Diurnal variation of fish and macrobenthic invertebrate community structure in an isolated oceanic island of the South Atlantic

PAUL EDWIN BREWIN^{1,2}, JUDITH BROWN^{1,3} AND PAUL BRICKLE^{1,4}

¹Shallow Marine Surveys Group, PO Box 598, Stanley, Falkland Islands FIQQ 1ZZ, South Atlantic, ²Falkland Islands Government Department of Fisheries, PO Box 598, Stanley, Falkland Islands FIQQ 1ZZ, South Atlantic, ³Ascension Island Government, Fisheries Department, Georgetown, Ascension Island ASCN 1ZZ, ⁴South Atlantic Environmental Research Institute, PO Box 609, Stanley, Falkland Islands, FIQQ 1ZZ, South Atlantic

The trophic structure of Ascension Island's sub-tidal reef assemblages is poorly understood. Unlike other tropical reef systems, sub-tidal habitats have very low abundance of both coral and macrophyte species. Visually dominant is a diverse assemblage of fish species, with particularly high densities of Melichthys niger, a voracious omnivore. In contrast, the nocturnal species assemblage is notably different, visually dominated by benthic invertebrates. To quantify the difference between day and night visible assemblages, we conducted day/night pairs of transect surveys of fish and invertebrates across three depths, and spanning 9 months, assigning all species to one of 10 functional groups. Multivariate analysis of surveys revealed significant turnover in species between day and night surveys and between survey periods, with concomitant changes in species rank-abundance distributions. Juveniles of a number of fish species were determinate in observed differences. Conversely, diversity of functional groups between day and night surveys and between seasons were not different, however there was significant species turnover within functional groups between day and night assemblages. The lack of proportional change in functional groups but a turn-over of species between day and night assemblages suggest that there may be a degree of functional redundancy in Ascension Island's marine trophic profile. Further investigation into the spatio-temporal variation in trophic profile and functional diversity around the island will benefit conservation and fisheries management in this isolated and poorly understood marine system.

Keywords: trophic, tropical reef, nocturnal, species turn-over, functional redundancy

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INTRODUCTION

Describing the trophic dynamics of marine communities can provide better understanding of the organization and functioning of assemblages (Paine, 1966; Menge, 1992; Hamilton et al., 2014), enable prediction of responses to disturbance or disease (Witman, 1985; Harley et al., 2006; Rodríguez-Barreras et al., 2014), and ultimately enhance their long-term management and conservation (Graham et al., 2003; Pauly & Watson, 2005; Mumby et al., 2007). In tropical reef systems, trophic dynamics tend to be 'top-down' (Carpenter & Edmunds, 2006; Hamilton et al., 2014), where the trophic interactions of predatory fish and their herbivore prey have direct and significant impact on fluctuations between coral and algal abundance (Floeter et al., 2005 for review). Consequently the relationships between predators, herbivores and primary producers have drawn a great deal of attention in shallow reef community studies due to their integrated role in structuring nutrient flow throughout the reef system (e.g. Barneche et al., 2014; Vinueza et al., 2014).

Corresponding author: P.E. Brewin Email: pbrewin@smsg-falklands.org

In response to typically high densities of visual predators on tropical reefs (e.g. fish species), prey species have evolved predator avoidance strategies such as migration into cracks and crevasses or other habitats that offer refuge during the day, and emerging at night when there is a lower encounter rate with visual predators (Dill, 1987; Barnes & Crook, 2001; Clark et al., 2003). In these cases, prey species must trade-off between optimal foraging opportunities and prey avoidance, because often foraging in predation refugia is sub-optimal (Holomuzki & Messier, 1993; Barnes & Crook, 2001 for review). For example, such daily retreat to predation refugia has been well documented for echinoids in tropical and temperate reefs (Nelson & Vance, 1979; Barnes & Crook, 2001), avoiding fish predators (see Tuya et al., 2004 for review). Similarly, a lack of daily migration in echinoids has been observed elsewhere when there are low abundances of predatory fish (Glynn et al., 1979). In these instances, echinoids are considered 'keystone' herbivores (e.g. D. antillarum in the tropical Atlantic: Hughes et al., 1987; Carpenter & Edmunds, 2006; Rodríguez-Barreras et al., 2014), where the reduced abundance of this species can result in dramatic increases in algal cover (Phinney et al., 2001). Therefore, when relating the importance of echinoids in reef trophic dynamics and their nocturnal feeding behaviour, it can be assumed that nocturnal grazing is when this species exerts maximum grazing pressure. However, herbivorous fish also play a major 'top-down' role in structuring tropical reef assemblages (Hughes *et al.*, 1987; Hamilton *et al.*, 2014 and references therein). Given that herbivorous fish are generally active in the day (Hay *et al.*, 1988; Choat & Clements, 1993) when historically most reef surveys have been carried out (although see Azzurro *et al.*, 2007), then a key period of critical trophic activity has been unstudied in Ascension Island; both night and day surveys of the reef assemblage are necessary for fully understanding trophic dynamics in tropical reef systems.

The inshore benthic habitats in Ascension Island are characterized by bedrock, boulders, cobble, maerl and sand, and conspicuously low coral substrate. Non-geniculate coralline red algae formations widely dominate the rocky bottom, where fleshy seaweeds are relatively inconspicuous, reduced to thin epibenthic turf-like mats, including diminutive and repent forms of various green, brown and red macroalgae (such as Wrangelia argus and Dictyota sp.) (Tsiamis et al., 2014). Visually dominating the daytime underwater seascape of Ascension Island is an abundant fish assemblage, with elements characteristic of both western and eastern Atlantic assemblages (Floeter et al., 2008, reviewed in Wirtz et al., 2014). Similar to other oceanic islands in the tropical Atlantic (Floeter et al., 2008; Pinheiro et al., 2011), richness is relatively low compared with coastal tropical Atlantic regions, with intermediate levels of endemism (Floeter et al., 2008). Characteristic of Ascension Island are the disproportionally dominant black triggerfish Melichthys niger in terms of both numbers and biomass (Price & John, 1980; Kavanagh & Olney, 2006). Melichthys niger is a voracious omnivore common throughout the tropical Atlantic. Analysis of stomach contents has revealed a wide spectrum of diet including benthic algae, invertebrates spanning most phyla, planktonic organisms and fish (Kavanagh & Olney, 2006) similar to other balistid fish (McClanahan, 2000). As such they are likely to exert significant grazing pressure on algal turf as well as predation pressure on invertebrate fauna. In contrast, preliminary night surveys (personal observation) showed an underwater seascape visually dominated by benthic invertebrate fauna, with a distinct lack of active fish present. Conspicuous nocturnal invertebrates include Diadema antillarum and the holothurian Euapta lappa. Interestingly, in addition to invertebrates, also seen are numerous M. niger in a state of deep rest, lying motionless on the rocky seabed.

There is currently no published quantitative information on the nocturnal benthic assemblage of Ascension Island. We therefore ask the following questions: (1) what are the differences between the day and night fish and invertebrate assemblages in terms of composition and abundance and (2) how are these assemblages structured in terms of their functional groups? To elucidate these questions, we carried out a series of day vs night scuba surveys collecting visual census data for the first time, providing a quantitative baseline of species that may otherwise be missed during day surveys as well as allowing for the comparison of the difference in relative abundance and species composition of marine fauna present between day and night.

MATERIALS AND METHODS

A total of 12 transect surveys were conducted at Wigan Pier, Ascension Island (7.894°S 14.384°W) spanning 9 months. The site is a small embayment at the base of a cliff headland. The seabed consists of very large (>1 m diameter) irregularly shaped and highly rugose boulders on very rugged bedrock, with numerous large fissures and crevasses. The reef slopes down to a sand/cobble/boulder field at a depth of 15-20 m, over a horizontal distance of approximately 200 m. In each survey period (season), six transect surveys were carried out; three day transects and three night transects (4 September 2012; day: 1500 h; night: 2025 h and 4 June 2013; day: 1500 h; night: 2030 h). In 2012, the three transects were placed at mean depths of 5.5 m (D1/N1), 5.1 m (D2/N2) and 7.0 m (D3/N3). In 2013, transects were placed at mean depths of 7.0 m (D1/N1), 6.3 m (D2/N2) and 8.5 m (D3/N3).

Surveys were carried out using standard visual survey methods, counting all fish and invertebrates within a 2 m belt (1 m either side of the tape measure) along a 50 m transect (based on methods in http://www.reeflifesurvey.com; Edgar et al., 2009). Transects were anchored on the seabed at three depths during the day surveys, suitably separated such that fish counts did not overlap between transects. The same transects were surveyed at night. Transects were marked with a single Cyalume-like glow stick at each end so they could be relocated by divers. Although the glow stick created a light field at night, it was deemed not significant enough to bias night survey results beyond the first metre of the transect. For the night survey, the survey diver wore a head-mounted torch. In all six night surveys, the head torch did not attract any fish species during the survey. All surveys in both seasons were carried out by the same diver surveyor, who was highly experienced in fish surveys in Ascension Island. Start and end depth and habitat type were noted for each transect.

All individuals were identified to species or closest practical taxonomic unit. Juvenile and adult forms of fish were recorded separately. All night observations of *M. niger* were of fish lying dormant on the seabed, and thus their presence was counted but classified separately from M. niger counted during the day. Categorizing these separately takes into account the different functional roles of M. niger in the day or night while at the same time, records their presence in the assemblage. Fish species were assigned to functional group codes based on Halpern & Floeter (2008), or from FishBase (http://www.fish base.org. Last accessed 23 October 2014) if not recorded in that study. Invertebrates were classified using the same terminology as fish, on the basis of a wide range of peer-reviewed sources. If specific species information could not be found, then functional grouping was assumed from the genus or family. Classifications are: Macrocarnivores (MCAR) consume mobile benthic organisms and fish; Strict piscivores (PISC) - consume fish only; Mobile benthic invertivores/cleaners (MINV) - consume primarily benthic mobile and parasitic invertebrates; Coral/colonial sessile invertivores (SINV) consume sessile benthic invertebrates; Planktivore (PLA) consume primarily macro- and micro-zooplankton; Turf grazing (TURF) - herbivores feeding on algae and epiphytic organisms; Scrapers (SCRP) - herbivores that leave shallow bite scars; General omnivores (OMNI) - consume a variety of animal and plant material in similar quantities; Detritivore (DETR) - consume sediments and decomposing material. The category DETR was needed for some species of invertebrates. The category INRT was used to classify M. niger at night.

All statistics were done in Rv3.1.0 (R Core Team, 2014). Multivariate analysis was conducted using the package 'vegan' $v_{2.0-10}$. Principal component analysis (PCA) was

	Code	Trophic guild	Day	Night
Fish species				
Triggerfish				
Melichthys niger (day)	MELNIG_d	OMNI	+	_
Melichthys niger (night)	MELNIG_n	INRT	_	+
Surgeonfish				
Acanthurus coeruleus	ACACOE	TURF	+	+
Acanthurus coeruleus (j)	ACACOE_j	TURF	+	_
Acanthurus bahianus	ACABAH	SCRP	+	+
Creolefish				
Paranthias furcifer	PARFUR	PLA	+	+
Paranthias furcifer (j)	PARFUR j	PLA	+	+
Butterflyfish				
Chaetodon sanctaehelenae	CHASAN	SINV	+	_
Boxfish				
Canthigaster sanctaehelenae	CANSAN	SINV	+	_
Damselfish	Gintonit	01111	'	
Chromis multilineata	CHRMIII	DI Δ		1
Stagastas lubhocki	STELLB	TUDE	- -	т 1
Stegustes lubbocki	STELUD :	TUDE	+	+
Stegastes IUDDOCKI ())	STELUE_J	TURF	+	+
Abudefduf saxatilis	ABUSAX	OMNI	+	-
Jacks				
Caranx lugubris	CARLUB	PISC	+	+
Hawkfish				
Amblycirrhitus earnshawi	AMBEAR	MINV	+	+
Amblycirrhitus earnshawi (j)	AMBEAR_j	MINV	+	-
Lizardfish				
Synodus synodus	SYNSYN	PISC	+	_
Trumpetfish				
Aulostomus strigosus	AULSTR	PISC	+	_
Tilefish				
Malacanthus plumieri (j)	MALPLU i	MCAR	+	_
Wrasse	/			
Thalassoma ascensionis	THAASC	PLA	+	_
Thalassoma ascensionis (i)	THAASC i	PLA	+	_
Thalassoma sanctahelenae	THASAN	DIA		_
Rodianus insularis	BODINS	MINV	- -	1
Podianus insularis (i)	PODINS ;	MINIV	- -	т _
Courses	BODINS_J	IVIIIN V	Ŧ	_
Epinophalua a dagamaiania	EDIADO	MCAD		
Epinepheius aascensionis	EPIADS	MCAR	+	+
Epinephelus adscensionis (j)	EPIADS_J	MCAR	+	+
Soapfish				
Rypticus saponaceus	RYPSAP	MCAR	+	+
Squirrelfish				
Holocentrus adscensionis	HOLADS	MINV	+	+
Holocentrus adscensionis (j)	HOLADS_j	MINV	+	+
Soldierfish				
Myripristis jacobus	MYRJAC	MCAR	+	+
Myripristis jacobus (j)	MYRJAC_j	MCAR	+	+
Moray eels	2			
Channomuraena vittata	CHAVIT	MCAR	+	+
Enchelycore anatina	ENCANA	MCAR	+	+
Quassiremus ascensionis	QUAASC	MINV	+	+
Gymnothorax moringa	GYMMOR	MCAR	+	+
Gymnothorax moringa	GYMMIL	MCAR	+	+
Gymnothorax unicolor	GYMUNI	MCAR	_	+
Goby	GIMOIN	1010/110		1
Prioletic accancionic	DRIASC	MINN	_1	_
Priotepis uscensionis	FRIASC	IVIIIN V	Ŧ	
Ophiahl	ODICD	TUDE		
Opniobiennius sp.	OPISP	IUKF	+	+
Cardinalfish		DI A		
	1 DO 1 371			
Apogon axillaris	APOAXI	PLA	+	+

Fable 1.	Species	presence/	absence	in all	surveys.	Trophic	c guilds	adapted
		from	Halpern	& Flo	oeter (200	<mark>08</mark>)		

Table 1. Continued

	Code	Trophic guild	Day	Night
Scorpaenodes insularis	SCOINS	MINV	-	+
Invertebrate species				
Echinodermata				
Ophidiaster guildingi	OPHGUI	PLA	+	_
Eucidaris tribuloides	EUCTRI	TURF	+	+
Diadema antillarum	DIAANT	TURF	+	+
Holothuria (Platyperona)	HOLSAN	DETR	+	+
sanctori				
Holothuria (Halodeima)	HOLGRI	DETR	_	+
grisea				
Euapta lappa	EUALAP	DETR	_	+
Crustacea				
Percnon abbreviatum	PERABB	MINV	+	+
Panulirus echinatus	PANECH	MINV	+	+
Corallianassa longiventris	CORLON	PLA	+	+
Dardanus imperator	DARIMP	OMNI	_	+
Euryozius pagalu	EURPAG	OMNI	_	+
Enoplometopus antillen	ENOANT	MINV	_	+
Decorator crab	DECCRA	OMNI	_	+
Stenopus hispidus	STEHIS	MINV	_	+
Hippolysmata grabhami	HIPGRA	MINV	+	+
Brachycarpus biunguiculatus	BRABIU	MINV	_	+
Polychaeta				
Amphinomidae	FIRWOR	SINV	_	+
Mollusca				
Bursa (Colubrellina)	BURCOR	MINV	_	+
corrugata				
Platydoris angustipes	PLAANG	SINV	_	+
Charonia variegata	CHAVAR	MINV	_	+
Octopus vulgaris	OCTVUL	MINV	_	+
Cnidaria				
Telmatactis sp1	TELSP1	PLA	_	+
Telmatactis sp2	TELSP ₂	PLA	_	+
Telmatactis sp3	TELSP3	PLA	_	+
Isarachnanthus maderensis	ISAMAD	PLA	_	+

carried out on the species matrix of day and night transect observations after 4th root transformation, reducing the weight of large values while retaining relative abundance information (Legendre & Legendre, 1998). Significance of site groupings was done using ANOSIM (analysis of similarity). SIMPER (similarity percentage) was used to determine the species having the highest 90% of influence on site groupings. Functional group diversity of day and night assemblages was examined in terms of richness (S), Shannon diversity (H') and Pielou's evenness (J') (Maurer & McGill, 2011). Morisita – Horn index was used to examine the difference (dissimilarity) in species composition between day and night surveys, within each functional group (Maurer & McGill, 2011). This is an abundance-based index of species overlap between two groups (day vs night).

RESULTS

Species composition and abundance

A total of 57 taxa were identified among a total of 6547 individuals counted, including 32 species of fish, 10 of which included juvenile forms, and 25 invertebrate species (Table 1). Of the fish species, 29 species were found during



Fig. 1. Density of total fish and invertebrates between paired transects (D1, D2, D3, N1, N2, N3) between seasons (September 2012 and June 2013).

day transects with 13 found exclusively during the day. At night 23 species were found, with the brown moray eel (*Gymnothorax unicolor*), cardinal fish (*Apogon* *pseudomaculatus*) and scorpion fish (*Scorpaenodes insularis*) found exclusively at night. Of the invertebrates, eight species were found during the day transects, with the ophiuroid *Ophidiaster guildingi* found exclusively during the day. All other invertebrates were recorded at night, with 16 taxa occurring exclusively at night.

In all transect pairs, the density of all fish decreased between day and night surveys by an average of 27.9% (\pm 15.6% SD) (Figure 1). Conversely, the density of invertebrates increased between day and night surveys by an average of 53.5% (\pm 8.4% SD). In 2012, density of combined fish and invertebrates (using day and night transects as replicates) increased with depth (ANOVA, df = 2, F = 22.645, P = 0.0155). There was no similar increase or decrease detected in 2013. Rank abundance analysis of all species from all transects shows that *M. niger* and *D. antillarum* were ranked among the highest three species in all day and night transects in both seasons (Table 2). In addition to these two species, juvenile *Paranthias furcifer* ranked highly in 2012, but were not recorded in 2013. *Ophioblennius* sp. and juvenile *Epinephelus adscensionis* also ranked highly in

Table 2. Abundance (100 m^{-2}) of the top 10 ranked species in all transects. Ranks are coded dark grey (highest rank) to light grey (lowest rank) for easeof interpretation.

D1 N1 D2 N2 D3 N3 D1 N1 D2 N2 D3 Paranthias furcifer (j) 135 101 180 200 360 370	N3
Paranthias furcifer (j) 135 101 180 200 360 370	95
	95
Melichthys niger d/n 120 124 100 57 100 135 100 74 120 127 60	
Diadema antillarum 55 132 121 301 329 510 107 237 119 208 166	337
Ophioblennius sp 24 24 19 18 15 9	
<i>Epinephelus adscensionis</i> (j) 20 5 28 <u>15</u> 33 6	_
Stegastes lubbocki 16 16 31 23 4 30 6 29	4
Thalassoma ascensionis 13 7 4	
Malacanthus plumieri (j) 7 7	
Abudefduf saxatilis 7	
Holocentrus adscensionis 5 10 15 8 7 12 6 4	
<i>Euapta lappa</i> 6 7 13 16 8	15
Apogon axillaris 5 4 21 19 4	14
Brachycarpus biunguiculatus 3 5	12
Paranthias furcifer 3 10 7 4	10
Holothuria (Platyperona) sanctori 2 3	
Myripristis jacobis 1 11 16	
Malacanthus plumieri (j) 10	
Acanthurus bahianus 7 3	
Acanthurus coeruleus 5	
Telmatactis sp2 7	
Telmatactis sp1 6	
Gymnothorax moringa 10	
Thalassoma ascensionis (j) 60 30	
Stegastes lubbocki (j) 16 27 25	
Percnon abbreviatum 6 4	-
Chaetodon sanctaehelenae 4	
Telmatactis sp3 4	24
Hippolysmata grabhami 4	
Scorpaenodes insularis 4	
Holocentrus adscensionis (j) 4 5	
Chromis multilineata 9	
Eucidaris tribuloides 4	
Myripristis jacobus (j) 4	5
Epinephelus adscensionis 4	
Corallianassa longiventris	
Priolepis ascensionis 4	
Channomuraena vittata	3



Fig. 2. Photo of resting M. niger at night (4 September 2014, N1Quadrat 5). Also seen are D. antillarum, and Euapta lappa (arrow). Quadrat is 0.5 × 0.5 m.

day surveys. In 2013 top ranked species also included juvenile *Thalassoma ascensionis* and juvenile and adult *Stegastes lubbocki* in day surveys, and *Apogon axillaris* and *Telmatactis* sp3 in night surveys. In total, there were 37 out of the total of 57 species ranked in the top 10 most abundant species among all transects.

Because M. niger and D. antillarum were found in large numbers consistently throughout all transects (M. niger; mean density = 101.0 individuals 100 m⁻², range = 57-135: D. antillarum; mean density = 218.5 individuals 100 m⁻², range = 55-510), these species are analysed separately. Two-sample Wilcoxon test (as data did not conform to parametric assumptions) was used to compare day and night densities among pairs of day/night transects in these species. There was no significant difference found between day and night densities of *M. niger* (W = 16, P = 0.81). However, all night observations of this species were of individuals in a state of deep rest (Figure 2). In contrast, pairs of day and night densities of D. antillarum were significantly different (W = 5, P = 0.0411). *Diadema antillarum* tended to hide during day transects, although many could be observed by the surveyor. At night however, urchins were observed out on open rock surfaces (Figure 2).

Principal component analysis of the day and night species matrix shows significant groupings of day and night surveys and surveys between seasons (ANOSIM, R-statistic = 0.8827, P = 0.001) (Figure 3). PC1 explains 39.5% of the variation in surveys, and represents the gradient between day (negative PC1) and night (positive PC1) surveys. This gradient seems to be strongly influenced by active or resting M. niger. As noted above, their day and night densities were not significantly different. Therefore in the PCA (Figure 3), the influence of M. niger effectively cancels each other out, and the abundance of other species can be interpreted with respect to their relative influence on survey groupings. Species driving this gradient are the echinoderms Euapta lappa, Holothuria (Platyperona) sanctori, Holothuria (Halodeima) grisea, D. antillarum at night, and the fish Ophioblennius sp., Thalassoma ascensionis, Stegastes lubbocki (adult) during the day. The gradient shown along the PC2 axis (explaining

29.5% of variation between surveys) represents the change in day and night assemblages between season (September 2012 and June 2013). These groupings are heavily influenced by the presence or absence of the juveniles of *Paranthias furcifer*, *Epinephelus adscensionis*, *Malacanthus plumieri*, *Myripristis jacobus* in 2012, and *Apogon axillaris*, *Stegastes lubbocki*, *Holocentrus adscensionis* and *Thalassoma ascensionis* in 2013. Analysis using SIMPER (per group identified in PCA) confirms these species are important for distinguishing groups; further detail of species driving the top 90% of these groupings are shown in Table 3.

Functional group diversity

Both day and night functional group profiles showed large proportions of herbivores (TURF) and planktivores (PLA) (Figure 4). The main difference in functional group profile was the presence (day) or absence (night) of omnivores (OMNI). These were composed almost entirely of *M. niger*, where active fish in the day were replaced by resting (INERT) *M. niger* at night. Other species of omnivore were present in the day and at night, although in low numbers (Table 1). Other functional groups were present in generally low proportions, varying between day and night transects in different ways, e.g. macrocarnivores (MCAR) and mobile benthic invertivores/cleaners (MINV). Detritivores were found primarily at night, being composed of three species of holothurians (Table 1).

Functional group diversity was examined between day and night transects for each season. There were no significant differences between day/night *S* (here being the number of functional groups), *H'* (relative abundance of groups present) and *J'* (evenness of groups present) in each season (after ANOVA and Tukey's HSD *post hoc* tests). Within functional groups, species overlap between day and night transect pairs was tested using Morisita–Horn index of dissimilarity (Table 4). Particularly low dissimilarity between day and night assemblages was found within the planktivores (PLA) in 2012, due in part to high abundances of juvenile *Paranthias furcifer* in both day and night surveys. However in 2013, there was



Fig. 3. Day and night species PCA. Top and bottom plots are identical, where the top plot names fish species, and the bottom plot names invertebrates. Species in black are those identified in SIMPER as being in the top 90% of species driving separation of day/night and 2012/2013 pattern.

high dissimilarity between day and night planktivore species, this time driven by the relatively high abundance of juvenile *Thalassoma ascensionis* in the day and high abundance *Apogon axillaris* at night. For mobile benthic invertivores/cleaners (MINV), low dissimilarity was seen in 2012, while high dissimilarity was seen in 2013, possibly driven by the presence of the shrimp *Brachycarpus biunguiculatus* at night, but also the presence or absence of a variety of low abundance species found in night or day surveys. There was a surprisingly low dissimilarity for herbivore species turnover between day and night transects (Table 4), suggesting that in general most herbivore species were found in both day and night surveys, although varying in their abundance only between day and night.

DISCUSSION

We quantitatively show clear differences between the day and night shallow reef assemblages on Ascension Island. Overall, invertebrate density increased by over 50% at night compared with the day, while fish density decreased. The sea urchin *Diadema antillarum* was a large component of this increase, emerging from crevasses at night; this predator avoidance behaviour has been well studied in the tropical Atlantic (e.g. Rodríguez-Barreras *et al.*, 2014). Indeed, the majority of invertebrates surveyed were found exclusively at night. For example, the holothurians *Holothuria* (*Platyperona*) sanctori, *H.* (*Halodeima*) grisea and *Euapta lappa* all increased activity

	Contr	SD	Cumsum		Contr	SD	Cumsum
D2012/N2012				D2013/N2013			
Diadema antillarum	0.14	0.10	0.28	Diadema antillarum	0.15	0.06	0.23
Melichthys niger (day)	0.09	0.03	0.46	Melichthys niger (night)	0.12	0.03	0.42
Paranthias furcifer (j)	0.09	0.06	0.63	Melichthys niger (day)	0.11	0.03	0.59
Melichthys niger (night)	0.08	0.04	0.79	Thalassoma ascensionis (j)	0.07	0.03	0.70
Ophioblennius sp.	0.02	0.01	0.83	Stegastes lubbocki	0.03	0.00	0.75
Stegastes lubbocki	0.01	0.00	0.86	Stegastes lubbocki (j)	0.03	0.01	0.79
Epinephelus adscensionis (j)	0.01	0.01	0.89	Apogon axillaris	0.02	0.00	0.82
Malacanthus plumieri (j)	0.01	0.00	0.90	Ophioblennius sp.	0.02	0.00	0.84
				Euapta lappa	0.02	0.00	0.86
				Telmatactis sp3	0.01	0.01	0.88
				Brachycarpus biunguiculatus	0.01	0.00	0.89
				Paranthias furcifer	0.01	0.00	0.90
D2012/D2013				N2012/N2013			
Paranthias furcifer (j)	0.21	0.05	0.41	Paranthias furcifer (j)	0.18	0.05	0.42
Diadema antillarum	0.09	0.06	0.58	Diadema antillarum	0.12	0.06	0.70
Thalassoma ascensionis (j)	0.06	0.03	0.70	Melichthys niger (night)	0.03	0.02	0.77
Epinephelus adscensionis (j)	0.02	0.00	0.75	Apogon axillaris	0.01	0.01	0.81
Melichthys niger (day)	0.02	0.03	0.79	Telmatactis sp3	0.01	0.01	0.82
Stegastes lubbocki (j)	0.02	0.01	0.84	Epinephelus adscensionis (j)	0.01	0.00	0.84
Stegastes lubbocki	0.01	0.01	0.86	Holocentrus adscensionis	0.01	0.00	0.85
Ophioblennius sp.	0.01	0.01	0.87	Euapta lappa	0.01	0.00	0.87
Malacanthus plumieri (j)	0.01	0.00	0.89	Myripristis jacobus	0.01	0.01	0.88
Thalassoma ascensionis	0.01	0.00	0.90	Paranthias furcifer	0.01	0.00	0.89
				Brachycarpus biunguiculatus	0.00	0.00	0.90

 Table 3. SIMPER results for day/night comparison, and 2012/2013 comparison. Results of this analysis are Contr, Average contribution to overall dissimilarity; SD, Standard deviation of contribution; Cumsum, Ordered cumulative contribution.



Fig. 4. Proportion of functional groups per transect. Functional group codes are found in Table 1.

at night. These large (10 s of cm in length), soft-bodied, slowmoving species may be particularly susceptible to predation, and their nocturnal emergence may be a predator avoidance behaviour (Hammond, 1982). PCA and ANOSIM analyses indicate significant differences between day and night assemblages, however these analyses also show that there is a great deal of variability not well explained, where the first two PCA axes explained only 69% of multivariate variation in the species matrix. Sample size is relatively low in the present study with limited replication, which may lead to a poorly resolved pattern. This study does provide strong guidance towards developing and testing hypotheses of important ecological gradients within the assemblage (e.g. day/night, depth, season) that can be translated into a more fully replicated and stratified design in the future, spanning a wider spatial extent.

It should be pointed out that day and night survey observations are based on those species that are in fact observable in either the day or night. In other words, with the exception of some fish that may have diurnal depth migrations, the actual assemblage is likely to be similar between day and night, where unobserved taxa are present but hidden. For example,

Functional group	2012	2012			2013		
	DN1	DN2	DN3	DN1	DN2	DN3	
TURF	0.18554	0.05413	0.01351	0.10669	0.12552	0.06537	
MCAR	0.11471	0.14544	0.38138	0.58729	0.32151	0.62298	
MINV	0.09576	0.03614	0.04261	0.45525	0.56627	0.88471	
PLA	0.00717	0.00141	0.00186	0.96602	0.93798	0.89502	
DETR	0.70492	Too few data					
SINV	Too few data						
SCRP	Too few data						
PISC	Too few data						
OMNI	Too few data						

 Table 4. Species overlap (Morisita – Horn index) within functional groups. High values represent high dissimilarity (i.e. low number of shared species).

 Instances marked 'too few data' mean that there were too few occurrences (many zeros) or the functional group was made up of only one species, meaning that comparison was not possible.

M. niger densities were not significantly different between day (observed to be swimming) and night (observed to be resting) transects, suggesting that they are somewhat resident to specific areas. Conversely, D. antillarum densities were significantly different between day and night surveys. However in this case, they are simply hidden from view in the day, and this is likely to be the case for all invertebrates observed at night. What this means in terms of trophic dynamics is unclear, however it is thought that species' refuge sites are suboptimal for food resources (Barnes & Crook, 2001 for review). Therefore it is reasonable to assume that the trophic ecology of the reef is roughly partitioned between day and night, described by a representative turnover of species and related functional groups. What this means in terms of the reef's day and night distribution of biomass and coincident diversity remains unknown; further investigation into the diversitybiomass (i.e. productivity) relationship would offer good insight into the overall ecosystem functioning of tropical reef systems (Gaston & Blackburn, 2000).

Although predator avoidance may be a driver for nocturnal activity in some species, many other reef species have evolved to take predatory advantage of their behaviour. For example, we show that anemone (Telmatactis sp. and Isarachnanthus *maderensis*) emergence is exclusively at night. This is likely to be related to feeding on the nocturnal emergence of benthic zooplankton (Sebens & DeRiemer, 1977). Increased zooplankton (e.g. copepods) and other small crustacean (cumaceans and amphipods, etc.) abundance at night has been reported widely for temperate and tropical shallow reefs (Ohlhorst, 1982; Annese & Kingsford, 2005; Nakajima et al., 2009; Heidelberg et al., 2010), where they swarm near the bottom or hide in sediment in the day, and emerge into the water column at night with depth and time of emergence varying with species (Alldredge & King, 1980). Increases in zooplankton at night were not measured in this study, however it can be inferred through the higher abundance of other planktivorous fish such as the cardinalfish (Apogon axillaris) (Marnane & Bellwood, 2002) reported in the present study at night, particularly in 2013. Nocturnal feeding on zooplankton may also drive the emergence of benthic decapods found in the present study (e.g. Brachycarpus biunguiculatus). Furthermore, the role of other more cryptic trophic associations such as diel patterns of fish parasites, fish, and their mutualistic cleaner fish and invertebrates commensals (Chambers & Sikkel, 2002) cannot be ignored.

In addition to day/night patterns of species abundance, there is also a strong seasonal pattern in both day and night assemblages. This pattern was driven primarily by the presence or absence of juveniles of a variety of fish species, where principal component analysis and rank abundance analyses show a difference in juvenile night assemblage between surveys (spanning 9 months) as well as differences in the day assemblages. Timing and location of, for example, spawning aggregations in reef fish are known to be highly precise and variable among species (Domeier & Colin, 1997). In particular, species of the Serranidae (e.g. Paranthias furcifer and Epinephelus adscensionis highlighted in the present study), are known to form seasonal spawning aggregations; their differences in reproductive timing may account for the seasonal variation in their presence or absence shown in Ascension Island. Difference in length of larval stages will also determine when the juveniles are visible in abundance counts. Certainly, variability between seasons in other species may be due to those species being rare or found in particularly low abundances in the area surveyed (i.e. those having low rank abundance). As such, sampling error may account for at least some of this variability. More comprehensive sampling is needed to find true seasonal or inter-annual differences in rare species. In addition, seasonal patterns related to reproductive timing also need to be considered in future trophic studies.

Herbivorous and planktivorous species dominated the functional group profile in all the day and night surveys. There was no difference in the diversity of functional groups between day and night, although there was significant turnover in species within each functional group between day and night assemblages. Given this high species turnover, this suggests that there may be a degree of 'functional redundancy' (sensu Rosenfeld, 2002) in this system. That is, because species within the group share a similar functional role, some species loss may have little effect on the overall functioning of the ecosystem. Functional redundancy, therefore, offers the ecosystem some resilience to perturbations. This has been studied in herbivorous reef fish (e.g. Hamilton et al., 2014), and in cases where there is significant exploitation of reef fish, the consequences of reduced functional redundancy can have significant negative effects on the reef system (e.g. Bellwood et al., 2003). Although we define functional diversity solely by the presence and relative abundance of trophic groups, true functional diversity of an assemblage may integrate a wider range of ecological traits including species abundance, body size, behaviour and habitat characteristics (e.g. Stuart-Smith et al., 2013; Whittaker et al., 2014). In the present study, we found numerical dominance in a few species, but low

abundance (local rarity) in others, drawing into question how much effective redundancy there actually is within functional groups. Indeed, our interpretation of functional redundancy in the present study can be an artefact of how we have defined our functional groups (Bellwood et al., 2003, Halpern & Floeter, 2008), where our functional group might be too broad, thereby being more descriptive of the species' fundamental niche space, when the reality is that each species occupies a very narrow realized niche space (Hutchinson, 1957; Morlon et al., 2014), which would result in a reduced functional redundancy overall. Future trophic studies in Ascension Island should focus on the importance of abundant versus rare species, their relationship to habitat type, and species' morphological features as has been shown to be important in other reef systems (e.g. Ellingsen et al., 2007; Stuart-Smith et al., 2013; Aguilar-Medrano & Calderon-Aguilera, 2015).

The presence/absence and abundance of species in the day compared with night, and hence the trophic profile of the reef, will likely be strongly correlated with habitat complexity as it relates to refuge space for fish and invertebrates (e.g. Beukers & Jones, 1997; Wilson et al., 2007 for review). The subtidal rocky seabed of Ascension Island can be extremely complex, with non-geniculate coralline red algae creating formations that add significantly to benthic rugosity. Such complexity will vary spatially throughout the island and at varying depths; in the present study the same habitat was surveyed in both seasons. However, future studies of trophic complexity or other ecological studies of assemblages across wider spatial scales should include habitat complexity as a co-variable. Furthermore, there is evidence of important yet highly cryptic habitats such as in the internal spaces of maerl balls and reef-forming coralline tower formations (SMSG unpublished observations) that should be more thoroughly explored.

Trophic processes of tropical marine communities can be placed on a spectrum of top-down (e.g. predation pressure) versus bottom-up (e.g. physical forcing) control of trophic dynamics, where placement along this spectrum is dependent on biotic and abiotic processes acting across a broad range of spatio-temporal scales (Vinueza et al., 2014). Moreover, in these systems dynamics are not static and can be modified by, for example, spatio-temporal variability in oceanographic processes (e.g. Vinueza et al., 2014). Our data show that there is significant daily and seasonal variability in species composition, abundance and functional group profile in Ascension Island. This suggests that there are a very wide range of species and processes that contribute to the trophic dynamics of the reef system. For example, at this location there have been significant seasonal cold-water upwellings detected compared with near-by sites (SMSG, unpublished data). Future research should examine interactions of such bottom up-processes with potential top-down processes (herbivory and predation on herbivores) for better understanding of reef assemblages and ecosystem function across local and wider-spatial scales in this highly isolated tropical reef system.

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REFERENCES

- Aguilar-Medrano R. and Calderon-Aguilera L.E. (2015) Redundancy and diversity of functional reef fish groups of the Mexican Eastern Pacific. *Marine Ecology*. doi: 10.1111/maec.12253.
- Alldredge A.L. and King J.M. (1980) Effects of moonlight on the vertical migration patterns of demersal zooplankton. *Journal of Experimental Marine Biology and Ecology* 44, 133–156.
- Annese D.M. and Kingsford M.J. (2005) Distribution, movements and diet of nocturnal fishes on temperate reefs. *Environmental Biology of Fishes* 72, 161–174.
- Azzurro E., Pais A., Consoli P. and Andaloro F. (2007) Evaluating daynight changes in shallow Mediterranean rocky reef fish assemblages by visual census. *Marine Biology* 151, 2245-2253.
- Barneche D.R., Kulbicki M., Floeter S.R., Friedlander A.M., Maina J. and Allen A.P. (2014) Scaling metabolism from individuals to reef-fish communities at broad spatial scales. *Ecology Letters* 17, 1067–1076.
- Barnes D.K.A. and Crook A.C. (2001) Quantifying behavioural determinants of the coastal European sea-urchin *Paracentrotus lividus*. *Marine Biology* 138, 1205–1212.
- Bellwood D.R., Hoey A.S. and Choat H. (2003) Limited functional redundancy in high diversity systems: resilience and ecosystem function on coral reefs. *Ecology Letters* 6, 281–285.
- Beukers J.S. and Jones G.P. (1997) Habitat complexity modifies the impact of piscivores on a coral reef fish population. *Oecologia* 114, 50-59.
- Carpenter R.C. and Edmunds P.J. (2006) Local and regional scale recovery of *Diadema* promotes recruitment of scleractinian corals. *Ecology Letters* 9, 271–280.
- **Chambers S.D. and Sikkel P.C.** (2002) Diel emergence patterns of ecologically important, fish-parasitic, Gnathiid isopod larvae on Caribbean coral reefs. *Caribbean Journal of Science* 38, 37–43.
- Choat J.H. and Clements K.D. (1993) Daily feeding rates in herbivorous labroid fishes. *Marine Biology* 117, 205–211.
- Clark K.L., Ruiz G.M. and Hines A.H. (2003) Diel variation in predator abundance, predation risk and prey distribution in shallow-water estuarine habitats. *Journal of Experimental Marine Biology and Ecology* 287, 37–55.
- Dill L.M. (1987) Animal decision making and its ecological consequences, the future of aquatic ecology and behaviour. *Canadian Journal of Zoology* 65, 803–811.
- Domeier M.L. and Colin P.L. (1997) Tropical reef fish spawning aggregations: defined and reviewed. *Bulletin of Marine Science* 60, 698-726.
- Edgar G.J., Barrett N.S. and Stuart-Smith R.D. (2009) Exploited reefs protected from fishing transform over decades into conservation

features otherwise absent from seascapes. *Ecological Applications* 19, 1967–1974.

- Ellingsen K.E., Hewitt J.E. and Thrush S.F. (2007) Rare species, habitat diversity and functional redundancy in marine benthos. *Journal of Sea Research* 58, 291–301.
- Floeter S.R., Behrens M.D., Ferreira C.E.L., Paddack M.J. and Horn M.H. (2005) Geographical gradients of marine herbivorous fishes: patterns and processes. *Marine Biology* 147, 1435–1447.
- Floeter S.R., Rocha L.A., Robertson D.R., Joyeux J.C., Smith-Vaniz W.F., Wirtz P., Edwards A.J., Barreiros J.P., Ferreira C.E.L., Gasparini J.L., Brito A., Falcón J.M., Bowen B.W. and Bernardi G. (2008) Atlantic reef fish biogeography and evolution. *Journal of Biogeography* 35, 22–47.
- Gaston K.J. and Blackburn T.M. (2000) Pattern and process in macroecology. Oxford: Blackwell Publishing, 377 pp.
- Glynn P.W., Wellington G.M. and Birkeland C. (1979) Coral reef growth in the Galápagos: limitation by sea urchins. *Science* 203, 47–49.
- Graham N.A.J., Evans R.D. and Russ G.R. (2003) The effects of marine reserve protection on the trophic relationships of reef fishes on the Great Barrier Reef. *Environmental Conservation* 30, 200-208.
- Halpern B.S. and Floeter S.R. (2008) Functional diversity responses to changing species richness in reef fish communities. *Marine Ecology Progress Series* 364, 147–156.
- Hamilton S.L., Smith J.E., Price N.N. and Sandin S.A. (2014) Quantifying patterns of fish herbivory on Palmyra Atoll (USA), an uninhabited predator-dominated central Pacific coral reef. *Marine Ecology Progress Series* 50, 141–155.
- Hammond L.S. (1982) Patterns of feeding and activity in deposit-feeding holothurians and echinoids (Echinodermata) from a shallow back-reef lagoon, Discovery Bay, Jamaica. *Bulletin of Marine Science* 32, 549– 571.
- Harley C.D.G., Hughes A.R., Hultgren K.M., Miner B.G., Sorte C.J.B., Thornber C.S., Rodriguez L.F., Tomanek L. and Williams S.L. (2006) The impacts of climate change in coastal marine systems. *Ecology Letters* 9, 228–241.
- Hay M.E., Paul V.J., Lewis S.M., Gustafson K., Tucker J. and Trindell R.N. (1988) Can tropical seaweeds reduce herbivory by growing at night? Diel patterns of growth, nitrogen content, herbivory, and chemical versus morphological defences. *Oecologia* 75, 233–245.
- Heidelberg K.B., O'Neil K.L., Bythell J.C. and Sebens K.P. (2010) Vertical distribution and diel patterns of zooplankton abundance and biomass at Conch Reef, Florida Keys (USA). *Journal of Plankton Research* 32, 75–91.
- Holomuzki J.R. and Messier S.H. (1993) Habitat selection by the stream mayfly *Paraleptophlebia guttata*. Journal of the North American Benthological Society 12, 126–135.
- Hughes T.P., Reed D.C. and Boyle M.-J. (1987) Herbivory on coral reefs: community structure following mass mortalities of sea urchins. *Journal of Experimental Marine Biology and Ecology* 113, 39–59.
- Hutchinson G.E. (1957) Concluding remarks. Cold Spring Harbor Symposia on Quantitative Biology 22, 415-427.
- Kavanagh K.D. and Olney J.E. (2006) Ecological correlates of population density and behavior in the circumtropical black triggerfish *Melichthys niger* (Balistidae). *Environmental Biology of Fishes* 76, 387–398.
- Legendre P. and Legendre L. (1998) Numerical ecology, 2nd edn. Amsterdam: Elsevier Science.
- Marnane M.J. and Bellwood D.R. (2002) Diet and nocturnal foraging in cardinalfishes (Apogonidae) at One Tree Reef, Great Barrier Reef, Australia. *Marine Ecology Progress Series* 231, 261–268.

- Maurer B.A. and McGill B.J. (2011) Measurement of species diversity. In Magurran A.E. and McGill B.J. (eds) *Biological diversity*. Oxford: Oxford University Press, pp. 55–65.
- McClanahan T.R. (2000) Recovery of a coral reef keystone predator, Balistapus undulatus, in East African marine parks. Biological Conservation 94, 191–198.
- Menge B.A. (1992) Community regulation: under what conditions are bottom-up factors important on rocky shores? *Ecology* 73, 755–765.
- Morlon H., Kefi S. and Martinez N.D. (2014) Effects of trophic similarity on community composition. *Ecology Letters* 17, 1495–1506.
- Mumby P.J., Harborne A.R., Williams J., Kappel C.V., Brumbaugh D.R., Micheli F., Holmes K.E., Dahlgren C.P., Paris C.B. and Blackwell P.G. (2007) Trophic cascade facilitates coral recruitment in a marine reserve. *Proceedings of the National Academy of Sciences* USA 104, 8362–8367.
- Nakajima R., Yoshida T., Othman B.H.R. and Toda T. (2009) Diel variation of zooplankton in the tropical coral-reef water of Tioman Island, Malaysia. *Aquatic Ecology* 43, 965–975.
- Nelson B.V. and Vance R.R. (1979) Diel foraging patterns of the sea urchin *Centrostephanus coronatus* as a predator avoiding strategy. *Marine Biology* 51, 251–258.
- **Ohlhorst S.L.** (1982) Diel migration patterns of demersal reef zooplankton. *Journal of Experimental Marine Biology and Ecology* 60, 1–15.
- Paine R.T. (1966) Food web complexity and species diversity. *American Naturalist* 100, 65–75.
- Pauly D. and Watson R. (2005) Background and interpretation of the 'Marine Trophic Index' as a measure of biodiversity. *Philosophical Transactions of the Royal Society B* 360, 415–423.
- Phinney J.T., Muller-Karger F., Dustan P. and Sobel J. (2001) Using remote sensing to reassess the mass mortality of *Diadema antillarum* 1983–1984. *Conservation Biology* 15, 885–881.
- Pinheiro H.T., Ferreira C.E.L., Joyeux J.-C., Santos R.G. and Horta P.A. (2011) Reef fish structure and distribution in a south-western Atlantic Ocean tropical island. *Journal of Fish Biology* 79, 1984–2006.
- Price J.H. and John D.M. (1980) Ascension Island, South Atlantic: a survey of inshore benthic macroorganisms, communities and interactions. Aquatic Botany 9, 251–278.
- R Core Team (2014) R: a language and environment for statistical computing. Vienna: R Foundation for Statistical Computing. http://www.R-project.org/.
- Rodríguez-Barreras R., Pérez M.E., Mercado-Molina A.E., Williams S.M. and Sabat A.M. (2014) Higher population densities of the sea urchin *Diadema antillarum* linked to wave sheltered areas in north Puerto Rico Archipelago. *Journal of the Marine Biological Association of the United Kingdom* 94, 1661–1669.
- Sebens K.P. and DeRiemer K. (1977) Diel cycles of expansion and contraction in coral reef anthozoans. *Marine Biology* 43, 247-256.
- Stuart-Smith R.D., Bates A.E., Lefcheck J.S., Duffy J.E., Baker S.C., Thomson R.J., Stuart-Smith J.F., Hill N.A., Kininmonth S.J., Airoldi L., Becerro M.A., Campbell S.J., Dawson T.P., Navarrete S.A., Soler G.A., Strain E.M.A., Willis T.J. and Edgar G.J. (2013) Integrating abundance and functional traits reveals new global hotspot of fish diversity. *Nature* 501, 539-542.
- Tsiamis K., Peters A.F., Shewring D.M., Asensi A.O., Van West P. and Küpper F.C. (2014) Marine benthic algal flora of Ascension Island, South Atlantic. *Journal of the Marine Biological Association of the United Kingdom.* doi: 10.1017/S0025315414000952.
- Tuya F., Martin J.A. and Luque A. (2004) Patterns of nocturnal movement of the long-spined sea urchin *Diadema antillarum* (Philippi)

in Gran Canaria (the Canary Islands, central East Atlantic Ocean). *Helgoland Marine Research* 58, 26–31.

- Vinueza L.R., Menge B.A., Ruiz D. and Palacios D.M. (2014) Oceanographic and climatic variation drive top-down/bottom-up coupling in the Galápagos intertidal meta-ecosystem. *Ecological Monographs* 84, 411–434.
- Whittaker R.J., Rigal F., Borges P.A.V., Cardoso P., Terzopoulou S., Casanoves F., Pla L., Guilhaumon F., Ladle R.J. and Triantis K. (2014) Functional biogeography of oceanic islands and the scaling of functional diversity in the Azores. PNAS 111, 13709-13714.
- Wilson S.K., Graham N.A.J. and Polunin N.V.C. (2007) Appraisal of visual assessments of habitat complexity and benthic composition on coral reefs. *Marine Biology* 151, 1069–1076.

Wirtz P., Bingeman J., Bingeman J., Frickle R., Hook T.J. and Young J. (2014) The fishes of Ascension Island, central Atlantic Ocean – new records and an annotated checklist. *Journal of the Marine Biological Association of the United Kingdom*. doi: 10.1017/S0025315414001301.

and

Witman J.D. (1985) Refuges, biological disturbance, and rocky subtidal community structure in New England. *Ecological Monographs* 55, 421–445.

Correspondence should be addressed to:

P.E. Brewin

Shallow Marine Surveys Group, PO Box 598, Stanley, Falkland Islands FIQQ 1ZZ, South Atlantic email: pbrewin@smsg-falklands.org