer than four queens. Nest number varied significantly with population type (medium abundance vs. high abundance  $z = -3.060$ ,  $P = 0.002$ ; low abundance vs. high abundance  $z = -4.981$ ,  $P < 0.001$ ) with high-abundance populations having the most nests and low-abundance populations having the fewest (tables 1 and 2). Population types had significantly different queen numbers (medium abundance vs. high abundance  $z = -0.444$ ,  $P = 0.657$ ; low abundance vs. high abundance  $z = 3.077$ ,  $P = 0.002$ ; table 2), with low-abundance populations having higher numbers of queens per nest. There was a clear negative relationship between nest number and queen number, with 19% of nests that had queens in low-abundance populations had five or more queens, medium-abundance populations had 13% and high-abundance populations had only 7%. The number of *Acacia* was the only environmental variable with a significant, positive association with nest presence. Queen number was significantly and positively associated with canopy cover, litter cover and the number of *Acacia* (table 2).



Table 1. Population type, site, the percentage of dissected queens with the total number of queens dissected in brackets, total number of nests excavated and nest density per site.

Population type	Site	% queens with nodules (queens dissected)	Total nests excavated	Nests/10 m <sup>2</sup> ± SD
Low abundance	L1	18 (34)	13	<0.01 <sup>1</sup>
Low abundance	L2	5 (40)	14	<0.01 <sup>1</sup>
Medium abundance	M1	3 (30)	21	0.9 ± 0.9
Medium abundance	M2	47 (73)	33	1.0 ± 0.8
High abundance	H1	23 (57)	43	2.6 ± 1.1
High abundance	H2	20 (44)	47	2.4 ± 0.8
Total	–	23 (278)	171	–

<sup>1</sup>Nest density could not be accurately calculated due to paucity of nests.

Table 2. (A) Results of a generalized mixed effect model (GLMM) testing the effect of population type, canopy cover, litter depth and cover, *Acacia* number and *Eucalyptus* number on nest presence and number of queens per nest. Random effects were site, year and transect. The least square means post hoc tests for population type are reported after the GLMM results. (B) Results of a GLMM testing the effect of population type, nodulation status (status), and their interaction on oocyte number and fat content. Random effects were site, year, nest, gaster width, head width, queen number per nest, ovariole number and longest basal oocyte.

(A)	Nest presence			Queen #		
	β <sup>1</sup>	SE <sup>2</sup>	P <sup>3</sup>	β	SE	P
Intercept	–1.683	0.280	***	–1.141	0.431	**
Population type						
Medium	–0.616	0.201	**	1.153	0.372	NS
Low	–1.836	0.369	***	0.311	0.112	**
Canopy cover	0.273	0.300	NS	0.007	0.003	*
Litter depth	0.037	0.024	NS	0.014	0.028	NS
Litter cover	0.502	0.327	NS	0.010	0.004	*
<i>Acacia</i> #	0.101	0.043	*	0.095	0.033	**
<i>Eucalyptus</i> #	0.076	0.047	NS	0.021	0.051	NS
Population type post hoc test						
High vs. medium	1.851	0.373	**	1.153	0.372	NS
High vs. low	6.281	2.317	***	0.311	0.112	**
Medium vs. low	3.393	1.333	**	0.269	0.091	***
(B)	Oocyte #			Fat		
	β	SE	P	β	SE	P
Intercept	3.012	0.414	***	–5.700	0.056	***
Population type						
Medium	0.148	0.408	NS	0.023	0.089	NS
Low	–0.111	0.424	NS	0.104	0.088	NS
Status	–0.1.89	0.073	**	–0.002	0.082	NS
Population type × status						
Medium × with nodules	–0.224	0.100	*	0.062	0.157	NS
Low × with nodules	0.362	0.142	*	0.007	0.212	NS

<sup>1</sup>β coefficient.

<sup>2</sup>Standard error.

<sup>3</sup>Statistical significance.

\**P* < 0.05; \*\**P* < 0.01; \*\*\**P* < 0.001.

### Potential fecundity

A total of 278 queens were dissected, of which 23% had melanized nodules (table 1). More than half the queens with nodules (34/65) were from population M2, which was a medium-abundance population (fig. 1). Oocyte number was significantly negatively associated with the presence of nodules ( $z = -2.585$ ,  $P = 0.010$ ) (table 2, fig. 3a). Across all populations, queens without nodules had on average 22% more oocytes in their ovaries than queens with nodules (mean ± SE: 30.73 ± 1.43; 23.97 ± 2.36). Oocyte number did not vary

significantly with population type (medium abundance vs. high abundance  $z = 0.363$ ,  $P = 0.793$ ; low abundance vs. high abundance  $z = -0.262$ ,  $P = 0.793$ ; table 2), although the interaction between population types and nodulation status were significant (medium abundance × with nodules vs. high abundance × without nodules  $z = -2.241$ ,  $P = 0.025$ ; low abundance × with nodules vs. high abundance × without nodules  $z = 2.539$ ,  $P = 0.011$ ; table 2, see Supplementary Material S1 for the *post-hoc* comparisons for this interaction term). On average, queens without nodules from medium-abundance populations had 38% more oocytes in their ovaries than queens

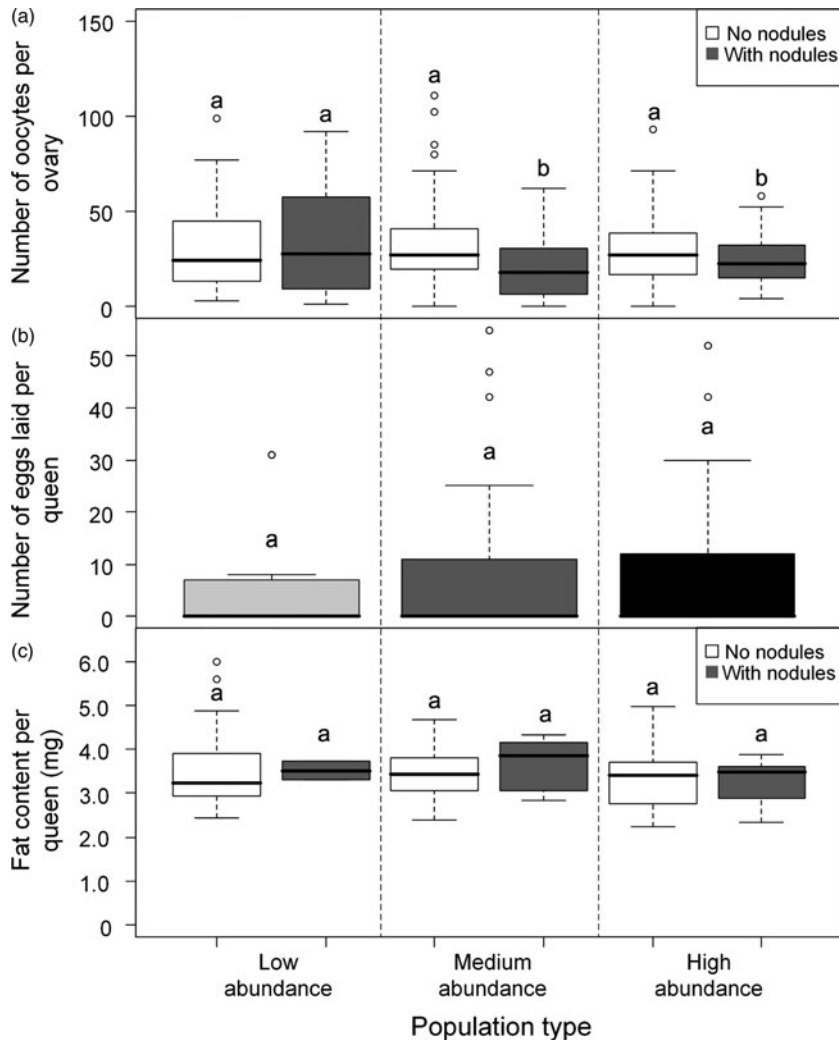


Fig. 3. (a) Box plot showing the median number of oocytes found in queens without nodules and queens with nodules at the three site types. Different letters above plots indicate significant differences ( $P < 0.05$ ). *Post-hoc* tests using least-square means were done using the R package 'lsmeans'. (b) Boxplot showing the median number of eggs laid by queens from low-abundance (light gray), medium-abundance (dark gray) and high-abundance (black) sites in the laboratory over a 24 h period. There were no differences in number of eggs laid between the different population types. (c) Box plot showing the median weight of fat found in queens without and with nodules at the three site types. There were no differences in fat content between queens from different population types.

with nodules (mean  $\pm$  SE:  $32.48 \pm 2.69$  vs.  $20.94 \pm 2.85$  oocytes, respectively). Queens without nodules from high-abundance populations had on average 17% more oocytes in their ovaries than queens with nodules (mean  $\pm$  SE:  $29.81 \pm 2.05$  vs.  $24.73 \pm 3.10$  oocytes, respectively), but queens without nodules from low-abundance populations had on average 17% fewer oocytes in their ovaries than queens with nodules (mean  $\pm$  SE:  $30.03 \pm 2.80$  vs.  $35.13 \pm 11.80$  oocytes, respectively).

#### Egg laying

There was no difference in the number of eggs laid by queens among population types (Kruskal–Wallis  $H$  test;  $\chi^2_2 = 0.272$ ,  $P = 0.873$ ; fig. 3b). However, only 35% of the 96 queens laid eggs; but excluding queens that did not lay eggs from the analysis also found no difference among the

population types (Kruskal–Wallis  $H$  test;  $\chi^2_2 = 1.005$ ,  $P = 0.605$ ). There was no difference in the number of eggs laid between queens with and without nodules (Mann–Whitney  $U$  test;  $U = 549.500$ ,  $P = 0.246$ ). As only 19% of queens in the egg-laying test had nodules, additional samples and analysis would be necessary to confirm the lack of statistical difference in egg-laying behavior between queen groups.

#### Fat content

Queens had a mean fat content of  $3.55 \times 10^{-3} \pm 1.33 \times 10^{-4}$  mg (mean  $\pm$  SE). Fat content did not vary with nodulation status ( $z = -0.030$ ,  $P = 0.978$ ) (fig. 3c) or population type (medium abundance vs. high abundance  $z = 0.260$ ,  $P = 0.797$ ; low abundance vs. high abundance  $z = -1.18$ ,  $P = 0.238$ ), nor was there an interaction effect (table 2).

## Discussion

This is the first study to find evidence that queens of the invasive ant *A. gracilipes* may be significantly affected by pathogens. Twenty-three percent of queens had melanized nodules in their ovaries or fat bodies, which was significantly associated with reduced oocyte number. It is unclear if the observed reduction in oocyte number was a direct effect of the invading pathogen, or a trade-off in immune response. Though a queen's immune system may have dealt effectively with the invader, such an immune function is biologically expensive and can result in physiological trade-offs, at both the individual and colony level (Rolff & Siva-Jothy, 2003; Siva-Jothy *et al.*, 2005). For example, *A. mellifera* colonies whose members mounted a strong immune response to pathogens produced significantly fewer larvae than those that did not (Evans & Pettis, 2005). Mosquitos that have encapsulated micro-filarial parasites also have reduced and delayed egg laying (Ferdig *et al.*, 1993).

The production of melanized nodules (nodulation) is an innate cellular immune response in insects (Stanley *et al.*, 2012). When an invading pathogen or parasite enters the hemocoel and is not phagocytized by a single hemocyte (the equivalent of white blood cells in insects), aggregations of hemocytes will surround the invader, smother and melanize it (Rolff & Siva-Jothy, 2003; Stanley *et al.*, 2012). The resulting nodule is attached to an organ or the inner body wall where it typically remains for the rest of the insect's life (Stanley *et al.*, 2012). This response may be induced by bacteria, fungus, parasitoid eggs and, in some cases, viruses (Carton *et al.*, 2002; Mullen & Goldsworthy, 2006; Durmus *et al.*, 2008; Gatschenberger *et al.*, 2013). It should be noted that though pathogens are the most likely cause of melanized nodules, ingestion of finely abrasive materials, such as cornstarch, may also cause them (L. Solter, personal communication). Bacteria have been the most commonly studied pathogen in relation to nodulation (Stanley *et al.*, 2012; Gatschenberger *et al.*, 2013). The nodulation response by hosts to viral challenge has been mixed. Azzami *et al.* (2012) found that injecting *A. mellifera* larvae and adults with *Acute bee paralysis virus* did not elicit a cellular immune response, though subjects did respond to *Escherichia coli* injection. However, larvae of the endoparasitoid wasp, *Pimpla turionellae*, produced nodules in response to an injection of the *Bovine Herpes simplex virus-1* (Durmus *et al.*, 2008). In addition to this work, we have unpublished laboratory tests that found no nodulation in *A. gracilipes* queens in response to injection of lipopolysaccharide, a major component of the outer membrane of Gram-negative bacteria, but a significant nodulation response when injected with fungal spores from *Metarhizium anisopliae* (see Supplementary Material S1). However PCR with universal fungal primers did not detect fungal DNA in *A. gracilipes* queens with a nodulation response (M. Cooling, unpublished data). We have not yet identified the cause of the nodules in this study.

There was no difference in fat content of queens between populations. It is possible that this pathogen or the observed immune response does not affect fat content in *A. gracilipes* queens, but there may also be a compensatory response from the ants. For example, it has been shown that the fungal pathogen *Nosema ceranae* can cause fat reduction in overwintering bees (Baily & Ball, 1991), but the fat content of *A. mellifera* queens infected with *N. ceranae* have been found to not differ from uninfected queens (Alaux *et al.*, 2011). The latter study suggested that because workers feed the queen, the

queen might be able to compensate for the nutritional stress of *N. ceranae* infection by increasing her food demand from the workers. This mechanism may also be the case in *A. gracilipes* queens, which are also fed by workers. Ants are known to develop fat during times of high resource abundance (Hahn, 2006; Lease & Wolf, 2011). We collected queens during February, which is the peak of the wet season, when *A. gracilipes* abundance, and presumably resources, is at their highest (Hoffmann, 2015). It is possible that differences in fat content may have been found among populations if queens had been collected in the dry season, when resources are scarcer and queens may be more likely to be stressed.

Surprisingly, low-abundance populations had the highest numbers of queens per nest and high-abundance populations had the lowest. This result does not support the hypothesis that queens are dying disproportionately at low- and medium-abundance populations compared with high-abundance populations. Perhaps the population must reach a critical mass before queens colonize new nests, as *A. gracilipes* queens require the attendance of workers to survive (Holldobler & Wilson, 1990). The relationship between queen number and nest density in polygynous ant species is variable and appears to be due to a complex interaction between ecological and genetic factors (Herbers & Stuart, 1996; Chapuisat *et al.*, 2004). For example, Ingram (2002a) found Argentine ant, *Linepithema humile*, queen numbers increased as nest density increased. However, another study by the same author discovered that low-density populations of *L. humile* had more queens per nest than a high-density population Ingram (2002b).

It is extremely difficult to document queen death in the field because workers dismember and dispose of dead queens immediately after death (personal observation). The replacement of dead queens by newly produced queens may also mask the depletion of queens. In Arnhem Land, however, new queens are only produced at one time of the year: at the end of the dry season around September (Hoffmann, 2015). The timing of this production means that in February, new queens should not be present to obscure functional queen numbers and we saw no evidence of this, as very few alates were observed.

Like other organisms, insects are constantly exposed to microbial pathogens and parasites, but few of these encounters result in a chronic or deleterious infection. We believe, though further work will be necessary to confirm it, we have documented an immune response in *A. gracilipes* queens to an unknown pathogen or parasite. While this work suggests that there may be a trade-off between immune response and oocyte production in *A. gracilipes* queens, we do not know if the pathogen that caused the immune response is still present in the population, or if the queens' immune system was capable of destroying it upon its initial invasion of the hemocoel. If it is still present in the population, it is also possible that the pathogen or pathogens are asymptomatic, like some *Picornavirales* RNA viruses that are commonly present as low-level chronic infections in hymenopterans like *S. invicta* and *A. mellifera* (de Miranda *et al.*, 2010). These viruses, such as SINV-1, usually appear to have no effect on colony health, though periods of high stress, or other, unknown causes, may trigger them to enter an acute-lethal stage (Bailey, 1967; de Miranda *et al.*, 2010; Porter *et al.*, 2013).

Our previous work has demonstrated that *A. gracilipes* are infected with various pathogens (Cooling *et al.*, 2016). Our current study provides a strong indication that pathogens

are likely to affect the reproductive biology of this invasive ant. However, it remains unclear whether these possible pathogens could be responsible for the observed population declines. Nodulation rate was not associated with population type, and the two high-abundance sites had the second and third highest nodulation rates after site M2 (23 and 20% respectively). Further analyses are needed to determine the nature of the microorganisms contained in the nodules. This may open a new avenue of biological control options for *A. gracilipes*, providing solutions to limitations currently experienced by toxic treatments.

### Supplementary material

The supplementary material for this article can be found at <https://doi.org/10.1017/S0007485317000967>.

### Acknowledgements

The authors thank many people that provided technical assistance, especially Daryl Lacey, Paul Augustin, Leanne Dzendolet, Tony Schultz and the many people involved with Conservation Volunteers Australia. Thanks to Dhimurru Aboriginal Corporation and the traditional landowners of northeast Arnhem Land for access to the region, and the staff of Rio Tinto for access to the mining leases and accommodation. The authors also thank the Lester laboratory group for constructive comments on the manuscript. This work was funded and supported by CSIRO Australia, Rio Tinto Alcan and the Dhimurru Aboriginal Corporation. M. Cooling was supported by a Victoria University of Wellington Doctoral Scholarship.

### References

- Alaux, C., Folschweiller, M., McDonnell, C., Beslay, D., Cousin, M., Dussaubat, C., Brunet, J. & Le Conte, Y. (2011) Pathological effects of the microsporidian *Nosema ceranae* on honey bee queen physiology (*Apis mellifera*). *Journal of Invertebrate Pathology* **106**, 380–385.
- Anderson, K., Sheehan, T., Eckholm, B., Mott, B. & DeGrandi-Hoffman, G. (2011) An emerging paradigm of colony health: microbial balance of the honey bee and hive (*Apis mellifera*). *Insectes Sociaux* **58**, 431–444.
- Azzami, K., Ritter, W., Tautz, J. & Beier, H. (2012) Infection of honey bees with acute bee paralysis virus does not trigger humoral or cellular immune responses. *Archives of Virology* **157**, 689–702.
- Bailey, L. (1967) The incidence of viral diseases in the honey bee. *Annals of Applied Biology* **60**, 43–48.
- Baily, L. & Ball, B. (1991) *Honey bee Pathology*, 2nd edn. Academic Press, London.
- Bates, D., Maechler, M., Bolker, B. & Walker, S. (2014) lme4: Linear mixed-effects models using Eigen and S4. R package version 1.1-7. Available online at <http://CRAN.R-project.org/package=lme4>.
- Briano, J. (2005) Long-term studies of the red imported fire ant, *Solenopsis invicta*, infected with the microsporidia *Vairimorpha invictae* and *Thelohania solenopsae* in Argentina. *Environmental Entomology* **34**, 124–132.
- Briano, J., Patterson, R. & Cordo, H. (1995a) Long-term studies of the black imported fire ant (Hymenoptera: Formicidae) infected with a microsporidium. *Environmental Entomology* **24**, 1328–1332.
- Briano, J., Patterson, R. & Cordo, H. (1995b) Relationship between colony size of *Solenopsis richteri* and infection with *Thelohania solenopsae* (Microsporidia: Thelohaniidae) in Argentina. *Journal of Economic Entomology* **88**, 1233–1237.
- Calleri, D., Rosengaus, R. & Traniello, J. (2006) Disease and colony establishment in the dampwood termite *Zootermopsis angusticollis*: survival and fitness consequences of infection in primary reproductive. *Insectes Sociaux* **53**, 204–211.
- Cameron, S., Lozier, J., Strange, J., Koch, J., Cordes, N., Solter, L. & Griswold, T. (2011) Patterns of widespread decline in North American bumble bees. *Proceedings of the National Academy of Sciences* **108**, 662–667.
- Carton, Y., Frey, F., Stanley, D., Vass, E. & Nappi, A. (2002) Dexamethasone inhibition of the cellular immune response of *Drosophila melanogaster* against a parasitoid. *Journal of Parasitology* **88**, 405–407.
- Chapuisat, M., Bocherens, S. & Rosset, H. (2004) Variable queen number in ant colonies: no impact on queen turnover, inbreeding, and population genetic differentiation in the ant *Formica selysi*. *Evolution* **58**, 1064–1072.
- Cooling, M. (2016) *Population dynamics and pathogens of the invasive yellow crazy ant (Anoplolepis gracilipes) in Arnhem Land, Australia*. Dissertation, Victoria University of Wellington.
- Cooling, M. & Hoffmann, B. (2015) Here today, gone tomorrow: declines and local extinctions of invasive ant populations in the absence of intervention. *Biological Invasions* **17**, 3351–3357.
- Cooling, M., Gruber, M., Hoffmann, B., Sébastien, A. & Lester, P. (2016) A metatranscriptomic survey of the invasive yellow crazy ant, *Anoplolepis gracilipes*, identifies several potential viral and bacterial pathogens. *Insectes Sociaux* **64**, 197–207.
- Cox-Foster, D., Conlan, S., Holmes, E., Palacios, G., Evans, J., Moran, N., Quan, P., Briese, T., Hornig, M., Geiser, D., Martinson, V., vanEngelsdorp, D., Kalkstein, A., Drysdale, A., Hui, J., Zhai, J., Cui, L., Hutchison, S., Simons, J., Egholm, M., Pettis, J. & Lipkin, W. (2007) A metagenomic survey of microbes in honey bee colony collapse disorder. *Science* **318**, 283–287.
- Dalecky, A., Gaume, L., Schatz, B., McKey, D. & Kjellberg, F. (2005) Facultative polygyny in the plant-ant *Petalomyrmex phylax* (Hymenoptera: Formicinae): sociogenetic and ecological determinants of queen number. *Biological Journal of the Linnean Society* **86**, 133–151.
- de Miranda, J., Cordon, G. & Budge, G. (2010) The acute bee paralysis virus-Kashmir bee virus-Israeli acute paralysis virus complex. *Journal of Insect Pathology* **103**, S30–S47.
- Dillon, R. & Dillon, V. (2004) The gut bacteria of insects: non-pathogenic interactions. *Annual Review of Entomology* **49**, 71–92.
- Dray, S. & Dufour, A. (2007) The ade4 package: implementing the duality diagram for ecologists. *Journal of Statistical Software* **22**, 1–20.
- Drescher, J., Feldhaar, H. & Bluthgen, N. (2011) Interspecific aggression and resource monopolization of the invasive ant *Anoplolepis gracilipes* in Malaysian Borneo. *Biotropica* **43**, 93–99.
- Dunn, A., Torchin, M., Hatcher, M., Kotanen, P., Blumenthal, D., Byers, J., Coon, C., Frankel, V., Holt, R., Hufbauer, R., Kanarek, A., Schierenbeck, K., Wolfe, L. & Perkins, S. (2012) Indirect effects of parasites in invasions. *Functional Ecology* **26**, 1262–1274.
- Durmuz, Y., Buyukguzel, E., Terzi, B., Tunaz, H., Stanley, D. & Buyukguzel, K. (2008) Eicosanoids mediate melanotic nodulation reactions to viral infection in larvae of the parasitic wasp, *Pimpla turionellae*. *Journal of Insect Physiology* **54**, 17–24.



- Elmes, G., Wardlaw, J., Nielsen, G., Kipyatkov, V., Lopatina, E., Radchenko, A. & Barr, B. (1999) Site latitude influences on respiration rate, fat content and the ability of worker ants to rear larvae: a comparison of *Myrmica rubra* (Hymenoptera: Formicidae) populations over their European range. *European Journal of Entomology* **96**, 117–124.
- Evans, J. & Pettis, J. (2005) Colony-level impacts of immune responsiveness in honey bees, *Apis mellifera*. *Evolution* **59**, 2270–2274.
- Fellous, S. & Lazzaro, B. (2010) Larval food quality affects adults (but not larval) immune gene expression independent of effects on general condition. *Molecular Ecology* **19**, 1462–1468.
- Ferdig, M., Spray, F., Li, J. & Christensen, B. (1993) Reproductive costs associated with resistance in a mosquito-filarial worm system. *American Journal of Tropical Medicine and Hygiene* **49**, 756–762.
- Gatschenberger, H., Azzami, K., Tautz, J. & Beier, H. (2013) Antibacterial immune competence of honey bees (*Apis mellifera*) is adapted to different life stages and environmental risks. *PLoS ONE* **8**, e66415.
- Gruber, M. (2012) *Genetic factors associated with variation in abundance of the invasive yellow crazy ant (Anoplolepis gracilipes)*. Dissertation, Victoria University of Wellington.
- Gruber, M., Burne, A., Abbott, K., Pierce, R. & Lester, P. (2012) Population decline but increased distribution of an invasive ant genotype on a Pacific atoll. *Biological Invasions* **15**, 599–612.
- Gruber, M., Cooling, M., Baty, J., Buckley, K., Friedlander, A., Quinn, O., Russell, J., Sébastien, A. & Lester, P. (2017) Single-stranded RNA viruses infecting the invasive Argentine ant, *Linepithema humile*. *Scientific Reports* **7**, 3304.
- Hahn, D. (2006) Two closely related species of desert carpenter ant differ in individual-level allocation to fat storage. *Physiological and Biochemical Zoology* **79**, 847–856.
- Haines, I. & Haines, J. (1978) Pest status of the crazy ant, *Anoplolepis longipes* (Jerdon) (Hymenoptera: Formicidae), in the Seychelles. *Bulletin of Entomological Research* **68**, 627.
- Herbers, J. & Stuart, R. (1996) Multiple queens in ant nests: impact on genetic structure and inclusive fitness. *American Naturalist* **147**, 161–187.
- Hoffmann, B. (2014) Quantification of supercolonial traits in the yellow crazy ant *Anoplolepis gracilipes*. *Journal of Insect Science* **14**, 25.
- Hoffmann, B. (2015) Integrating biology into invasive species management is a key principle for eradication success: the case of yellow crazy ant *Anoplolepis gracilipes* in northern Australia. *Bulletin of Entomological Research* **105**, 141–151.
- Hölldobler, B. & Wilson, E.O. (1990) *The Ants*. Cambridge, Belknap Press.
- Holway, D., Lach, L., Suarez, A., Tsutsui, N. & Case, T. (2002) The causes and consequences of ant invasions. *Annual Review of Ecology, Evolution and Systematics* **33**, 181–233.
- Ingram, K. (2002a) Plasticity in queen number and social structure in the invasive Argentine ant (*Linepithema humile*). *Evolution* **56**, 2008–2016.
- Ingram, K. (2002b) Flexibility in nest density and social structure in invasive populations of the Argentine ant, *Linepithema humile*. *Oecologia* **133**, 492–500.
- Kaspari, M. (1993) Body size and microclimate use in Neotropical granivorous ants. *Oecologia* **96**, 500–507.
- Lease, H. & Wolf, B. (2011) Lipid content of terrestrial arthropods in relation to body size, phylogeny, ontogeny and sex. *Physiological Entomology* **36**, 29–38.
- Lester, P., Bosch, P., Gruber, M., Kapp, E., Peng, L., Brenton-Rule, E., Buchanan, J., Stanislawek, W., Archer, M., Corley, J., Masciocchi, M., Van Oystaeyen, A. & Wenseleers, T. (2015) No evidence of enemy release in pathogen and microbial communities of common wasps (*Vespa vulgaris*) in their native and introduced range. *PLoS ONE* **10**, e0121358.
- McGlynn, T. (2010) Polygyny in thief ants responds to competition and nest limitation but not food resources. *Insectes Sociaux* **57**, 23–28.
- Moran, M. (2003) Arguments for rejecting the sequential Bonferroni in ecological studies. *Oikos* **11**, 403–405.
- Mullen, L. & Goldsworthy, G. (2006) Immune responses of locusts to challenge with the pathogenic fungus *Metarhizium* or high doses of laminarin. *Journal of Insect Pathology* **52**, 389–398.
- Nguyen, B., Ribiere, M., vanEngelsdorp, D., Snoeck, C., Saegerman, C., Kalkstein, A., Schurr, F., Brostaux, Y., Faucon, J. & Haubruge, E. (2011) Effects of honey bee virus prevalence, *Varroa destructor* load and queen condition on honey bee colony survival over the winter in Belgium. *Journal of Apicultural Research* **50**, 195–202.
- Nonacs, P. (1991) Less growth with more food: how insect-prey availability changes colony demographics in the ant, *Camponotus floridanus*. *Journal of Insect Pathology* **37**(12), 891–898.
- Oi, D. & Valles, S. (2009) Fire ant control with entomopathogens in the USA. pp 237–257 in Hajek, T., Glare, T. & O’Callaghan, M. (Eds.) *Use of Microbes for Control and Eradication of Invasive Arthropods*, vol 6. Dordrecht, The Netherlands, Springer.
- Oldroyd, B. (2007) What’s killing American honey bees? *PLoS ONE* **5**(6), e168.
- Porter, S., Valles, S. & Oi, D. (2013) Host specificity and colony impacts of the fire ant pathogen, *Solenopsis invicta virus 3*. *Journal of Insect Pathology* **114**, 1–6.
- R Core Team. (2013) R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna. Available online at <http://www.R-project.org>.
- Richards, M. & Packer, L. (1994) Trophic aspects of caste determination in *Halictus ligatus*, a primitively eusocial sweat bee. *Behavioural Ecology and Sociobiology* **34**, 385–391.
- RNAlater Handbook. (2006) Qiagen Corporation, Velno.
- Rolff, J. & Siva-Jothy, M. (2003) Invertebrate ecological immunology. *Science* **301**, 472–475.
- Russell, V., Hervão, L. & Hervão, M. (2015) lsmeans: least-square means. R package version 2.16. Available online at <http://CRAN.R-project.org/package=lsmeans>.
- Schmid-Hempel, P. (1998) *Parasites in Social Insects*. Princeton, Princeton University Press.
- Sébastien, A., Gruber, M. & Lester, P. (2011) Prevalence and genetic diversity of three bacterial endosymbionts (*Wolbachia*, *Arsenophonus*, and *Rhizobiales*) associated with the invasive yellow crazy ant (*Anoplolepis gracilipes*). *Insectes Sociaux* **59**, 33–40.
- Sébastien, A., Lester, P., Hall, R., Wang, J., Moore, N. & Gruber, M. (2015) Invasive ants carry novel viruses in their new range and form reservoirs for a honeybee pathogen. *Biology Letters* **11**, 20150610.
- Siva-Jothy, M., Moret, Y. & Rolff, J. (2005) Insect immunity: an evolutionary ecology perspective. *Advances in Insect Physiology* **32**, 1–48.
- Stanley, D., Haas, E. & Miller, J. (2012) Eicosanoids: exploiting insect immunity to improve biological control programs. *Insects* **3**, 492–510.

- Valles, S., Oi, D., Yu, F., Tan, X. & Buss, E. (2012) Metatranscriptomics and pyrosequencing facilitate discovery of potential viral natural enemies of the invasive Caribbean crazy ant, *Nylanderia pubens*. *PLoS ONE* **7**(2), e31828.
- Valles, S., Porter, S., Choi, M. & Oi, D. (2013) Successful transmission of *Solenopsis invicta virus 3* to *Solenopsis invicta* fire ant colonies in oil, sugar and cricket bait formulations. *Journal of Invertebrate Pathology* **113**, 198–204.
- vanEngelsdorp, D., Evans, J., Saegerman, C., Mullin, C., Haubruge, E., Nguyen, B., Frazier, M., Frazier, J., Cox-Foster, D., Chen, Y., Underwood, R., Tarpay, D. & Pettis, J. (2009) Colony collapse disorder: a descriptive study. *PLoS ONE* **4**(8), e6481.
- Wetterer, J. (2005) Worldwide distribution and potential spread of the long-legged ant, *Anoplolepis gracilipes* (Hymenoptera: Formicidae). *Sociobiology* **45**(1), 77–97.