

¹⁴C DISTRIBUTION IN ATMOSPHERIC AND AQUATIC ENVIRONMENTS AROUND QINSHAN NUCLEAR POWER PLANT, CHINA

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ABSTRACT. Atmospheric CO₂ and aquatic water samples were analyzed to evaluate the environmental ¹⁴C enrichment due to operation of the Qinshan nuclear power plant (NPP), where two heavy-water reactors and five pressurized-water reactors are employed. Elevated ¹⁴C-specific activities (2–26.7 Bq/kg C) were observed in the short-term air samples collected within a 5-km radius, while samples over 5 km were close to background levels. The ¹⁴C-specific activities of dissolved inorganic carbon (DIC) in the surface seawater samples ranged from 196.8 to 206.5 Bq/kg C (average 203.4 Bq/kg C), which are close to the background value. No elevated ¹⁴C level in surface seawater was found after 20 years of operation of Qinshan NPP, indicating that the ¹⁴C discharged was well diffused. The results of the freshwater samples show that excess ¹⁴C-specific activity (average 17.1 Bq/kg C) was found in surface water and well water samples, while no obvious ¹⁴C increase was found in drinking water (groundwater and tap water) compared to the background level.

INTRODUCTION

Radiocarbon is produced not only naturally by neutron interaction with nitrogen in the atmosphere, but also by anthropogenic activities: atmospheric nuclear weapon tests and the nuclear power industry. A series of bomb tests, which happened mainly in the 1950s and 1960s, released 220 PBq ¹⁴C to the environment (IAEA 2004). After the nuclear moratorium on atmospheric nuclear bomb tests signed in 1963, the enhanced ¹⁴C level has been declining to the natural level (Kim et al. 2000). ¹⁴C is produced in nuclear power reactors by ¹⁴N(n, p)¹⁴C reactions with nitrogen in fuels, moderators, and coolants as a primary impurity; by ¹⁷O(n, α)¹⁴C reactions in oxide fuels, moderators, and coolants; and by ¹³C(n, γ)¹⁴C reactions in graphite moderators (Yim and Caron 2006; Povinec et al. 2008). The mean ¹⁴C production rates for pressurized-water reactors (PWRs) and heavy-water reactors (HWRs) are 1.48 and 10.13 TBq GW_e⁻¹ yr⁻¹, respectively (IAEA 2004).

¹⁴C produced in nuclear power reactors is mainly disposed in the solid radioactive waste, but still a small quantity of ¹⁴C is released in gaseous and liquid effluents. Previous studies have shown that nearly all ¹⁴C is emitted in the form of ¹⁴CO₂, except in PWRs where ¹⁴C is mainly released as ¹⁴CH₄ (Hertelendi et al. 1989; Yim and Caron 2006). ¹⁴C in aqueous liquid waste is usually in chemical forms of carbonate and bicarbonate, depending on the solution pH (IAEA 2004). Investigations of ¹⁴C activity around nuclear power stations have found that elevated ¹⁴C concentrations were observed in environmental samples (Magnusson et al. 2004; Molnár et al. 2007; Varlam et al. 2007; Povinec et al. 2009).

The Qinshan nuclear power plant (NPP), with an electrical output of 4156 MW_e, consists of five PWRs and two HWRs. The two HWRs were expected to be the main contributor of ¹⁴C enrichment in the surrounding area (Wang et al. 2013). In previous studies, ¹⁴C spatial and temporal distribution in the terrestrial environment was investigated by analyzing vegetation samples, and the internal dose of ¹⁴C by ingestion of foodstuffs was evaluated (Wang et al. 2012, 2013). The aim of the present investigation was to study a spatial distribution of ¹⁴C in the atmospheric and aquatic environment where ¹⁴C gaseous and liquid waste was directly released. This study includes the ¹⁴C measurements of CO₂ in air, as well as the ¹⁴C of dissolved inorganic carbon (DIC) in water.

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MATERIAL AND METHODS

Site Description

Qinshan NPP, located in Haiyan County, Zhejiang Province, in southeast China, comprises three power plants, referred to as Plant I, Plant II, and Plant III. The 300-MW_e PWR reactor of Plant I began operation in April 1994. Furthermore, four PWR reactors (600 MW_e) of Plant II began commercial operation in April 2002, May 2004, August 2010, and April 2012. Plant III, which used two HWRs (CANDU type) of 728 MW_e, started to operate in December 2002 and June 2003. The ¹⁴C released from Qinshan NPP was through airborne and liquid discharge. The positions of the venting stacks and outlets of each plant are shown in Figures 1 and 2. The active liquid waste of Qinshan NPP was discharged discontinuously, depending on the running condition of the plant.

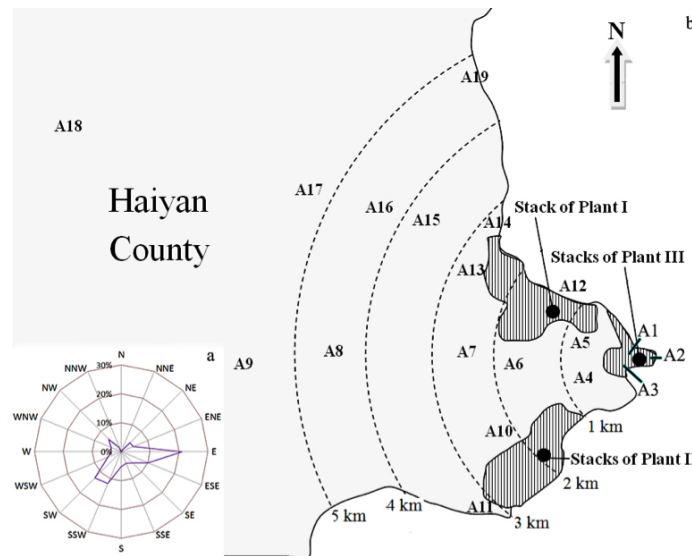


Figure 1 (a) Wind frequency at the Qinshan site in the period of air sampling (from 5 to 12 August 2011). (b) Sites of collection of the air samples at the Qinshan NPP.

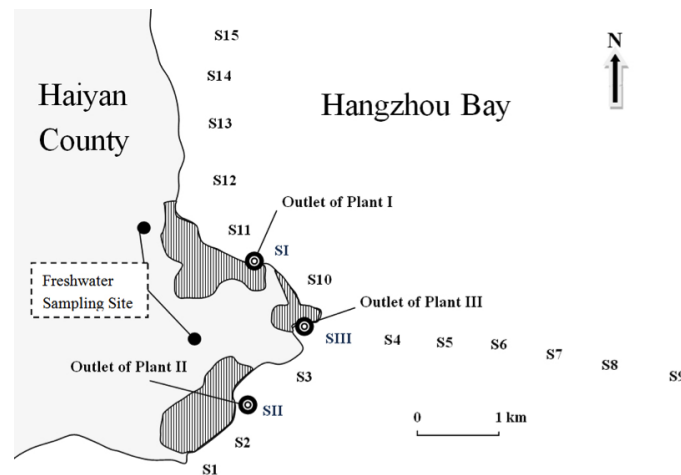


Figure 2 Locations of the sampling sites where seawater samples were collected on 28 May 2013.

The terrain around Qinshan NPP is basically flat, except for the small hill (170 m high) named Qinshan that lies to the west of the Plant III stacks (samples A4 and A5 in Figure 1 were located on the hill). Meteorological data were obtained from three meteorological monitoring stations within the Qinshan site.

Sample Collection

To determine the ¹⁴C spatial distribution in local air, sampling was done at a flow rate of 1.5 L/min for 3 hr using CO₂ collection equipment, in which CO₂ was absorbed in two bubbling bottles with 500 mL of NaOH solution (2 mol/L) with the collection efficiency better than 98% (Yang et al. 1997). During 5–12 August 2011, 19 atmospheric CO₂ samples were collected in the vicinity of Qinshan NPP (Figure 1b), taking into account wind direction, the distance to the stacks of Plant III, and the availability of sample collection. During the 1-week sampling period, ¹⁴C was continuously discharged from three plants under steady meteorological conditions, where seven reactors were all in operation. Two reference samples were collected at a site 20 km from Qinshan NPP, an area outside of the plant's ¹⁴C contamination. Reference sampling points were situated in areas with an estimated similar density of buildup and with similar intensity of local road transport.

In May 2013, surface seawater samples were collected from the Qinshan site for DIC ¹⁴C analyses using two 1-L glass bottles, following the sampling procedure described by Hinger et al. (2010). In order to evaluate the DIC ¹⁴C concentrations in the outlets of Qinshan NPP, possible ¹⁴C enrichment near the shore, and ¹⁴C diffusion in seawater, 18 surface seawater samples were distributed around Qinshan NPP (Figure 2). Freshwater samples, including underground water, surface water (river water and pool water), and tap water, were collected in a village next to Qinshan NPP (Figure 2). The reference samples of water, whose ¹⁴C concentrations were used as background, were collected at Jiading in Shanghai (100 km away). Both sampled rivers do not cross the karst area.

PRETREATMENT

Air Sample

¹⁴C in the NaOH solution was recovered as BaCO₃ precipitate by adding BaCl₂ and NH₄Cl according to the method of Yang et al. (2010). After filtrating and drying, about 20 mg of BaCO₃ was decomposed with H₃PO₄ to re-evolve CO₂, which was then transferred to a vacuum system for purification and quantification. This was followed by transferring CO₂ (approximately 1 mg C) to a reaction tube containing zinc (Aldrich, #324930), titanium hydride (Alfa Aesar, #12857), and Fe catalyst (Alfa Aesar, 350 mesh, #39813). The reaction tube was sealed and heated at 540°C in a furnace for 8 hr, and graphite formed at the surface of the catalyst (Xu et al. 2007).

Water Sample

As shown in Figure 3, an apparatus extracting DIC from water was set up and applied to pretreat the water samples. The extracting device was connected to a vacuum system; the pressure of the entire system can reach 2 Pa. The extraction procedure comprised the following. About 300 mL of water sample and appropriate phosphoric acid were introduced into the bunsen flask and separating funnel (4 and 5 in Figure 3) in a glove box. After sealing the valves, the vessel was then connected to the extracting system, followed by pumping the vacuum system up to the valve of the glass vessel. In the next step, the valve of the nitrogen cylinder was opened and the flow rate of N₂ was controlled at 450 mL/min. The water sample was acidified and heated (at 80°C) to release CO₂, which was then transferred to the water trap by purified N₂ gas. The last stage was needed to convert CO₂ to graphite, similar to the pretreatment of the air samples described above.

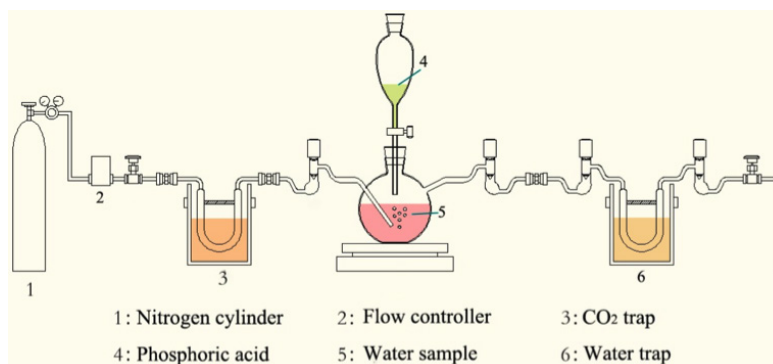


Figure 3 Diagram of the DIC extraction device for water samples

Measurement

¹⁴C analysis was performed at the 0.6MV accelerator mass spectrometry (AMS) facility of Peking University, with a measurement accuracy better than 0.4% and a machine background lower than 0.03 pMC (Liu et al. 2007). Blank coal samples and NIST oxalic acid standard (SRM-4990C) samples were also graphitized and measured at the same time. After isotopic fractionation correction (Stuiver and Polach 1977), the final results are presented as normalized specific activities (Bq/kg C) (Wang et al. 2012) and $\Delta^{14}\text{C}$ (‰) (Roussel-Debet et al. 2006).

RESULTS AND DISCUSSION

Air Sample

As shown in Table 1, the average ¹⁴C value of two reference air samples is 231.2 Bq/kg C (15.2‰ in $\Delta^{14}\text{C}$), which is lower than the western Europe ¹⁴C background (237.0 Bq/kg C in specific activity, 40.1‰ in $\Delta^{14}\text{C}$) measured at Schauinsland station (47°55'N, 7°54'E) in August 2011 (Levin et al. 2013). The disagreement of these two values is probably due to the disturbance of regional ¹⁴C-free fossil fuel CO₂ emission. To eliminate the regional CO₂ influence, the ¹⁴C value of the reference samples was used as the background in the present study.

The ¹⁴C results of the air samples (Table 1) show that the specific activity ranges from 231.2 to 257.9 Bq/kg C (average 239.5 Bq/kg C). Compared to the background, excess or equal levels of ¹⁴C were found in all sampling sites. The maximum excess ¹⁴C-specific activity of 26.7 Bq/kg C (11.5% above the background) was observed in A3, which was located in the prevailing wind direction of the 55-m-high stack of Plant III.

In a previous study of vegetation samples (Wang et al. 2013), it was well demonstrated that the diffusion of ¹⁴C discharged from Qinshan NPP was predominantly affected by meteorological conditions. This conclusion is further validated by this study. In the region mainly contaminated by Plant III-derived ¹⁴C (radius <3 km), samples located downwind in the eastern direction have higher ¹⁴C concentrations than those of other directions, though the distances to Plant III stack were almost the same (A5>A4>A12, A7>A13>A10 in specific activity). As shown in Table 1, ¹⁴C values of the samples collected beyond 5 km (A9, A17, and A18) are close to the background, indicating that the ¹⁴C-influenced zone of Qinshan NPP was limited to a 5-km radius in this investigation.

In contrast to the atmospheric CO₂, data of vegetation samples collected at Qinshan in 2010 show a higher maximum excess value (18.1% above the background) and larger range of influence (6.5 km radius) (Wang et al. 2013). The discrepancy is probably attributable to different sampling periods

of these two kinds of sample. While atmospheric CO₂ samples represent a 1-week sampling period, the plant moss represents a 1-yr assimilation period. Since both ¹⁴C emission and meteorological conditions do not remain steady in a year, the annual average concentrations should be different from 1-week average concentrations. Similar discrepancies between short-term air and vegetation data were observed in studies of other countries (Milton et al. 1995; Dias et al. 2009).

Table 1 ¹⁴C concentrations in air samples collected at Qinshan site (from 5 to 12 August 2011).

Sample ID	¹⁴ C-specific activity (Bq/kg C)	1σ error	Δ ¹⁴ C (‰)	1σ error
A1	242.8	0.7	66.2	3.1
A2	237.7	0.9	43.8	4.0
A3	257.9	0.7	132.5	3.1
A4	243.1	0.8	78.5	3.5
A5	245.6	0.8	67.5	3.5
A6	249.8	1.0	96.9	4.4
A7	241.6	0.7	60.9	3.1
A8	233.2	0.8	24.0	3.5
A9	232.2	0.7	19.6	3.1
A10	236.5	1.0	38.5	4.4
A11	238.1	0.7	45.5	3.1
A12	243.0	0.8	67.1	3.5
A13	239.6	0.8	52.1	3.5
A14	242.2	0.7	63.5	3.1
A15	235.9	0.8	35.9	3.5
A16	233.3	0.7	24.5	3.1
A17	231.2	0.9	15.2	4.0
A18	232.3	0.7	20.1	3.1
A19	233.7	1.1	34.6	3.5
Background ^a	231.2	0.9	15.2	4.0

a: average ¹⁴C values of two air samples collected at reference site.

Water Sample

Tables 2 and 3 show the ¹⁴C concentrations in surface seawater and freshwater samples, respectively. The ¹⁴C-specific activities in the seawater samples were in the range of 196.8–206.5 Bq/kg C (average 203.4 Bq/kg C). Compared to the background value (205.9 ± 1.8 Bq/kg C) measured from two reference samples, no obvious difference was observed.

As indicated by Figure 4, the ¹⁴C values of the samples collected near the coast were close to the background, while samples whose sites were not adjacent to the coast had lower ¹⁴C values. Considering that the reference samples were collected near the coast, it can be concluded that coastal surface seawater samples had higher ¹⁴C concentrations, in comparison with the samples collected far away from the coast. A possible explanation is that the ¹⁴C concentrations in coastal surface seawater samples could be influenced by terrestrial freshwater samples whose ¹⁴C levels were higher (Table 3) (Shen et al. 2004).

Table 2 ^{14}C concentrations in seawater samples collected in May 2013 in the vicinity of Qinshan NPP.

Sample ID	^{14}C -specific activity (Bq/kg C)	1 σ error	$\Delta^{14}\text{C}$ (‰)	1 σ error
S1	204.7	1.5	-101.1	6.6
S2	205.9	1.4	-95.9	6.1
S3	206.5	1.3	-93.2	5.7
S4	204.4	1.4	-102.4	6.1
S5	198.8	1.0	-127.0	4.4
S6	199.5	1.2	-124.0	5.3
S7	196.8	1.7	-135.8	7.5
S8	198.2	1.6	-129.7	7.0
S9	201.7	1.3	-114.3	5.7
S10	204.2	1.5	-103.3	6.6
S11	206.0	1.4	-95.4	6.1
S12	203.8	1.3	-105.1	5.7
S13	206.1	1.2	-95.0	5.3
S14	205.7	1.0	-96.7	4.4
S15	205.5	1.1	-97.6	4.8
SI	205.9	1.2	-95.9	5.3
SII	202.2	1.3	-112.1	5.7
SIII	206.1	1.2	-95.0	5.3
Background ^a	205.9	1.3	-95.9	5.7

a: average ^{14}C values of two seawater samples collected at reference site.

Table 3 Results of freshwater samples around Qinshan NPP collected in May 2013.

Sample	Qinshan site		Reference site ^a	
	^{14}C -specific activity (Bq/kg C)	$\Delta^{14}\text{C}$ (‰)	^{14}C -specific activity (Bq/kg C)	$\Delta^{14}\text{C}$ (‰)
Surface water1	225.1 \pm 1.2	-11.5 \pm 5.3	207.3 \pm 1.5	-89.7 \pm 6.6
Surface water2	232.0 \pm 1.4	18.8 \pm 6.1	210.1 \pm 0.8	-77.4 \pm 3.5
Well water	220.6 \pm 1.2	31.3 \pm 5.3	208.9 \pm 1.1	-82.7 \pm 4.8
Groundwater	235.3 \pm 1.8	33.2 \pm 7.9	236.9 \pm 2.1	40.3 \pm 9.2
Tap water	204.6 \pm 1.3	-101.6 \pm 5.7	194.2 \pm 1.5	-147.2 \pm 6.6

^aLocated 100 km from the Qinshan site.

The liquid waste of Qinshan NPP was discharged discontinuously, but unfortunately we failed to obtain the accurate discharge time, leading to the fact that no liquid waste was released during the sampling period. As a consequence, the results of samples SI, SII, and SIII, which were collected to evaluate the ^{14}C levels in the outlets of Qinshan NPP, show no increase in ^{14}C concentration. It can be concluded from the results that the ^{14}C discharged previously was diluted and diffused well, resulting in no ^{14}C enrichment in seawater observed.

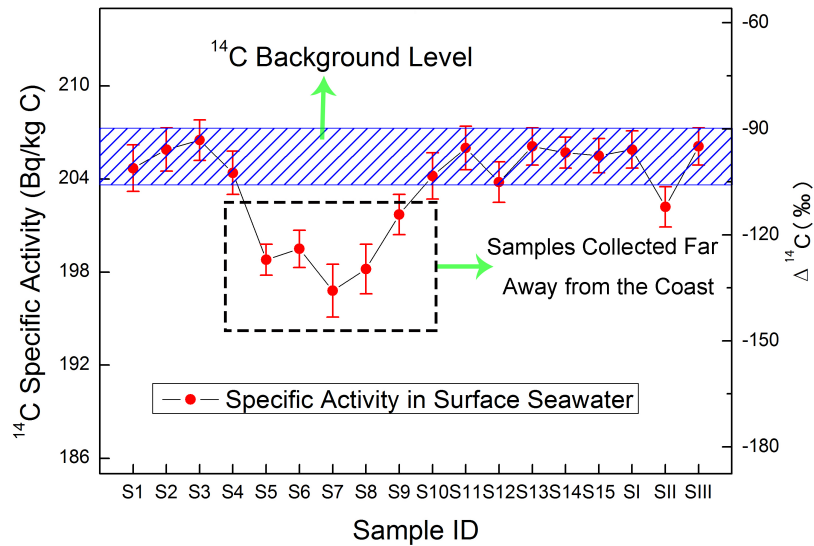


Figure 4 ¹⁴C concentration distribution in surface seawater samples

For the freshwater samples collected at Qinshan, elevated ¹⁴C concentrations were found in the surface water and well water, compared to the values of reference samples. The excess values are probably caused by the atmosphere-water gas exchange process (Shen et al. 2004), through which the ¹⁴CO₂ released from Qinshan NPP finds its way into water. The ¹⁴C levels in underground water and tap water samples are similar to those collected at the reference site, resulting in no obvious increase in ¹⁴C concentration at the Qinshan site.

CONCLUSION

The ¹⁴C concentrations in the vicinity of Qinshan NPP have been studied by analyzing air and water samples using AMS. Enhanced ¹⁴C values were observed in the air samples, which were collected within a 5-km radius in this short-term air investigation. The maximum ¹⁴C excess value was 26.7 Bq/kg C, 11.5% higher than the background. The results of surface seawater samples show no enhanced activity, indicating that the ¹⁴C discharged previously diffused well and no enrichment in the seawater existed after 20 yr of Qinshan NPP operation. In comparison with the reference site, an average of 17.1 Bq/kg C excess ¹⁴C-specific activity was observed in the surface water and well water collected at Qinshan, which is probably attributed to the atmosphere-water gas exchange. No obvious enhanced ¹⁴C activity in drinking water (underground water and tap water) was found in this study.

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REFERENCES

Dias CM, Stenström K, Leão ILB, Santos RV, Nícoli IG, Skog G, Ekström P, Corrêa RS. 2009. ¹⁴CO₂ dispersion around two PWR nuclear power plants in Brazil. *Journal of Environmental Radioactivity* 100(7): 574–80.

Hertelendi E, Uchrin G, Ormai P. 1989. ¹⁴C release in various chemical forms with gaseous effluents from the Paks nuclear power plant. *Radiocarbon* 31(3):754–61.

Hinger EN, Santos GM, Druffel ERM, Griffin S. 2010.

- Carbon isotope measurements of surface seawater from a time-series site off southern California. *Radiocarbon* 52(1):69–89.
- International Atomic Energy Agency (IAEA). 2004. Management of waste containing tritium and carbon-14. Technical Report Series No 421. Vienna: IAEA.
- Kim CK, Lee SK, Rho BH, Lee YG. 2000. Environmental distribution and behavior of ^3H and ^{14}C around Wolsong Nuclear Power Plant. *Health Physics* 78(6):693–9.
- Levin I, Kromer B, Hammer S. 2013. Atmospheric $\Delta^{14}\text{CO}_2$ trend in Western Europe background air from 2000 to 2012. *Tellus B* 65:20092, <http://dx.doi.org/10.3402/tellusb.v65i0.20092>.
- Liu KX, Ding XF, Fu DP, Pan Y, Wu XH, Guo ZY, Zhou LP. 2007. A new compact AMS system at Peking University. *Nuclear Instruments and Methods in Physics Research B* 259(1):23–6.
- Magnusson Å, Stenström K, Skog G, Adliene D, Adlyš G, Hellborg R, Olariu A, Zakaria M, Rääf C, Mattsson S. 2004. Levels of ^{14}C in the terrestrial environment in the vicinity of two European nuclear power plants. *Radiocarbon* 46(2):863–8.
- Milton GM, Kramer SJ, Brown RM, Repta CJW, King KJ, Rao RR. 1995. Radiocarbon dispersion around Canadian nuclear facilities. *Radiocarbon* 37(2):485–96.
- Molnár M, Bujtás T, Svingor É, Futó I, Světlík I. 2007. Monitoring of atmospheric excess ^{14}C around Paks nuclear power plant, Hungary. *Radiocarbon* 49(2):1031–43.
- Povinec PP, Chudý M, Šivo A, Šimon J, Holý K, Richtáriková M, Morávek J. 2008. Impact of the Bohunice nuclear power plant on atmospheric radiocarbon. *Applied Radiation and Isotopes* 66(11):1686–90.
- Povinec PP, Chudý M, Šivo A, Šimon J, Holý K, Richtáriková M. 2009. Forty years of atmospheric radiocarbon monitoring around Bohunice nuclear power plant, Slovakia. *Journal of Environmental Radioactivity* 100(2):125–30.
- Roussel-Debet S, Gontier G, Siclet F, Fournier M. 2006. Distribution of carbon 14 in the terrestrial environment close to French nuclear power plants. *Journal of Environmental Radioactivity* 87(3):246–59.
- Shen CD, Yi WX, Yu KF, Sun YM, Yang Y, Zhou B. 2004. Interannual ^{14}C variations during 1977–1998 recorded in coral from Daya Bay, south China Sea. *Radiocarbon* 46(2):595–601.
- Stuiver M, Polach HA. 1977. Discussion: reporting of ^{14}C data. *Radiocarbon* 19(3):355–63.
- Varlam C, Stefanescu I, Varlam M, Popescu I, Faurescu I. 2007. Applying the direct absorption method and LSC for ^{14}C concentration measurement in aqueous samples. *Radiocarbon* 49(2):281–9.
- Wang ZT, Xiang YY, Guo QJ. 2012. Levels of ^{14}C in the tree rings in the vicinity of Qinshan nuclear power plant, China. *Radiocarbon* 54(2):195–202.
- Wang ZT, Xiang YY, Guo QJ. 2013. Terrestrial distribution of ^{14}C in the vicinity of Qinshan nuclear power plant, China. *Radiocarbon* 55(1):59–66.
- Xu XM, Trumbore SE, Zheng SH, Southon JR, McDuffee KE, Luttgen M, Liu JC. 2007. Modifying a sealed tube zinc reduction method for preparation of AMS graphite targets: reducing background and attaining high precision. *Nuclear Instruments and Methods in Physics Research B* 259(1):320–9.
- Yang HY, Liu W, Wen XL. 1997. Sampling and determination of carbon-14 in air. *Radiation Protection* 17(1):68–72. In Chinese.
- Yang YH, Kang DW, Lee GB. 2010. Evaluation of ^{14}C release characteristics in Korean standard pressurized water reactor. *Nuclear Engineering and Design* 240(10):3611–5.
- Yim M-S, Caron F. 2006. Life cycle and management of carbon-14 from nuclear power generation. *Progress in Nuclear Energy* 48(1):2–36.